

# Benthic diversity and assemblage structure of a north Patagonian rocky shore: a monitoring legacy of the NaGISA project

M.E. RECHIMONT<sup>1</sup>, D.E. GALVÁN<sup>2</sup>, M.C. SUEIRO<sup>3</sup>, G. CASAS<sup>4</sup>, M.L. PIRIZ<sup>4</sup>, M.E. DIEZ<sup>5</sup>, M. PRIMOST<sup>1</sup>, M.S. ZABALA<sup>1</sup>, F. MÁRQUEZ<sup>6</sup>, M. BROGGER<sup>7</sup>, J.E.F. ALFAYA<sup>1</sup> AND G. BIGATTI<sup>1</sup>

<sup>1</sup>Laboratorio de Reproducción y Biología Integrativa de Invertebrados Marinos (LARBIM), CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina, <sup>2</sup>Laboratorio de Ecología y Manejo de Peces de Arrecife, CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina, <sup>3</sup>Grupo de Ecología en Ambientes Costeros, CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina, <sup>4</sup>Laboratorio de Algas Marinas, Grupo de Ecología en Ambientes Costeros (GEAC), CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina, <sup>5</sup>Laboratorio de Parasitología, CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina, <sup>6</sup>Biología evolutiva de moluscos marinos, CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina, <sup>7</sup>Laboratorio de Ecosistemas Costeros, MACN–CONICET, Museo Argentino de Ciencias Naturales ‘Bernardino Rivadavia’, Avenida Ángel Gallardo 470, C1405DJR, Buenos Aires, Argentina

*The rocky shore of Punta Este, Golfo Nuevo (Patagonia, Argentina), was sampled by means of the standardized NaGISA (CoML) protocol, that was aimed to generate biodiversity baseline data in six levels, from high intertidal to 10 m depth. Based on the generated data, we analysed the benthic assemblage structure, species richness, mean abundance and the distribution pattern of invertebrate functional groups, typifying species in each intertidal and subtidal level. The intertidal sampled is exposed to extreme physical conditions higher than any other rocky shore system studied, with air temperature variation of 40°C during the year, maximum winds of 90 km/h and semidiurnal tides of 5 m amplitude; on the other hand subtidal presents less thermal variation ( $\Delta T$  10°C throughout the year) and more homogeneous physical conditions. We identified 64 taxa represented by six animal phyla: Mollusca, Arthropoda, Annelida (Polychaeta), Echinodermata, Cnidaria and Nemertea; and three algal phyla: Chlorophyta, Rhodophyta and Heterokontophyta (class: Phaeophyceae). Ordination non-metric multidimensional scaling plots showed three different assemblages in terms of species composition (intertidal, subtidal 1 m level and subtidal 5–10 m levels). The intertidal was represented by suspension feeders, mainly Mollusca. The tiny mussels *Brachidontes rodriguezii* and *Brachidontes purpuratus*, and the algae *Corallina officinalis* dominated the intertidal and promote positive interaction (amelioration), preventing other species from the extreme physical stress. The subtidal was mainly represented by grazers. Our results showed a tendency of suspension feeders to decrease and grazers and predators to increase from high intertidal to subtidal, probably driven by decreasing physical stress. The gastropod *Tegula patagonica*, the sea urchins *Arbacia dufresnii* and *Pseudechinus magellanicus* and the invasive algae *Undaria pinnatifida* were the most abundant species in 1 m–10 m levels. Based on previous work performed in the region, we hypothesize that the differences registered between intertidal and subtidal levels could be explained in part by an increase in physical stress at the intertidal, with low predation pressure that promotes positive interactions, while in the subtidal the increase in consumers and decrease of physical stress could lead to associational defences. Our results could be useful as baseline data to develop a sustainable network for long-term monitoring of benthic community changes due to anthropogenic activities.*

**Keywords:** biodiversity, rocky shores, Patagonia, physical stress, Census of Marine Life

Submitted 22 March 2013; accepted 1 July 2013; first published online 13 August 2013

## INTRODUCTION

Making inventories and monitoring biodiversity are crucial to identifying the natural processes and the human activities that affect ecosystems. The loss of biodiversity should be of considerable concern to everyone because it threatens the functioning of a system (Rigby *et al.*, 2007). There is compelling

evidence that climate is changing and this change is driving important alterations in the composition and structure of a very diverse array of natural assemblages: terrestrial, marine and aquatic (Cruz-Mota *et al.*, 2010). Destruction of natural habitats, loss of strong interacting species (such as habitat formers and keystone species), introduction of non-native species (Orensanz *et al.*, 2002), over-exploitation of natural resources, such as overfishing (Jackson *et al.*, 2001), urban pollution (Díaz *et al.*, 2002) and climate change (Smith *et al.*, 2006) are all potential threats to biodiversity. The first step towards understanding these complexities, on the large scale, is to investigate global patterns of biodiversity (i.e.

**Corresponding author:**

G. Bigatti

Email: gbigatti@cenpat.edu.ar

taxonomic composition and diversity of assemblages) over a broad range of spatial scales, and to develop a sustainable network for long-term monitoring (Rigby *et al.*, 2000). To detect changes in natural communities, and unequivocally relate them to anthropogenic impacts or climate disruptions, proper baseline data are of utmost importance (Cruz-Mota *et al.*, 2010), as well as to have knowledge about the processes governing the patterns observed locally in rocky intertidal habitats (Underwood, 2000). In this sense, it is important to perform experiments to test hypotheses about the processes that structure and maintain diversity in such communities. Alternatively, standardized global monitoring programmes can be put in place to assess changes in biodiversity and relate those changes to possible anthropogenic causes (Darling & Côté, 2008).

