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Antioxidant responses of peanut roots exposed to realistic groundwater doses of arsenate: Identification of glutathione S-transferase as a suitable biomarker for metalloid toxicity

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1 **Antioxidant responses of peanut roots exposed to realistic groundwater doses**  
2 **of arsenate: identification of glutathione S-transferase as a suitable biomarker for**  
3 **metalloid toxicity**  
4

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17 **Abstract**

18 Arsenic (As)-polluted groundwater constitutes a serious problem for peanut plants, as roots can  
19 accumulate the metalloid in their edible parts. Characterization of stress responses to As may  
20 help to detect potential risks and identify mechanisms of tolerance, being the induction of  
21 oxidative stress a key feature. Fifteen-day old peanut plants were treated with arsenate in order  
22 to characterize the oxidative stress indexes and antioxidant response of the legume under  
23 realistic groundwater doses of the metalloid. Superoxide anion ( $O_2^{\cdot-}$ ) and hydrogen peroxide  
24 ( $H_2O_2$ ) histochemical staining along with the activities of NADPH oxidase, superoxide dismutase  
25 (SOD), catalase (CAT) and thiol (glutathione and thioredoxins) metabolism were determined in  
26 roots. Result showed that at 20  $\mu M$   $H_2AsO_4^-$ , peanut growth was reduced and the root  
27 architecture was altered.  $O_2^{\cdot-}$  and  $H_2O_2$  accumulated at the root epidermis, while lipid  
28 peroxidation, NADPH oxidase, SOD, CAT and glutathione S-transferase (GST) activities  
29 augmented. These variables increased with increasing As concentration (100  $\mu M$ ) while  
30 glutathione reductase (GR) and glutathione peroxidase/peroxiredoxin (GPX/PRX) were  
31 significantly decreased. These findings demonstrated that the metalloid induced physiological  
32 and biochemical alterations, being the NADPH oxidase enzyme implicated in the oxidative burst.  
33 Additionally, the strong induction of GST activity, even at the lowest  $H_2AsO_4^-$  doses studied, can  
34 be exploited as suitable biomarker of As toxicity in peanut plants, which may help to detect risks  
35 of As accumulation and select tolerant cultivars.

36

37 *Keywords:* arsenic; glutathione S-transferase; oxidative stress; peanut

38 *Abbreviations:*

39 1-chloro-2, 4-dinitrobenzene (CDNB)

40 3,3-diaminobenzidine (DAB)

41 Arsenic (As)

42 Catalase (CAT)

43 Glutathione (GSH)

44 Glutathione peroxidase/peroxiredoxin (GPX/PRX)

45 Glutathione reductase (GR)

46 Glutathione S-transferase (GST)

47 Hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>)

48 Nitroblue tetrazolium (NBT)

49 Potassium iodide (KI)

50 Superoxide anion (O<sub>2</sub><sup>•-</sup>)

51 Superoxide dismutase (SOD)

52 Thiobarbituric-reactive substances (TBARS)

## 53 1. Introduction

54 Arsenic (As) is an ubiquitous metalloid present in many ecosystems. The average concentration  
55 of As in the lithosphere ranges between 20 and 30  $\mu\text{M}$ , occupying the position number 52 in the  
56 list of the most abundant elements in nature (Adriano 2001). Arsenate ( $\text{H}_2\text{AsO}_4^-$ , As(V)) and  
57 arsenite ( $\text{H}_2\text{AsO}_3^-$ , As(III)) are the prevalent inorganic chemical species, being more toxic than  
58 the organic forms (Mandal and Suzuki 2002; Zhao et al. 2010). These contaminants are  
59 receiving much attention due to human population chronic poisoning caused by drinking water  
60 polluted in several regions of Asia and South America (Zhao et al. 2010). In addition, crop plants  
61 can accumulate large amounts of As in edible parts when irrigated with As-containing water;  
62 acting as the first stage of As distribution in the trophic chain (Smedley and Kinniburgh, 2002). In  
63 Argentina, particularly in Córdoba province, As has become a public health problem due to  
64 numerous cases of human poisoning. Francisca et al. (2006) found that 90% of this region is  
65 affected by groundwater As concentrations that exceeds the maximum allowed level of the  
66 metalloid in drinking water (0.1  $\mu\text{M}$ ) (FAO, OMS-OPS, 2007), reaching in some areas up to 20  
67  $\mu\text{M}$  (Cabrera et al. 2005). Peanut is a very important crop in Argentina, specifically in Córdoba,  
68 as it represents approximately 90% of the Argentinean production (Ministry of Agroindustry,  
69 Argentina, 2006). The fruit of this legume is a pod with one to five seeds that develops  
70 underground, in putative intimate contact with soil contaminants. One of the major constraints  
71 that Córdoba province has, given the water scarcity in the area, is that crops require artificial  
72 irrigation by sprinkler systems using groundwater rich in As (Bustingorri and Lavado, 2012;  
73 2014) representing a potential risk for crops. This is aggravated by the deep penetration of plant  
74 roots in soil strata, which is a typical response of plants suffering from drought stress (Sharp and  
75 LeNoble 2002).

76 Arsenate form ( $\text{H}_2\text{AsO}_4^-$  As(V)) is an analog of phosphate ( $\text{H}_2\text{PO}_4^-$ ) and it can be easily  
77 absorbed by plant by  $\text{H}_2\text{PO}_4^-$  transporters (Tripathi et al. 2007; Verbruggen et al. 2009). Once  
78 inside the cell, most of  $\text{H}_2\text{AsO}_4^-$  (As(V)) is reduced to  $\text{H}_2\text{AsO}_3^-$  (As(III)), by different  $\text{H}_2\text{AsO}_4^-$

79 reductases (Dhankher et al. 2006; Duan et al. 2007; Sánchez-Bermejo et al. 2014). Therefore,  
80 As is present predominantly in roots and shoots as  $\text{H}_2\text{AsO}_3^-$ , which binds to sulfhydryl groups of  
81 thiol-rich peptides such as glutathione (GSH) and phytochelatins (PCs) (Pickering et al. 2000).  
82 Although the precise mechanisms of toxicity are not completely understood, plants exposed to  
83 both As species showed substantial redox homeostasis imbalance, with accumulation of reactive  
84 oxygen species (ROS) (Requejo and Tena 2005; Talukdar 2013), and changes in antioxidant  
85 responses (Mylona et al. 1998). It is feasible that the As high affinity for sulfhydryl groups causes  
86 GSH depletion and changes in enzymatic activity that leads to oxidative stress (Hernández et al.  
87 2015). Biomolecules such as lipids, proteins and nucleic acids are principal targets of ROS,  
88 which cause the characteristic oxidative damage symptoms (Verbruggen et al. 2009). Another  
89 source of ROS is constituted by the plasma membrane-associated NADPH oxidase family that  
90 has a low redox potential able to reduce  $\text{O}_2$  to superoxide anion ( $\text{O}_2^{\cdot-}$ ) (Mittler et al. 2011).  
91 Moreover, under metal(loid)s stress these enzymes are activated increasing ROS accumulation  
92 (Hernández et al. 2015).  
93 Plants possess an effective antioxidant defense system that scavenges ROS accumulated by  
94 aerobic metabolism, and is comprised of antioxidant enzymes and metabolites (Noctor et al.  
95 2012). Superoxide dismutase (SOD) constitutes the first line of defense dismutating  $\text{O}_2^{\cdot-}$  into  
96 hydrogen peroxide ( $\text{H}_2\text{O}_2$ ), which is then reduced to water and oxygen by ascorbate peroxidase  
97 (APX) and catalase (CAT) (Mishra et al. 2006). These enzymes constitutes the major ROS  
98 scavenging mechanisms of plants (Mittler 2012) and their induction under different metal(loid)  
99 treatments has been demonstrated in several species (Sharma and Dietz, 2009). One of the  
100 most studied antioxidant metabolite essential to cope with the oxidative stress induced by toxic  
101 elements such as Cd, Hg and As is the tripeptide GSH (Li et al. 2006; Bianucci et al. 2012;  
102 Sobrino-Plata et al. 2013). GSH transits from its reduced form (thiol) to its oxidized form (GSSG,  
103 disulfide), keeping a strong reducing environment in the cell which plays an important role in