In the particular case of coastal ecosystems, studies on benthic community structure are an adequate tool to assess the impact caused by pollution (Díaz *et al.*, 2002), mainly due to the reduced mobility of benthic species, which cannot avoid the potentially dangerous environmental conditions. For instance, on the southern Atlantic shores, the structure of the intertidal community characterized by the tiny mussel *Brachidontes rodriguezii*, has been successfully used to detect environmental impact due to domestic sewage (López Gappa *et al.*, 1990, 1993; Vallarino, 2002; Vallarino *et al.*, 2002). Therefore, baseline information on benthic community structure is critical to avoid confounding the consequences of an impact, given the high level of variability in natural assemblages (Smith & Simpson, 1995; Gelin *et al.*, 2003). Several authors have pointed out that pre-impact data are essential to understand effects of pollution such as an oil spill (Paine *et al.*, 1996). The problem is that the databases available at the moment are restricted to a few regions and locations throughout the world (Cruz-Motta *et al.*, 2010). Despite the ecological and economic importance of Patagonian coasts, its benthic biodiversity remains poorly understood, except for some records in the literature (Olivier *et al.*, 1966; Bertness *et al.*, 2006; Cuevas *et al.*,

2006; Kelaher *et al.*, 2007; Torres & Caille, 2009; Irigoyen *et al.*, 2010; Miloslavich *et al.*, 2011; Silliman *et al.*, 2011; Sueiro *et al.*, 2011). In this context, the rocky shore of Punta Este (Golfo Nuevo, Patagonia) was selected as one of the sites distributed along the world for the Census of Marine Life (CoML), an international science programme for inventorying and monitoring marine biodiversity all over the world (finished in 2010). The main goal of this project was to establish an initial census for future comparisons and monitoring programmes over time (Iken & Konar, 2003). Through the combination of a simple, cost-efficient and intentionally low-tech sampling protocol and a hierarchically nested design, the project NaGISA (Natural Geography in Shore Areas), carried out a series of globally-distributed standard transects in more than 250 sites within 28 countries. At present, the NaGISA database, growing since 2003, is available through the Ocean Biogeographic Information System (OBIS) and contains nearly 54,666 records for 2867 taxa (Plos collections, 2011). Based on the data generated by the NaGISA project, the aim of the present study was directed to describe the benthic assemblages in terms of richness, specific and functional composition at different levels (intertidal and subtidal) of Punta Este rocky shore, as baseline data that could be useful for future temporal monitoring programmes, allowing to detect changes in biodiversity or ecosystem function due to anthropogenic impacts or environmental changes.

## MATERIALS AND METHODS

### Study area

The rocky shore of Punta Este ( $42^{\circ}47'5.7''S$   $64^{\circ}57'7.3''W$ ) is ~8 km from the city of Puerto Madryn, located in northeast Patagonia, Argentina and on the west coast of Golfo Nuevo (Figure 1). The region is characterized by extreme weather conditions. A characteristic of the Patagonian climate is the

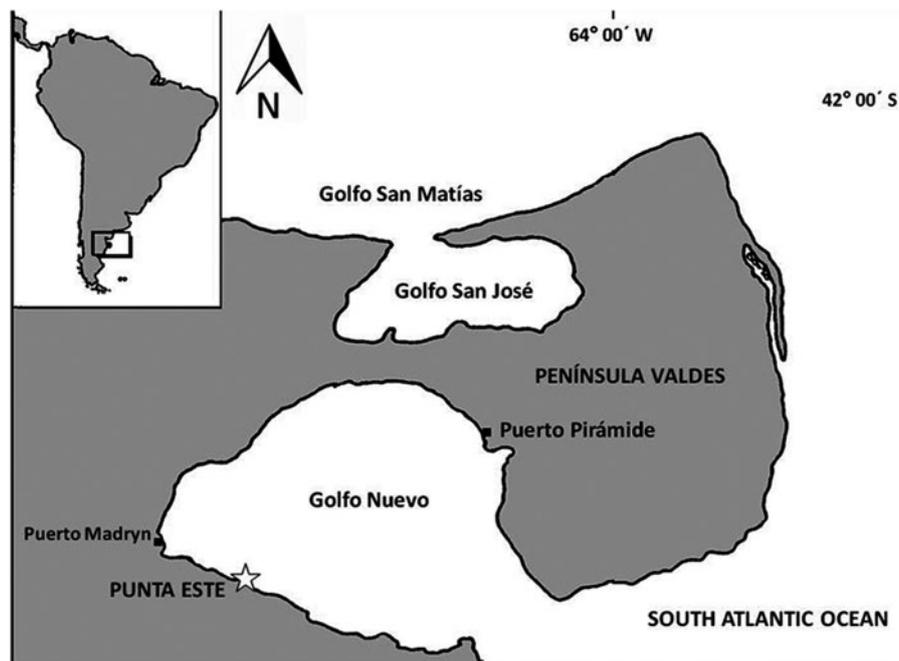


Fig. 1. Sampling site (star) at Punta Este, Golfo Nuevo, Patagonia, Argentina.

**Table 1.** Mean values ( $\pm$ SD) of abundance (ind.m<sup>-2</sup>), total abundance (ind.m<sup>-2</sup>), diversity (H) and taxonomic richness (S) of invertebrates found in each level sampled.

<b>Invertebrate taxa</b>							
<b>Mollusca</b>	<b>FG</b>	<b>HT</b>	<b>MT</b>	<b>LT</b>	<b>1 m</b>	<b>5 m</b>	<b>10 m</b>
<i>Brachidontes purpuratus/Brachidontes rodriguezii</i>	SF	24640 $\pm$ 17991.6	16,115.2 $\pm$ 9234	6752 $\pm$ 4092.8	688 $\pm$ 1406.8		
<i>Lithophaga patagonica</i>	SF	3.2 $\pm$ 7.1	76.8 $\pm$ 97.5	22.4 $\pm$ 14.3	140.8 $\pm$ 254.3	41.6 $\pm$ 76.3	
<i>Mytilus edulis</i>	SF	41.6 $\pm$ 44.6	48 $\pm$ 107.3				
<i>Aulacomya atra atra</i>	SF		3.2 $\pm$ 7.1	28.8 $\pm$ 64.3			
<i>Tegula patagonica</i>	G	3.2 $\pm$ 7.1	28.8 $\pm$ 42.9	76.8 $\pm$ 92.8	278.4 $\pm$ 164.5	352 $\pm$ 234	80 $\pm$ 45.2
<i>Trophon geversianus</i>	PS	19.2 $\pm$ 20.8	25.6 $\pm$ 48.7	9.6 $\pm$ 14.3	9.6 $\pm$ 8.7	3.2 $\pm$ 7.1	
<i>Epitonium frabrizioi</i>	PS	12.8 $\pm$ 17.5				3.2 $\pm$ 7.1	
<i>Siphonaria lessoni</i>	G	480 $\pm$ 398.3	140.8 $\pm$ 55.8	32 $\pm$ 55.4			
<i>Fissurella radiosa tixierae</i>	G	6.4 $\pm$ 14.3	25.6 $\pm$ 57.2	32 $\pm$ 46.6	9.6 $\pm$ 21.4		
<i>Plaxiphora aurata aurata</i>	G	12.8 $\pm$ 28.6	35.2 $\pm$ 70.1	51.2 $\pm$ 26.2	3.2 $\pm$ 7.1		
<i>Chaetopleura isabellei</i>	G	3.2 $\pm$ 7.1					
<b>Arthropoda</b>							
<i>Balanus glandula</i>	SF	105.6 $\pm$ 88	19.2 $\pm$ 20.8	3.2 $\pm$ 7.1			
<i>Exosphaeroma</i> sp.	DF	28.8 $\pm$ 64.3	105.6 $\pm$ 193.1	105.6 $\pm$ 80.4	96 $\pm$ 105.5	12.8 $\pm$ 20.8	
<i>Amphitoe valida</i>	DF		3.2 $\pm$ 7.1	3.2 $\pm$ 7.1			
<i>Tanais dulongii</i>	DF		102.4 $\pm$ 228.9	6.4 $\pm$ 14.3		35.2 $\pm$ 48.5	16 $\pm$ 22.6
<i>Cyrtograpsus angulatus</i>	DF	4.8 $\pm$ 8.6	9.6 $\pm$ 14.3	12.8 $\pm$ 28.6			
<i>Cyrtograpsus altimanus</i>	DF	19.2 $\pm$ 34.6	41.6 $\pm$ 54.9	3.2 $\pm$ 7.1			
<i>Pachycheles chubutensis</i>	DF				28.8 $\pm$ 55.8		
<i>Leucippa pentagona</i>	DF					9.6 $\pm$ 14.3	
<b>Annelida</b>							
<i>Platynereis</i> sp.	AFG	6.4 $\pm$ 14.3	44.8 $\pm$ 70.1	6.4 $\pm$ 14.3	144 $\pm$ 202.7	131.2 $\pm$ 88.6	40 $\pm$ 33.9
<i>Syllidia armata</i>	PS	6.4 $\pm$ 14.3	137.6 $\pm$ 307.6	3.2 $\pm$ 7.1		28.8 $\pm$ 28.6	24 $\pm$ 33.9
<i>Harmothoe</i> sp.	PS			6.4 $\pm$ 8.7	3.2 $\pm$ 7.1	22.4 $\pm$ 8.7	
<i>Lepidasthenia</i> sp.	PS			9.6 $\pm$ 8.7	3.2 $\pm$ 7.1		
<i>Terebellidae</i> sp.2	DF	3.2 $\pm$ 7.1	6.4 $\pm$ 14.3	6.4 $\pm$ 8.7	3.2 $\pm$ 7.1		
<i>Terebellidae</i> sp.1	DF			3.2 $\pm$ 7.1	3.2 $\pm$ 7.1	6.4 $\pm$ 14.3	
<i>Marphysa aenea</i>	PS	3.2 $\pm$ 7.1	16 $\pm$ 35.7	38.4 $\pm$ 60.5	6.4 $\pm$ 14.3		
<i>Eulalia</i> sp.	PS	9.6 $\pm$ 21.4	25.6 $\pm$ 29				
<b>Echinodermata</b>							
<i>Ophioplocus januarii</i>	SF		3.2 $\pm$ 7.1		3.2 $\pm$ 7.1	9.6 $\pm$ 8.7	16 $\pm$ 22.6
<i>Allostichaster capensis</i>	PS				6.4 $\pm$ 8.7	3.2 $\pm$ 7.1	
<i>Pseudechinus magellanicus</i>	O				3.2 $\pm$ 7.1	19.2 $\pm$ 26.2	320 $\pm$ 135.7
<i>Arbacia dufresnii</i>	O		3.2 $\pm$ 7.1		12.8 $\pm$ 13.3	3.2 $\pm$ 7.1	64 $\pm$ 0
<b>Nemertea</b>							
<i>Neolineus</i> sp.	PS	3.2 $\pm$ 7.1	6.4 $\pm$ 8.7	19.2 $\pm$ 17.5			
<i>Zygonemertes</i> sp.	PS		3.2 $\pm$ 7.1	16 $\pm$ 22.6			
<b>Cnidaria</b>							
<i>Parabunodactis imperfecta</i>	PS	16 $\pm$ 22.6	19.2 $\pm$ 28.6	44.8 $\pm$ 17.5	19.2 $\pm$ 28.6		