104 preventing excessive oxidation of cellular biomolecules under different stressing conditions  
105 (Jozefczak et al. 2012). Besides, a series of antioxidant enzymes related to GSH metabolism are  
106 involved in the detoxification of free radicals (Foyer and Noctor 2011). Glutathione reductase  
107 (GR) regenerates GSH from GSSG at the expense of NADPH to maintain a high GSH/GSSG  
108 ratio (Gill et al. 2013). Glutathione-S-transferases (GSTs) are a large and diverse group of  
109 enzymes involved in the detoxification of dangerous xenobiotics catalyzing their conjugation with  
110 GSH. Moreover, GSTs can bind GSH with organic peroxides and radicals generated by ROS  
111 attack, which can amplify in a chain reaction the oxidative damage caused by toxic elements,  
112 preventing their reaction with many important biomolecules (Gill and Tuteja 2010). The enzyme  
113 named misleadingly glutathione peroxidase (GPX), a peroxidase able to use GSH and  
114 thioredoxin (TRX), constitutes a special case (Herbette et al., 2002). The particularity is that the  
115 TRX is known to be a more efficient reducing substrate than GSH in certain plant species (Iqbal  
116 et al., 2006). Thus, these enzymes should be considered as peroxiredoxins (PRX) rather than  
117 GPX (Navrot et al., 2006) and are named accordingly in our work as GPX/PRX.

118 Arsenate and arsenite are highly soluble and stable in water under a wide range of pH and  
119 redox potential conditions. However, under oxidizing environments as occur in many areas of  
120 Córdoba province, the prevalent arsenic form is arsenate (Blarasin et al. 2014). Thus, our  
121 experiments were conducted to characterize oxidative stress indexes and the antioxidant  
122 response of peanut (*Arachis hypogaea* L.) plants exposed to environmentally realistic As(V)  
123 levels, considering the economic relevance of this legume crop in Córdoba province. This study  
124 will provide information about the mechanisms of toxicity evoked by As(V), which can be used as  
125 bioindexes of putative risks to consumers. Most sensitive stress indexes could be used to select  
126 tolerant cultivars with lower toxic symptoms and able to exclude As from edible parts of the  
127 plants.

## 128 **2. Material and Methods**

## 129 **2.1. Plant material and experimental design**

130 Peanut seeds, kindly supplied by "El Carmen S.A." Córdoba, Argentina, were surface sterilized  
131 following the method previously described by Vincent (1970). Then, they were germinated at  
132 28°C in Petri dishes on a layer of Whatman N°1 filter paper and moistened cotton, until the  
133 radicle reached 3-5 cm. Seedlings were transferred to an hydroponic system with Hoagland's  
134 nutrient solution (Hoagland and Arnon 1950) devoid of As (control) or containing different  
135 concentrations of Na<sub>2</sub>HAsO<sub>4</sub>·7H<sub>2</sub>O. The metalloid doses used are equivalent to those found in  
136 groundwater in Córdoba province (6 and 20 µM) (Cabrera et al., 2005) and a concentration  
137 above those mentioned (100 µM).

138 Plants were grown in a greenhouse under controlled environment (light intensity of 200 µmol m<sup>-2</sup>  
139 s<sup>-1</sup>, 16/8 h day/night cycle, constant temperature of 28°C and relative humidity of 50%) for 15  
140 days. At harvest, plant length was measured and shoots and roots were used for different  
141 analysis.

142

## 143 **2.2. Root structure and histochemical detection of H<sub>2</sub>O<sub>2</sub> and O<sub>2</sub><sup>•-</sup>**

144 For anatomical and histological studies, the main root of a fresh peanut plant was cut into 5 mm  
145 length portions, at 1 cm from the root tip as described in Bianucci et al. (2012). The tissues were  
146 cut with a rotary microtome and the samples were stained as described by Johansen (1940) and  
147 O'Brien and Mc Cully (1981). The photomicrographs were taken using an Axiophot Carl Zeiss  
148 microscope (Germany).

149 *In situ* localization of superoxide anion (O<sub>2</sub><sup>•-</sup>) was done incubating freshly roots segments in 1  
150 mM nitroblue tetrazolium (NBT), prepared in 10 mM sodium citrate buffer pH 6, for 8 hours,  
151 following the procedure described by Frahy and Schopfer (2001). H<sub>2</sub>O<sub>2</sub> was visually detected  
152 incubating freshly roots segments in 1 mg mL<sup>-1</sup> 3,3-diaminobenzidine (DAB) as substrate for 8  
153 hours (Orozco-Cárdenas and Ryan 1999). The roots were observed and photographed under a



154 stereoscopic microscope Stemi SV6, Carl Zeiss (Germany), with a digital camera Canon  
155 (China).

156

### 157 **2.3. Plant arsenic concentration**

158 Peanut leaves and roots were dried at 50 °C for 72 h, homogenized with a mortar and pestle,  
159 and digested under acidic conditions in pressurized chromatographic vials (4 ml capacity) sealed  
160 with polytetrafluoroethylene stoppers. A 0.1 g aliquot of the dried powder was mixed with 1 ml  
161 digestion reagent (H<sub>2</sub>O:HNO<sub>3</sub>:H<sub>2</sub>O<sub>2</sub> (5:3:2)) in closed glass vials, using an autoclave  
162 (Presoclave-75 Selecta, Barcelona, Spain) at 120°C and 1.5 atm for 30 min (Sobrino-Plata et al.  
163 2009). The digests were filtered through a PVDF filter and diluted in milli-RO water to 5 ml. As  
164 concentration was measured using inductively coupled plasma mass spectroscopy (ICP-MS)  
165 NexION 300 (Perkin-Elmer Sciex, San Jose, CA, USA).

166

167

### 168 **2.4. Stress indexes**

169 NADPH oxidase activity was determined spectrophotometrically by NBT reduction at 560 nm  
170 (Sagi and Fluhr 2001). Roots were homogenized according to Sobrino-Plata et al. (2009)  
171 procedure. The activity was measured in 1 ml reaction buffer containing: 0.02 mg protein extract,  
172 0.5 mg ml<sup>-1</sup> NBT, 0.2 mM NADPH, 4 mM CaCl<sub>2</sub> and 0.2 mM MgCl<sub>2</sub>. One unit of NADPH oxidase  
173 was defined as the quantity of enzyme needed to reduced 1 μmol NADPH min<sup>-1</sup>.

174 Hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) content was measured spectrophotometrically after reaction of root  
175 extracts with potassium iodure (KI) following the procedure described by Alexieva et al. (2001).

176 The amount of H<sub>2</sub>O<sub>2</sub> was calculated using a standard calibration curve prepared with known  
177 H<sub>2</sub>O<sub>2</sub> concentrations.

178 Lipid peroxidation was analyzed in peanut roots by determining the concentration of

179 thiobarbituric-reactive substances (TBARS) as described by Heath and Packer (1968). TBARS

180 were quantified by measuring absorbance at 535 nm, corrected for non-specific turbidity by  
181 subtracting the absorbance at 600 nm, using a UV-visible light spectrophotometer (Spectronic®  
182 Genesys 2, USA).

183

## 184 **2.5. Enzymatic assay**

185 The superoxide dismutase (SOD, EC 1.15.1.1) and catalase (CAT, EC 1.11.1.6) root samples  
186 were obtained by homogenization of roots with an extraction buffer containing 50 mM potassium  
187 phosphate, pH 7.8, 0.5 mM EDTA, insoluble polyvinylpyrrolidone, and 0.5% (v/v) Triton X-100  
188 after centrifugation at 10,000 g for 10 minutes at 4°C. GPX/PRX (EC 1.11.1.9), GR (EC 1.6.4.2)  
189 and GST (EC 2.5.1.18) root samples were obtained by homogenization of roots with an  
190 extraction buffer containing 100 mM potassium phosphate, pH 7.4, 1 mM EDTA after  
191 centrifugation at 10,000 g for 10 minutes at 4°C. The supernatants were used to determine  
192 enzymatic activities. Total protein content was assessed according to Bradford (1976) using  
193 bovine serum albumin as standard.

194 SOD activity was determined according to Beauchamp and Fridovich (1973) by measuring the  
195 ability of the enzyme extract to inhibit the photochemical reduction of NBT in the presence of  
196 riboflavin. SOD specific activity was determined spectrophotometrically at 560 nm and  
197 expressed as units  $\text{mg}^{-1}$  protein. One unit of SOD activity was defined as the amount of enzyme  
198 required to inhibit in a 50% the reduction of NBT. CAT activity was measured using the method  
199 described by Aebi (1984). The assay mixture contained 50 mM phosphate buffer at pH 7.4, 12.5  
200 mM  $\text{H}_2\text{O}_2$  and 100  $\mu\text{g}$  protein extract. The reaction was measured by the  $\text{H}_2\text{O}_2$  decomposition at  
201 240 nm. One unit of CAT was defined as the quantity of enzyme needed to degrade 1 mmol of  
202  $\text{H}_2\text{O}_2$ . GPX/PRX activity was determined as described by Flohé and Günzler (1984). One ml of  
203 the assay mixture contained 50 ml of protein extract, 0.1 M phosphate buffer at pH 7.4, 0.24U  
204 GR and 10mM GSH. Then, 1.5 mM NADPH and 1.5 mM  $\text{H}_2\text{O}_2$  were added. The reaction was

205 measured following NADPH oxidation at 340 nm. One unit of GPX/PRX was defined as the  
206 quantity of enzyme needed to produce 1 mmol NADP<sup>+</sup> min<sup>-1</sup>. This assay likely describes  
207 glutathione or glutaredoxin-dependent thiol peroxidase activity since plants do not contain Se-  
208 dependent GPX. Thus, plant GPX functions as a TRX-linked thiol peroxidase (Navrot et al.  
209 2006). GR activity of root extracts was determined as described by Sheadle and Bassham  
210 (1977): the reduction of GSSG was followed by the NADPH oxidation at 340 nm. One unit of GR  
211 was defined as the quantity of enzyme needed to reduced 1 μmol NADPH min<sup>-1</sup>. Finally, GST  
212 activity was assayed according to Habig et al. (1974), by measuring the conjugation of GSH with  
213 1-chloro-2, 4-dinitrobenzene (CDNB) as substrate, which was measured at 340 nm. One unit of  
214 GST was defined as the quantity of enzyme needed to produce 1 μmol conjugated CDNB-GSH  
215 min<sup>-1</sup>.

216

## 217 **2.6. Analysis of glutathione content**

218 Peanut roots were homogenized in 5 % (w/v) 5-sulfosalicylic acid and centrifuged at 10,000 g for  
219 10 min at 4°C to remove cell debris. The supernatant was used for measuring the total GSH  
220 equivalents content by the method of Anderson (1985) using baker's yeast GR. The reaction  
221 was followed at 412 nm and total GSH-Eq was calculated from a standard curve using GSH (0-  
222 30 μM). The intracellular GSH-Eq content was expressed as nmol GSH g<sup>-1</sup> fresh weight (FW).

223

## 224 **2.7. Statistical analysis**

225 Experiments were conducted in a completely randomised design and repeated three times. The  
226 data were analyzed using ANOVA and Duncan's test at P<0.05. Prior to the test of significance,  
227 the normality and homogeneity of variance were verified using the modified Shapiro-Wilk and  
228 Levene tests, respectively.

229

### 230 3. Results

#### 231 3.1. Arsenate-induced peanut growth inhibition and root architecture alteration

232 Arsenic addition caused plant growth alterations as revealed by some morphological features as  
233 the diminution of root volume and altered architecture (Fig. 1). Exposure to 20  $\mu\text{M}$  As(V) led to  
234 limited proliferation of secondary roots (Fig. 1c), while at 100  $\mu\text{M}$  As(V) root growth was severely  
235 inhibited and suffered an evident darkening possibly associated with an induction of oxidative  
236 stress (Fig. 1d). These apparent changes were confirmed by measuring of growth variables.  
237 Peanut root dry weight (DW) was significantly decreased at 6  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$  however, shoot DW  
238 only showed a reduction at the higher arsenate dose evaluated (100  $\mu\text{M}$ ) compared to control.  
239 At 20  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$  root and shoot length was significantly decreased (Table 1). Roots  
240 accumulated the largest proportion of As(V) in plants, reaching values of two to three orders of  
241 magnitude in compare to shoots (Table 2). At 100  $\mu\text{M}$  As(V), the amount of metalloid in root had  
242 no significant difference with those plants exposed to 20  $\mu\text{M}$  As(V), however, a significant  
243 increase in As concentration of shoot was observed at the higher As concentration tested (Table  
244 2). The histological cuts of roots showed three well-developed tissue systems at the epidermis,  
245 the cortex and the central cylinder (Fig. 2), where we could observe small deposits of unknown  
246 amorphous materials on the cell walls of epidermal cells of plants exposed to 6  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$   
247 (Fig. 2b). These deposits were more numerous in roots of plants treated with higher As(V) doses  
248 (20 and 100  $\mu\text{M}$ ), where thicker cell walls were also found (Fig. 2c,d). In addition, there were  
249 several epidermal and parenchymal cells of the cortex that suffered disintegration and rupture at  
250 the highest dose indicating acute cellular damage (Fig. 2d).