Continued

Table 1. Continued.

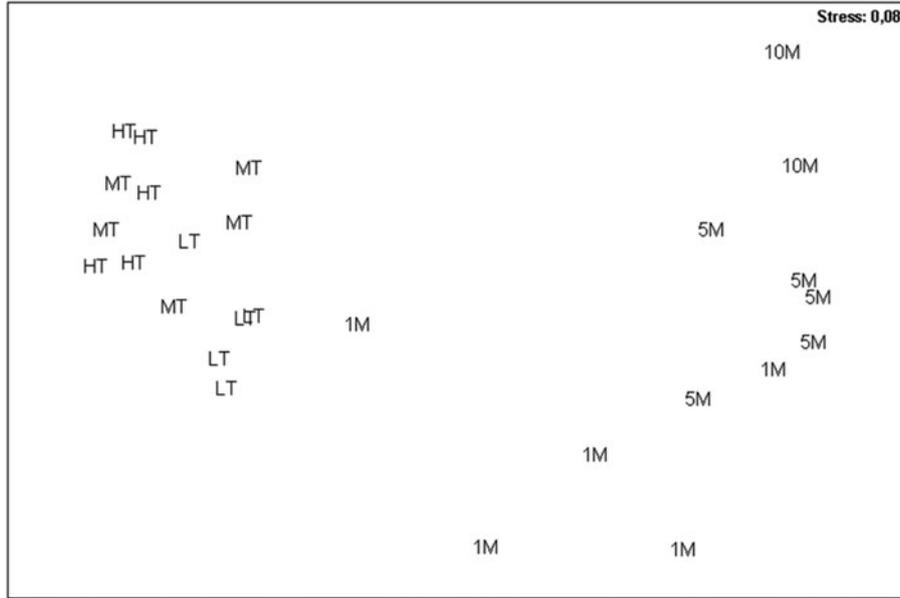
Invertebrate taxa	FG	HT	MT	LT	1 m	5 m	10 m
Mollusca							
Mean Total abundance (N) ± SD	25432 ± 18,115	17062 ± 9,849.2	7309 ± 4156.7	1497 ± 1445.7	761 ± 206.6	600 ± 79.2	
Mean Diversity (H) ± SD	0.2 ± 0.1	0.3 ± 0.2	0.4 ± 0.1	1.2 ± 0.4	1.3 ± 0.4	1.3 ± 0.2	
Mean Taxonomic richness (S) ± SD	8 ± 1.4	10.4 ± 5.5	12.4 ± 3.2	9.2 ± 4.6	7.8 ± 2.1	2.4 ± 0	

Levels are indicated as follows: high (HT), middle (MT) and low (LT) intertidal and 1 m depth (1 m), 5 m depth (5 m) and 10 m depth (10 m). The functional group (FG) of each taxon is indicated as: suspension feeder (SF), deposit–detritus feeder (DF), grazer (G), predator–scavenger (PS), opportunist (O) and all functional groups (AFG).

predominance of winds from the west. Westerly winds are characterized not only by the persistence during the year but also by their intensity (Paruelo *et al.*, 1998), recording an annual mean speed of 16.6 km/h and reaching up to 90 km/h (Labraga & De Davies, updated 2013). Also low humidity content characterizes this wind, so most of the region belongs to the arid category (Paruelo *et al.*, 1998). The mean annual precipitation over the region is 235.9 mm (Labraga & De Davies, updated 2013) and the mean annual temperature is 13.4°C with annual amplitude reaching 40°C and daily variations of up to 30°C. These strong, dry winds combined with low rainfall, make the Patagonian intertidal the place with the highest desiccation stress recorded for rocky shore communities (Bertness *et al.*, 2006). The tide regime is semidiurnal in the zone and the mean amplitude is 3.8 m, but spring tides can reach 5.7 m. Coastal shoreline is predominantly of consolidated limestone, locally known as toasca. Subtidal shallow waters present less thermal variation than the intertidal, recording surface water temperature of 18°C at the end of the summer and minimum of 8°C in spring (Rivas & Beier, 1990), and more homogeneous environmental conditions resulting in less physical stress.

### Sampling protocol

Samples were collected during November 2006 at low tide, following the NaGISA protocol (see Rigby *et al.*, 2007 for further details). Six levels were sampled: the high (hereafter HT); middle (hereafter MT) and low intertidal zone (hereafter LT); and 1 m depth (hereafter 1 m) (only exposed during spring tides); 5 m depth (hereafter 5 m); and 10 m depth (hereafter 10 m) in the subtidal zone. The levels in the intertidal zone were selected noting patterns in the distribution coverage of the most conspicuous benthic invertebrates and algae and topography that determines different exposure to water. The HT was defined where barnacles and tiny mussels were abundant; this is the zone most exposed to desiccation during low tides. The MT was defined as the tidal zone covered by a compact tiny mussel bed. The LT was defined as the zone that was dominated by dense coralline algal cover and above the low tide limit. Subtidal levels were selected by measuring depth with a digital depth-meter during low tide (1 m, 5 m, 10 m). For invertebrates samples were taken randomly at each level by five 25 × 25 cm quadrats along a 30 m transect parallel to the shoreline. All organisms were carefully removed from the substratum with a scraper, placed into mesh bags, fixed in 10% formalin and later transferred to 70% ethanol. At the laboratory each sample was sorted through a sieve of 0.5 mm mesh. Organisms retained were identified to the lowest possible taxonomic level and counted to calculate abundance ( $\text{ind.m}^{-2}$ ). Each organism from each level was assigned to a functional group according to their feeding type: suspension feeder (SF); deposit–detritus feeder (DF); predator–scavenger (PS); grazer (G); and opportunist (O). A voucher of each specimen collected was deposited in the CENPAT Invertebrate Collection with numbers CNP\_INV\_0101 to 0136. The mytilids *Brachidontes rodriguezii* and *Brachidontes purpuratus*, the latter historically placed into the genus *Perumitylus*, form a ‘mussel matrix’ and were counted and grouped together under the name of ‘tiny mussels’ because of their morphological similarities, high phenotypic variability (Rechimont, 2011; Trovant *et al.*, 2013) and shell damage during sampling procedure that



**Fig. 2.** Non-metric multidimensional scaling ordination for the invertebrate assemblages in the different levels sampled in the shore of Punta Este based on Bray–Curtis similarities and square root transformed data. HT (empty diamond); MT (empty square); LT (empty circle); 1 m (solid diamond); 5 m (solid circle) and 10 m (solid square).

make it difficult to discriminate between species. In the 10 m depth only two samples were possible on hard substrate due to the randomness of the sampling and the characteristics of hard substrates in the studied area, so soft bottoms were presented in the other three quadrats, and were sampled following the NaGISA protocol for hard bottoms as in the other quadrats.

Algae samples were collected following the same protocol and routine (Rigby *et al.*, 2007), but using a 50 × 50 cm quadrat. Algae were removed, placed in plastic bags and stored in a cooler until they were frozen at −18°C. Algae were identified in the laboratory to the lowest possible taxonomic level. Dry weight of each species or higher taxa were measured to the nearest 0.01 g in a Sartorius analytical balance after stove samples at 60°C for about 24 h or up to the weight was constant. Because of the precision of the balance samples with less than 0.01 g dry weight were not detected. Algae were not present in the 10 m level, due to the quadrats falling onto sandy bottoms or bare hard substrate.

**Data analysis**

Multivariate approaches were used to examine invertebrate assemblages in the six different levels. In order to investigate differences in assemblage structure the one-way analysis of similarity (ANOSIM) procedure was used (Clarke, 1993). A graphical comparison of assemblage structure between levels was produced by ordination of the data using two-dimensional non-metric multidimensional scaling (nMDS, Clarke, 1993). In order to identify taxa that dominated or characterized the levels sampled, a similarity percentage analysis (SIMPER) was performed (Clarke, 1993). Data were fourth-root transformed and the Bray–Curtis coefficient was used as a resemblance parameter (Clarke, 1993).

In addition, taxonomic richness, diversity (Shannon’s index H), and total abundance (ind.m<sup>−2</sup>) were compared

between intertidal and subtidal levels using the non-parametric Kruskal–Wallis test because variances were heterogeneous and could not be stabilized after different transformations.

**RESULTS**

**Invertebrate samples**

Among invertebrates, a total of 16,344 individuals were counted encompassing 35 species or higher taxa. In terms of species, the dominant phyla were molluscs (12 species), arthropods (eight species) and annelids (eight species). Also the presence of echinoderms (four species), nemertean (two species) and cnidarians (one specie) were detected (Table 1). Throughout all levels sampled in the intertidal zone (HT,

**Table 2.** Summary of ANOSIM results for Punta Este shore: (P<0.001 for all tests). R Global for level 0.6.

Test	R statistic
LT, MT	0.348
LT, HT	0.764
LT, 1 m	0.588
LT, 5 m	1
LT, 10 m	1
MT, HT	0.084
MT, 1 m	0.796
MT, 5 m	1
MT, 10 m	1
HT, 1 m	0.896
HT, 5 m	1
HT, 10 m	1
1 m, 5 m,	0.280
1 m, 10 m	0.745
5 m, 10 m	0.564

MT, LT), as well as in 1 m samples, the tiny mussels *Brachidontes rodriguezii* and *Brachidontes purpuratus* were dominant, their abundances being several orders of magnitude higher than those accounted for in the rest of the sampled taxa (Table 1). Although the non-native barnacle *Balanus glandula* was observed in all levels at the intertidal zone, its highest abundance was recorded on HT samples (Table 1). Other taxa showing high abundances at intertidal samples were the gastropod *Siphonaria lessoni*, the isopod *Exosphaeroma* sp., the tanaid *Tanais dulongii* and the polychaete *Syllidia armata* (Table 1). At the subtidal level, 1 m samples, showed high abundances of the bivalve *Lithophaga patagonica* while the gastropod *Tegula patagonica*, and the polychaete *Platynereis* sp. showed to be abundant at all

subtidal levels (Table 1). Finally, echinoderms were almost exclusively present in the subtidal samples (Table 1).