251

252

#### 253 3.2. ROS production and oxidative damage caused by arsenate

254 Oxidative stress parameters were determined in peanut roots, the organ that accumulated the  
255 largest proportion of As. Supply of 20  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$  led to the subsequent induction of NADPH

256 oxidase activity as the metalloid dose was increased, with the maximum activity at 100  $\mu\text{M}$  (Fig.  
257 3a). The same response was observed in SOD activity (Fig. 3b). In parallel,  $\text{H}_2\text{O}_2$  concentration  
258 rose at doses over 20  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$ , reaching levels approximately 2-fold higher than in control  
259 roots (Fig. 3d).  $\text{H}_2\text{O}_2$  content significantly enhanced CAT activity (Fig. 3c). Finally, ROS  
260 production induced to lipid peroxidation (Fig. 3e), which reflects the onset of oxidative damage  
261 caused by the exposure of peanut roots to  $\text{H}_2\text{AsO}_4^-$ . The accumulation of  $\text{H}_2\text{O}_2$  observed in roots  
262 (Fig. 3b) is consistent with the histochemical detection of  $\text{H}_2\text{O}_2$  production on peanut roots (Fig.  
263 4). Generation of  $\text{H}_2\text{O}_2$  *in situ* was detected by DAB reaction, and there was a noticeable  
264 accumulation of  $\text{H}_2\text{O}_2$  at the epidermis and cortex of plants treated with 20  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$  (see  
265 arrows in Fig. 4). DAB staining was much stronger when  $\text{H}_2\text{AsO}_4^-$  concentration augmented to  
266 100  $\mu\text{M}$  with deposits only in the epidermis cells, probably the thickening of the outer tangential  
267 wall did not allow the colorant entrance to the cell (Fig. 4). A similar pattern was observed with  
268 the histochemical staining of superoxide ( $\text{O}_2^{\bullet-}$ ), which accumulated remarkably at the epidermis  
269 of plants exposed to 6  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$  (Fig. 5). Interestingly,  $\text{O}_2^{\bullet-}$  was also found in cortex cells of  
270 roots under 20  $\mu\text{M}$  dose, while only epidermis was heavily stained when  $\text{H}_2\text{AsO}_4^-$  treatment rose  
271 to 100  $\mu\text{M}$  As, possibly as observed in  $\text{H}_2\text{O}_2$  detection, at this concentration the epidermis cells  
272 were thick and did not allow the colorant entrance consequently, blue precipitates were not  
273 detected (Fig. 5).

### 274 3.3. Arsenate alters thiol metabolism

275 The concentration of GSH increased significantly only in peanut plants exposed to 100  
276  $\mu\text{M H}_2\text{AsO}_4^-$  (Fig. 6a). On the other hand, at this As(V) concentration GPX/PRX and GR  
277 enzymatic activities decreased by almost 50% of control plants (Figs. 6b,c). In contrast, GST  
278 activity augmented in a dose-dependent manner from 6  $\mu\text{M H}_2\text{AsO}_4^-$ , reaching values of almost  
279 10-times bigger at the highest dose of  $\text{H}_2\text{AsO}_4^-$  (Fig. 6d).

280

### 281 4. Discussion

282 Peanut is an annual herbaceous legume plant with a remarkable importance as crop for food  
283 uses. Argentina is one of the major peanut producers in the world and 90% of this production is  
284 located in Córdoba province. In this region, there is a serious health problem since  
285 approximately 90% of the groundwater is naturally contaminated with As (Francisca et al., 2006).  
286 This water is used as artificial irrigation of crops or can be directly absorbed by roots. Therefore,  
287 contamination of crops with heavy metals and metalloids constitute not only an agronomic  
288 problem but also a human health issue due the transference along the chain food. The  
289 experiments performed in this manuscript describe the oxidative stress parameters of peanut  
290 plants exposed to As(V) as well as the antioxidant response evoked by the metalloid. The doses  
291 used are equivalent to those found in contaminated groundwater in Córdoba province (Cabrera  
292 et al., 2005) and a concentration above those mentioned. Arsenic is a non-essential element  
293 that is highly toxic even at relatively low concentrations and causes growth inhibition (Stoeva  
294 and Bineva 2003; Sobkowiak and Dekert 2003; Alaoui-Sossé et al. 2004; Finnegan and Chen  
295 2012). In our experiments, the strongest growth inhibitory effect was observed in roots, where  
296 dry weight and length were significantly decreased above 6 and 20  $\mu\text{M H}_2\text{AsO}_4^-$ , respectively.  
297 Reduction in plant growth was also observed in *Phaseolus vulgaris*, *Pisum sativum*, *Lupinus*  
298 *albus* L. exposed to large doses of As(V) (Liu et al. 2005; Vázquez Reina et al. 2005;  
299 Bundschuh et al. 2012; Talukdar, 2013). Interestingly, at realistic environmental doses of

300  $\text{H}_2\text{AsO}_4^-$  (above 3  $\mu\text{M}$ ) there was also a significant reduction in root growth of pea seedlings,  
301 which matched with inhibition of cell mitosis and cytokinesis (Dho et al. 2010). In fact, As has  
302 potent genotoxic effects that causes aberrant chromosomal arrangement and cell cycle arrest,  
303 resulting in poor root meristematic activity (Wu et al. 2010). Moreover, root growth inhibition  
304 caused by As(V) was associated with the onset of oxidative stress in rice seedlings (Shri et al.  
305 2009). This can explain the darkening of the peanut root system as arsenic doses increased, in  
306 the same manner as there was this blackish response in white lupin roots treated with at 100  $\mu\text{M}$   
307 As (Vázquez et al. 2008). The evolution of alterations observed in peanut plant is in agreement  
308 with As accumulation trend in roots and shoots. The metalloid content in roots was higher than  
309 in shoots, showing that more than 97% of the As taken up by peanut plants was retained in the  
310 root system. This is a common response of non-hyperaccumulator plants where roots  
311 accumulate the major metalloid concentration (Dhankher et al. 2006; Carbonell-Barrachina et al.  
312 2009; Sobrino-Plata et al. 2013). Probably, this toxic level of contaminant inhibited peanut  
313 growth, metabolism and transport allowing the passage of the metalloid to shoots disrupting the  
314 root protective “barrier” against As(V). It is feasible that the retention of As in roots occurs mainly  
315 at the cell walls of epidermal cells by the formation of an iron plaque as a product of the  
316 oxidation of the ferrous iron to ferric iron, that is precipitated as iron (Fe) onto the root surface,  
317 developing the plaque (Armstrong 1967). This plaque could acts as a barrier to the uptake of  
318 some potentially harmful metals or metalloids (Hansel et al 2002; Liu et al 2004). At the root  
319 endodermis As can be accumulated in vacuoles bound probably with GSH-derived ligands  
320 (Moore et al. 2011). The alterations of root anatomy, morphology and structure observed in  
321 peanut plants also occurred in plants grown with toxic metals like Cd (Vitória et al. 2001;  
322 Rodríguez-Serrano et al. 2006; Groppa et al. 2008; Gallego et al. 2012; Bianucci et al. 2012,  
323 2013). Furthermore, in this work, plants growing with As showed a deposit of unknown material  
324 over the epidermis and cortex cells. Considering that lateral root emergency begins at pericycle  
325 level, the amount of these unknown deposits in roots, that augmented as  $\text{H}_2\text{AsO}_4^-$  concentration

326 increased, could be limiting lateral root development, maintaining an invariable root:shoot ratio  
327 at all  $\text{H}_2\text{AsO}_4^-$  dose tested, but modifying root architecture. This depends not only on metalloid  
328 toxicity but in probable changes in phytohormone balances, as may occur with cytokinins  
329 (Mohan et al., 2016). Similar deposits were detected in peanut roots under Cd stress (Bianucci  
330 et al. 2012; 2013). Thus, the deposition of a material on peanut epidermis induced by As could  
331 be the response of the dermal tissue preventing  $\text{H}_2\text{AsO}_4^-$  entrance in the plant.