The multivariate analysis indicated that the assemblage composition of benthic invertebrates was different among levels (Figure 2; Table 2). We found significant differences between all pairs of samples from intertidal and subtidal zones and in the case of the intertidal, samples from HT showed to be different from samples from LT, while in the case of subtidal, samples from 1 m differed from 10 m samples ( $P < 0.001$ ; Table 2).

The SIMPER analysis showed the characteristic taxa from each level (Table 3). While the mussels *Brachidontes rodriguezii* and *Brachidontes purpuratus* characterized intertidal levels, the gastropod *Tegula patagonica* was characteristic from 1 m

**Table 3.** Summary of SIMPER results for Punta Este shore: average abundance (ind.m<sup>-2</sup>) of typifying species in each level, their contribution (%) to the within-group similarity, and cumulative total (%) of contributions (90% cut-off).

	Av. Abundance	Contribution %	Cumulative %
<b>HT (59.98)</b>			
<i>Brachidontes purpuratus</i> – <i>Brachidontes rodriguezii</i>	24,640	56.76	56.76
<i>Siphonaria lessoni</i>	480	21.7	78.46
<i>Balanus glandula</i>	105	9.94	88.4
<i>Mytilus edulis</i>	42	4.54	92.95
<b>MT (51.38)</b>			
<i>Brachidontes purpuratus</i> – <i>Brachidontes rodriguezii</i>	16,115	65.06	65.06
<i>Siphonaria lessoni</i>	140	18.31	75.73
<i>Eulalia</i> sp.	25	4.81	80.54
<i>Exosphaeroma</i> sp.	105	3.5	84.04
<i>Lithophaga patagonica</i>	76	3.46	87.49
<i>Cyrtograpsus altimanus</i>	41	3.06	90.55
<b>LT (61.54)</b>			
<i>Brachidontes purpuratus</i> – <i>Brachidontes rodriguezii</i>	6752	36.45	36.45
<i>Plaxiphora aurata aurata</i>	51	10.97	47.42
<i>Parabunodactis imperfecta</i>	44	10.81	58.23
<i>Exosphaeroma</i> sp.	105	8.44	66.67
<i>Tegula patagonica</i>	76	6.61	73.28
Sipunculid uncertain	28	5.87	79.15
<i>Lithophaga patagonica</i>	22	5.87	85.02
<i>Neolineus</i> sp.	19	2.85	87.87
<i>Fissurella radiosa tixierae</i>	32	2.62	90.49
<b>1 m (43.86)</b>			
<i>Tegula patagonica</i>	278	34.17	34.17
<i>Exosphaeroma</i> sp.	96	13.48	47.65
<i>Platynereis</i> sp.	144	12.80	60.46
<i>Brachidontes purpuratus</i> – <i>Brachidontes rodriguezii</i>	688	12.09	72.55
<i>Lithophaga patagonica</i>	140	5.64	78.19
Sipunculid uncertain	25	5.51	83.7
<i>Arbacia dufresnii</i>	12	4.49	88.19
<i>Trophon geversianus</i>	9	4.44	92.63
<b>5 m (55.96)</b>			
<i>Tegula patagonica</i>	352	31.31	31.31
<i>Platynereis</i> sp.	131	25.88	57.19
<i>Harmothoe</i> sp.	22	17.41	74.59
<i>Syllidia armata</i>	28	5.61	80.2
<i>Tanais dulongii</i>	35	5.38	85.59
<i>Ophioplocus januarii</i>	9	4.97	90.56
<b>10 m (64.49)</b>			
<i>Pseudechinus magellanicus</i>	320	34.15	34.15
<i>Arbacia dufresnii</i>	64	24.97	59.11
<i>Tegula patagonica</i>	80	23.23	82.35
<i>Platynereis</i> sp.	40	17.65	100