332 Previous research demonstrated that As induces generation of ROS leading to oxidative stress  
333 in plants (Hartley-Whitaker et al. 2001; Mascher et al. 2002; Molina et al. 2008; Shri et al. 2009;  
334 Finnegan and Chen 2012). ROS overproduction is the most harmful process that contributes to  
335 cellular damage however, the mechanisms by which As(V) causes the ROS burst are yet  
336 unknown (Sharma 2012). Among different putative sources of ROS production in plants, plasma  
337 membrane NADPH oxidase enzymatic activity constitutes an important one. This plasma  
338 membrane enzyme is involved on  $\text{O}_2^-$  production which is rapidly disrupted to  $\text{H}_2\text{O}_2$ ,  
339 spontaneously or *via* enzymatic reactions (Torres and Dangl 2005), and is known to be activated  
340 under metal stress (Hernández et al. 2015). The enzymatic reaction by which  $\text{O}_2^-$  is disrupted to  
341  $\text{H}_2\text{O}_2$  is catalyzed by SOD, therefore is considered the first enzyme to cope with oxidative stress.  
342 Then, the conversion of  $\text{H}_2\text{O}_2$  to  $\text{H}_2\text{O}$  is accomplished in peroxisomes by CAT enzyme (Noctor  
343 and Foyer 1998). Several studies demonstrated that SOD and CAT activities are enhanced by  
344 arsenate (Srivastava et al. 2005; 2010; Geng et al. 2006; Singh et al. 2007; Duman et al. 2010;  
345 Rai et al. 2011) as observed in this work.

346 Our findings revealed that NADPH oxidase was significantly enhanced by metalloid addition  
347 moreover,  $\text{O}_2^-$  was found in cortex cells of peanut roots under 20  $\mu\text{M}$  As(V), while epidermis  
348 was heavily stained when  $\text{H}_2\text{AsO}_4^-$  treatment rose to 100  $\mu\text{M}$ . The epidermis comprises cells  
349 that are readily exposed to toxic metals and metalloids, and are thought to be the first to respond  
350 to the induced stress (Hernández et al. 2015). Some metal trigger a quick accumulation of ROS  
351 and cell oxidative stress, that can be observed mainly at epidermal cells even after few minutes



352 of treatment with Hg and Cd (Ortega-Villasante et al. 2007). All these findings allowed to confirm  
353 that As induced oxidative stress in young peanut plants and not only reduced growth but also  
354 modified the architecture of the root system. Finally, the accumulation of ROS also causes the  
355 induction of lipid peroxidation (Cho and Park 2000; Hartley-Whitaker et al. 2001), which explains  
356 the strong level observed in peanut roots at the highest dose of  $\text{H}_2\text{AsO}_4^-$ .

357 GSH is the most abundant intracellular non-protein thiol in all living cells and acts as an  
358 important antioxidant to ameliorate excessive ROS generated by different stresses (Millar et al.  
359 2003; Foyer and Noctor 2005; Shao et al. 2008). This thiol is also involved in several  
360 mechanisms of As detoxification in cells, and has a special interest due its ability to detoxify  
361 xenobiotics and heavy metals by conjugation (Jozefczak et al. 2012). Phytochelatins are the  
362 major ligand of metabolized  $\text{H}_2\text{AsO}_4^-$ , which forms an array of different As-PCs complexes in  
363 different plant species (Moore et al. 2011), which are synthesized by the condensation of GSH  
364 (Sobrino-Plata et al. 2014). This imposes a demand of GSH to generate sufficient PCs to limit  
365 the amount of toxic As in plant cells, leading to transient GSH accumulation (Hernández et al.  
366 2015). Therefore, this mechanism may explain the significant increase in GSH concentration  
367 detected in peanut roots exposed to the highest dose of  $\text{H}_2\text{AsO}_4^-$ .

368 One of the hypothesized mechanisms of ROS accumulation under metal stress is the inhibition  
369 of antioxidant enzymatic activity (Sharma and Dietz 2009), particularly when stress symptoms  
370 are strong enough (Gratão et al. 2005). In this sense, we found a significant inhibition of  
371 GPX/PRX and GR activities only in plants treated with the higher arsenate dose. As-sensitive *B.*  
372 *juncea* genotype TM-4 suffered also from GPX/PRX and GR activity inhibition, upon exposure to  
373 acute toxic conditions (Srivastava et al. 2010). However, contradictory responses were observed  
374 in roots of *P. sativum* and *Oryza sativa* L. exposed to Cd and As, respectively, where the  
375 GPX/PRX activity was decreased and GR activity was significantly augmented (Dixit et al. 2001;  
376 Shri et al. 2009; Kumar et al. 2014); implying alterations specific to toxic metal(loid)s. On the  
377 opposite, GST activity rose remarkably at increasing concentrations of  $\text{H}_2\text{AsO}_4^-$  in peanut-

378 treated plants. Similar induction of GST activity occurred in *Oryza sativa* and *Prosopis* sp. plants  
379 exposed to As (Mokgalaka-Matlala et al. 2009; Tripathi et al. 2012; Singh et al. 2016). GST  
380 comprise a large family of isoenzymes involved in the detoxification of organic peroxides,  
381 radicals and xenobiotics at the expense of GSH, and are induced by Cd and  $\text{H}_2\text{AsO}_4^-$  (Schröder  
382 et al. 2009). GST deplete GSH cytosolic content directly to reduce  $\text{H}_2\text{O}_2$  and to detoxify oxidation  
383 subproducts, promoting oxidizing conditions (Anjum et al. 2012). It is feasible that the pool of  
384 GSH will cover the demand imposed for GST activity and allow peanut plant to grow up to  $20\ \mu\text{M}$   
385  $\text{H}_2\text{AsO}_4^-$  with minor observable symptoms of toxicity.

386 Taken together, the results presented in this manuscript revealed that  $20\ \mu\text{M}$  As(V) was the  
387 maximum concentration that allowed growth even when induced an overproduction of ROS in  
388 peanut plants with a relative tolerable As accumulation in shoots. Furthermore, root NADPH  
389 oxidase activity and lipid peroxidation contents increased after this As(V) supply, meanwhile  
390 GST enzyme activity was activated at  $6\ \mu\text{M}$   $\text{H}_2\text{AsO}_4^-$ . This phenomenon shows the ability of  
391 peanut GST enzyme to be induced and stimulate free metal binding before any ROS  
392 overproduction or oxidative damage to macromolecules can be detected. In this sense, it is  
393 important to highlight that  $20\ \mu\text{M}$  of arsenate supply resembles to As(V) levels found in  
394 contaminated Córdoba groundwater zones. Therefore, this finding contribute to understand how  
395 low doses of As could generate contamination of crops, affecting not only economically the  
396 peanut production but also generating a contamination of the food chain.