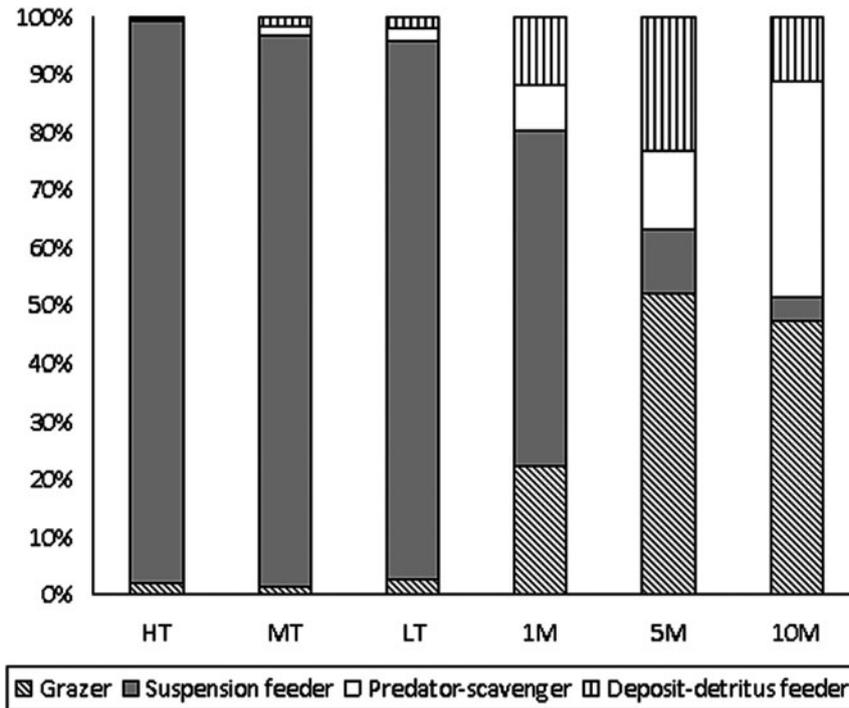


Fig. 3. Percentage proportion of the different feeding types present in each sampled level.

Table 4. Mean values ( $\pm$ SD) of biomass (g of dry weight.m<sup>-2</sup>) and non-detected weights (n/d; less than 0.01 g of dry weight.m<sup>-2</sup>) of algae found in each level sampled.

Algae taxa	HT	MT	LT	1 m	5 m
<i>Anotrichium furcellatum</i> (J. Agardh) Baldock					n/d
<i>Aphanocladia robusta</i> Pujals	n/d	n/d		n/d	n/d
<i>Bossiella orbigniana</i> (Decaisne) P.C. Silva				n/d	
<i>Callithamnion gaudichaudii</i> C. Agardh					n/d
<i>Ceramium strictum</i> Harvey		n/d			
<i>Ceramium virgatum</i> Roth	8.8 $\pm$ 6.9	5 $\pm$ 11.2	n/d	n/d	0.1 $\pm$ 0.3
<i>Cladophora falklandica</i> (J.D. Hooker and Harvey)		1.2 $\pm$ 2.4	n/d		
<i>Cladostephus spongiosus</i> (Hudson) C. Agardh				0.4 $\pm$ 0.8	
<i>Codium vermilara</i> (Olivieri) Delle Chiaje				2.8 $\pm$ 5	81.8 $\pm$ 119.1
<i>Colpomenia sinuosa</i> (Mertens ex Roth) Derbès and Solier		0.06 $\pm$ 0.09	5.1 $\pm$ 11.41		
<i>Corallina officinalis</i> Linnaeus	254.99 $\pm$ 209.84	402.85 $\pm$ 371.75	76.15 $\pm$ 109.14	6.5 $\pm$ 13.83	
<i>Litophyllum</i> sp.				n/d	n/d
<i>Lithothamnion</i> sp.	0.13 $\pm$ 0.3	0.35 $\pm$ 0.56		0.3 $\pm$ 0.68	3.68 $\pm$ 5.22
<i>Dictyota dichotoma</i> (Hudson) J.V. Lamouroux	0.12 $\pm$ 0.24	1.35 $\pm$ 1.61	4.46 $\pm$ 7.55	14.64 $\pm$ 17.86	23.08 $\pm$ 18.88
<i>Ectocarpus</i> sp.		0.71 $\pm$ 1.59			
<i>Gracilaria gracilis</i> (Stackhouse) M. Steentoft, L.M. Irvine and W.F. Farnham				n/d	n/d
<i>Halopteris funicularis</i> (Montagne) Sauvageau				n/d	n/d
<i>Heterosiphonia merenia</i> Falkenberg				n/d	n/d
<i>Hymenena laciniata</i> (J.D. Hooker and Harvey) Kylin		0.09 $\pm$ 0.19		0.74 $\pm$ 1.06	4.63 $\pm$ 5.13
<i>Leathesia difformis</i> Areschoug		n/d	n/d		
<i>Polysiphonia abscissa</i> J.D. Hooker and Harvey			n/d		
<i>Polysiphonia brodiei</i> (Dillwyn) Sprengel		n/d			
<i>Polysiphonia harveyi</i> J.W. Bailey	0.22 $\pm$ 0.22	0.24 $\pm$ 0.49	0.15 $\pm$ 0.34	0.05 $\pm$ 0.09	n/d
<i>Sphacelaria</i> sp.					n/d
<i>Sphacelaria rigidula</i> Kützing		n/d			
<i>Streblocladia camptoclada</i> (Montagne) Falkenberg	n/d				n/d
<i>Ulva rigida</i> C. Agardh	0.35 $\pm$ 0.79	n/d			
<i>Ulva</i> sp.	4.9 $\pm$ 10.78	0.11 $\pm$ 0.24		1.85 $\pm$ 2.55	4 $\pm$ 2.22
<i>Undaria pinnatifida</i> (Harvey) Suringar			10.81 $\pm$ 11.93	24.16 $\pm$ 38.93	442.74 $\pm$ 273.74

Levels are indicated as follows: high (HT), middle (MT) and low (LT) intertidal and 1 m depth (1 m) and 5 m depth (5 m).

and 5 m samples and the sea urchin *Pseudechinus magellanicus* was from 10 m samples.