397

## 398 **5. Conclusion and perspectives**

399 This is the first report of As(V) impact on peanut plant which demonstrates that physiological and  
400 biochemical alterations are induced by the metalloid being NADPH oxidase enzyme implicated  
401 in the oxidative burst. Furthermore, the most sensitive stress parameter was the GSH related  
402 enzyme GST, which could be used in biotechnological approaches to select cultivars with lower

403 toxic symptoms, more tolerant and able to exclude As(V) from edible parts of the plants in order  
404 to reduce As contamination.

405

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413

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**Fig. 1.** Alteration of plant architecture of peanut plants treated with different concentrations of arsenate ( $\mu\text{M}$ ): (a) 0 (control), (b) 6, (c) 20 and (d) 100.

**Fig. 2.** Cross section of peanut root treated with (a) 0 (Control), (b) 6, (c) 20, and (d) 100  $\mu\text{M}$  arsenate. c: cortex, e: epidermis, en: endodermis, ma: unknown material

**Fig. 3.** Effects of arsenate exposure on root (a) NADPH oxidase activity, (b) SOD activity, (c) CAT activity, (d) hydrogen peroxide concentration, (e) lipid peroxidation. Data represent the mean  $\pm$  SE ( $n = 10$ ). Different letters denote significant statistical differences ( $P < 0.05$ ) according to Duncan's test.

**Fig. 4.** Distribution of hydrogen peroxide in peanut root sections treated with 0 (control), 6, 20 and 100  $\mu\text{M}$  arsenate. The arrows indicate the dark deposits resulting from the reaction of  $\text{H}_2\text{O}_2$  with DAB.

**Fig. 5.** Distribution of superoxide radical in peanut root sections treated with 0 (control), 6, 20 and 100  $\mu\text{M}$  arsenate. The arrows indicate reduced formazan blue deposits caused by the reduction of NBT with  $\text{O}_2^{\cdot-}$ .

**Fig. 6.** Changes in GSH content and thiol-related enzyme activities in peanut roots in response to 0 (control), 6, 20 and 100  $\mu\text{M}$  arsenate. (a) Total GSH-Eq concentration, (b) GPX/PRX activity, (c) GR activity, and (d) GST activity. Data represent the mean  $\pm$  SE ( $n = 10$ ). Different letters indicate significant statistical differences ( $P < 0.05$ ) according to Duncan's test.

**Table 1.** Penaut plant growth treated with different doses of arsenate.

As addition ( $\mu\text{M}$ )	Shoot Dry weight (g)	Root Dry weight (g)	Shoot Length (cm)	Root Length (cm)
0	$0.38 \pm 0.02^{\text{B}}$	$0.14 \pm 0.01^{\text{C}}$	$9.46 \pm 0.20^{\text{C}}$	$13.44 \pm 0.57^{\text{C}}$
6	$0.41 \pm 0.01^{\text{B}}$	$0.11 \pm 0.00^{\text{B}}$	$9.65 \pm 0.28^{\text{C}}$	$12.13 \pm 0.34^{\text{C}}$
20	$0.37 \pm 0.01^{\text{B}}$	$0.09 \pm 0.01^{\text{B}}$	$8.38 \pm 0.26^{\text{B}}$	$10.20 \pm 0.28^{\text{B}}$
100	$0.14 \pm 0.01^{\text{A}}$	$0.05 \pm 0.01^{\text{A}}$	$5.04 \pm 0.32^{\text{A}}$	$6.15 \pm 0.36^{\text{A}}$

Data represent the mean  $\pm$  SE (n =10). Different superscript letters indicate significant statistical differences between treatments ( $P < 0.05$ ) according to the Duncan's test.

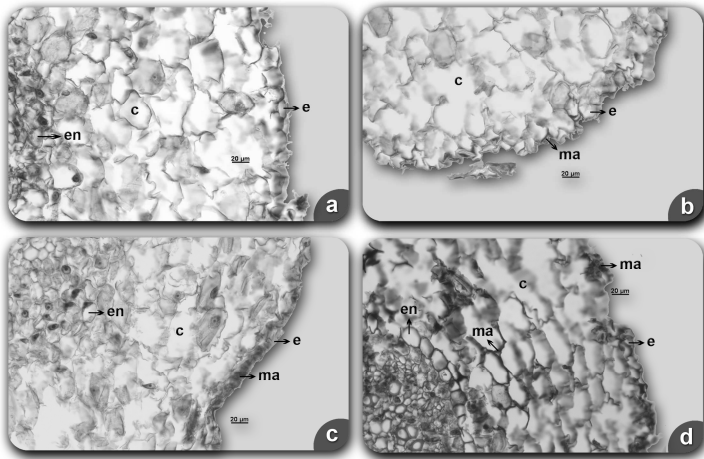
**Table 2.** Concentration of As in peanut plants treated with different doses of arsenate.

As treatments ( $\mu\text{M}$ )	As concentration ( $\mu\text{g g}^{-1}$ dry weight)	
	Shoot	Root
6	$0.28 \pm 0.05^{\text{B}}$	$149.40 \pm 25.87^{\text{B}}$
20	$0.66 \pm 0.12^{\text{B}}$	$520.20 \pm 40.45^{\text{A}}$
100	$7.80 \pm 0.62^{\text{A}}$	$538.60 \pm 84.79^{\text{A}}$

Data represent the mean  $\pm$  SE (n =10). Different superscript letters indicate significant statistical differences between treatments (P <0.05) according to the Duncan's test.



ACCEPTED MANUSCRIPT



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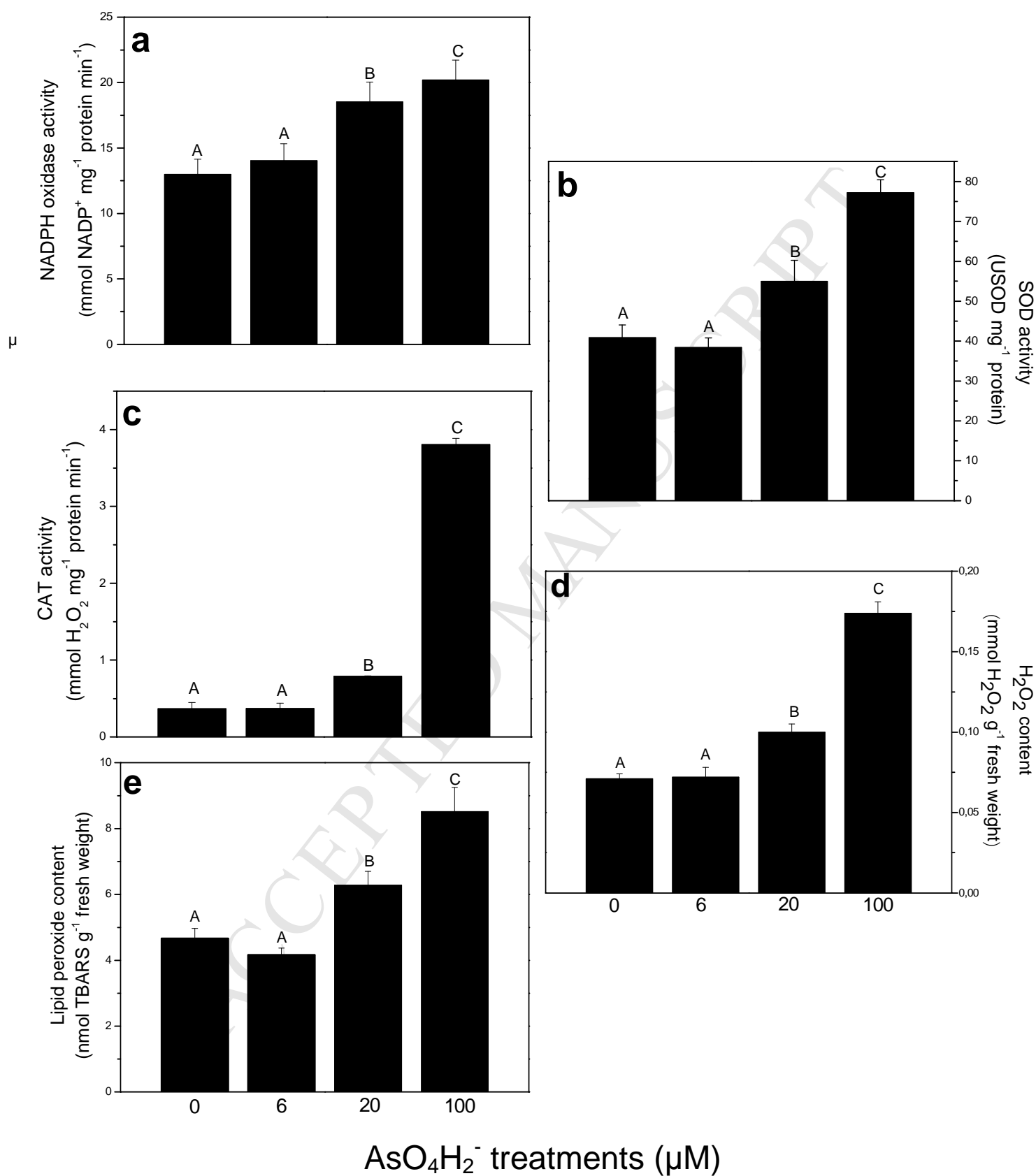
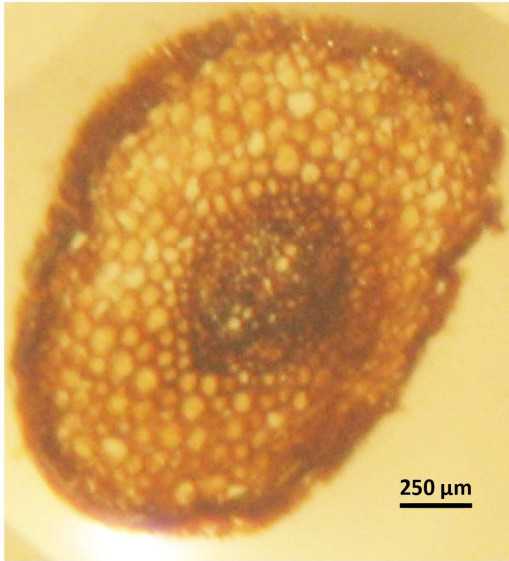
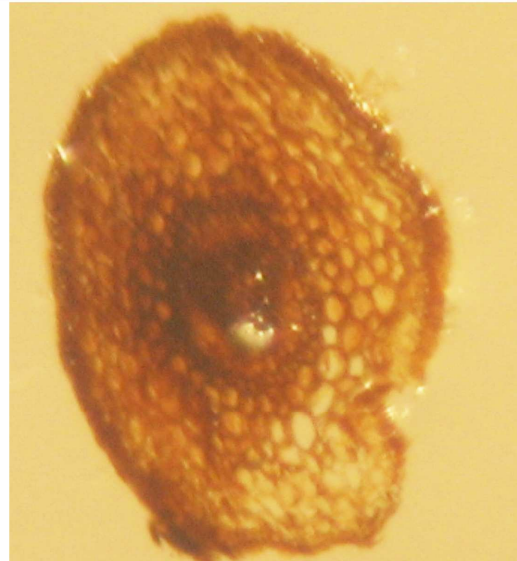
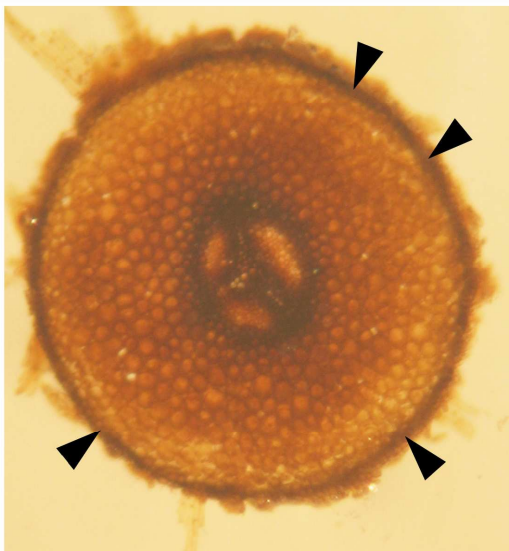
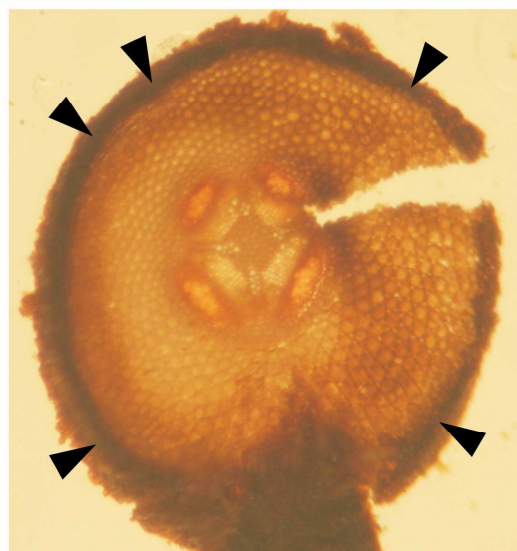
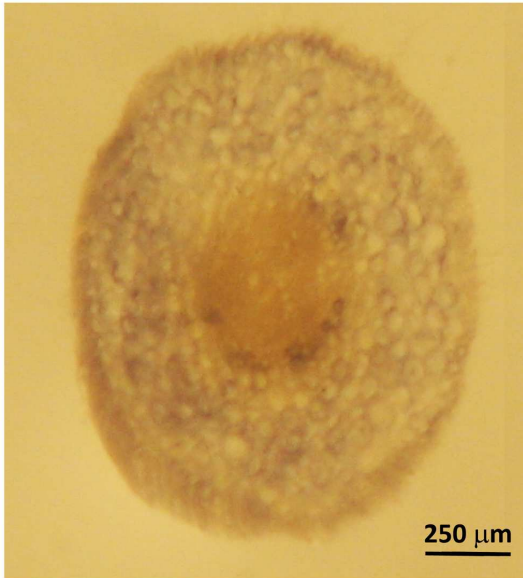
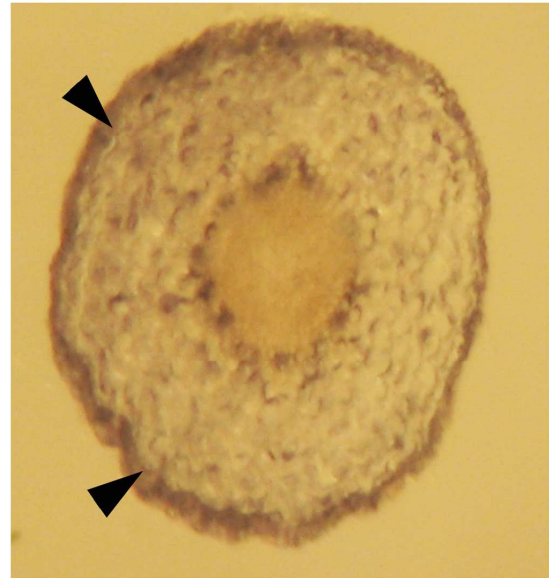
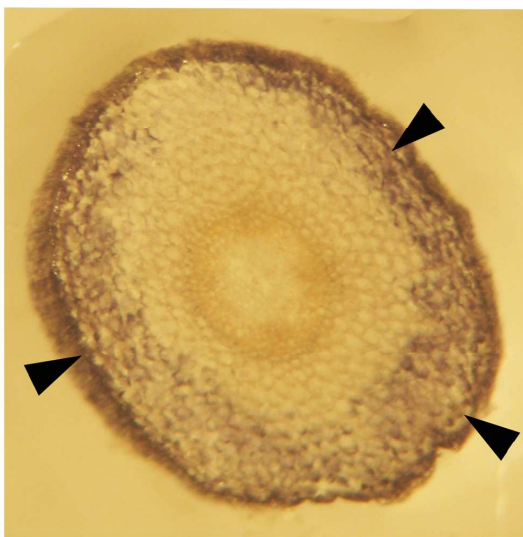
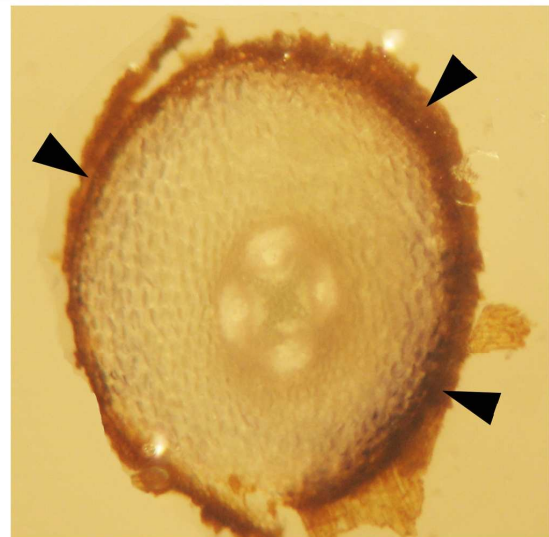


Fig 3.

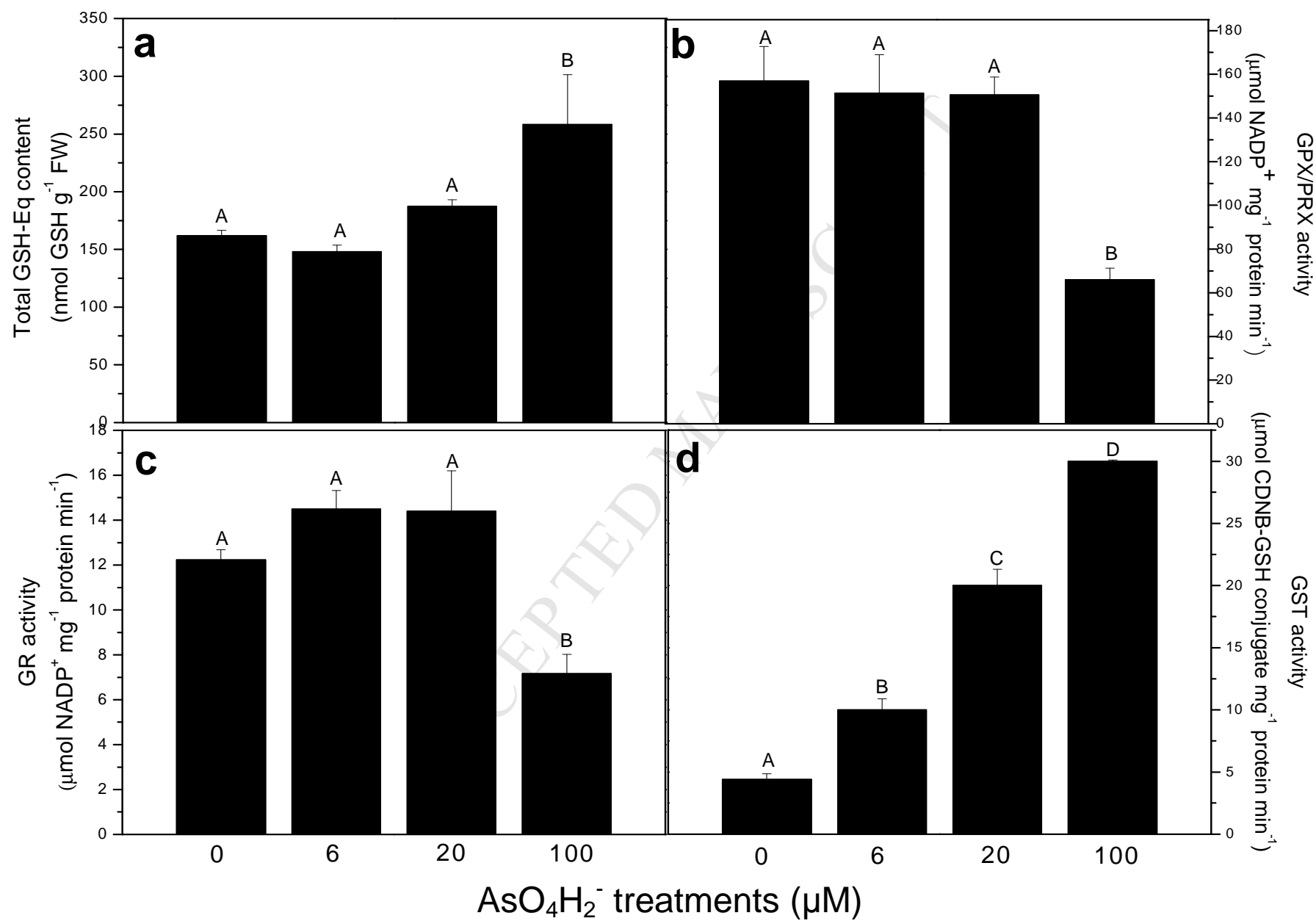


**Control****6  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$** **20  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$** **100  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$** 

ACC

**Control****6  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$** **20  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$** **100  $\mu\text{M}$   $\text{H}_2\text{AsO}_4^-$** 

ACC



### Highlights

This is the first report showing the effect of arsenic on peanut plant in hydroponics

Root histological structure modification and damage to macromolecules are induced by As

Glutathione S-Transferase exhibits a remarkable increase upon exposure to As

GST activity is proposed as a suitable biomarker of peanut arsenic contamination