Diversity and total abundance differed significantly between intertidal and subtidal levels, while subtidal samples were more diverse than intertidal samples (Kruskal–Wallis test  $H = 20.61$ ,  $P \leq 0.001$ ; Table 1), intertidal showed higher total abundances (Kruskal–Wallis test  $H = 21.05$ ,  $P \leq 0.001$ ; Table 1). On the other hand, no differences were observed for species richness between intertidal and subtidal samples (Kruskal–Wallis test  $H = 5.133787$ ,  $P = 0.399$ ).

Based on the mean abundances of each feeding type, the intertidal level was clearly dominated by suspension feeders while in 1 m samples this feeding group showed the same percentage as that of grazers. On the other hand, 5 m and 10 m samples were dominated by grazers and predators (Figure 3). There was a tendency of decreasing suspension feeders species and increasing of grazers and predators from intertidal to subtidal level.

### Algae samples

From the algae samples, 29 species or higher taxa were identified and weighted (Table 4). *Corallina officinalis* was the main algae species in the intertidal level with the higher biomass (HT:  $254.99 \text{ g.m}^{-2}$ , MT:  $402.85 \text{ g.m}^{-2}$ , LT:  $76.17 \text{ g.m}^{-2}$ ) (Table 4). This alga forms a matrix that retains humidity and is related to a great abundance of invertebrates (see Table 1). In the LT level the invasive algae *Undaria pinnatifida* was present with a biomass of  $24.16 \text{ g.m}^{-2}$ . The 1 m level was characterized by *Dictyota dichotoma* with  $14.64 \text{ g.m}^{-2}$  and *U. pinnatifida* (Table 4). Despite *D. dichotoma* presented high biomass in the 5 m level also ( $23.08 \text{ g.m}^{-2}$ ), *Codium vermilara* showed the highest biomass of native algae in this level ( $81.80 \text{ g.m}^{-2}$ ) (Table 4). Additionally, in the 5 m level the biomass of *U. pinnatifida* was elevated, with values of  $442.74 \text{ g.m}^{-2}$ . At the 10 m level no algae were registered as quadrats fell onto soft bottom or bare hard substrate.

### DISCUSSION

The rocky shore of Punta Este, selected as one of the sites distributed along the world for the Census of Marine Life, showed a high benthic biodiversity which encompassed 64 taxa of invertebrates and algae. Given the characteristics of the standardized sampling protocol used, it could be highlighted that the composition of invertebrate assemblages showed a vertical zonation where the greatest differences were observed between intertidal and subtidal samples. In this sense, it was found that the high intertidal was dominated by the tiny mussels *Brachidontes purpuratus* and *Brachidontes rodriguezii*, forming a conspicuous mussel bed (mussel matrix), while in the lowest levels the tiny mussel bed disappeared and the zone is dominated by a mixture of bigger mussels and the coralline alga *Corallina officinalis*. Rocky intertidal communities on Argentinean coasts are structured by the dominance of mytilids, and particularly by *B. rodriguezii* in the warm temperate and *B. purpuratus* in the cold-temperate sector, coexisting in the same beds along the transition zone ( $41\text{--}43^\circ\text{S}$ ) (Penchaszadeh, 1973; Adami et al., 2008; Trovant et al., 2013; Van der Molen et al., 2013). These mussel beds are critical in the structure and function of the intertidal community because light, temperature,

wave action and potentially lethal desiccation stress are significantly reduced within them, while sediment loading and relative humidity increase (Seed & Suchanek, 1992; Silliman et al., 2011). Although diversity may decline on the rocky surface due to mussel dominance, the development of the beds creates a secondary substratum with enriched infaunal and epibenthic assemblages (Dittman & Robles, 1991; Lohse, 1993; Ragnarsson & Raffaelli, 1999). Likewise, the turf-forming coralline alga also provides a more benign habitat for the recruitment and survival of several invertebrate species (Kelaher et al., 2007). Due to the high abundance of the tiny mussels registered, the intertidal zone was characterized by suspension-feeders. This kind of feeding is favoured in this environment due to the plankton abundance in the waters around the rocky shores and the wave movement that contributes to the nutrient circulation (Bertness, 1999).

With regard to subtidal samples, the 1 m level showed a merged invertebrate assemblage with both components of LT and 5 m samples. Even though this level was characterized by the grazer gastropod *Tegula patagonica* and that all feeding types were present, the suspension feeders remained dominant. On the other hand, 5 m and 10 m samples clearly differed in composition assemblages from intertidal. At subtidal zone the tiny mussels suspension feeders were not present, and that zone was dominated by grazers and predator–scavenger. That dominance of grazers may be associated with the macroalgae, composed mainly of the invasive species *Undaria pinnatifida*. The dominance of herbivore gastropods in Punta Este are concordant to some other mid to high-latitude assemblages (from LMEs: Beaufort Sea and Gulf of Alaska), which represent more than 90% within the gastropods; such dominance in herbivores with high densities could be related to high macroalgal biomass (Miloslavich et al., in press). Irigoyen et al. (2010) have found that the presence of *U. pinnatifida* is associated with increased invertebrates species richness and diversity, and could potentially produce a bottom-up effect on local food chains by increasing the abundance of prey for a wide range of predators. An example are the sea *Arbacia dufresnii* and *Pseudechinus magellanicus* that fed on the invasive kelp *U. pinnatifida*, whereas the snail *T. patagonica* scraped off biofouling adhered to its surface (Teso et al., 2009). We observed this tendency in our results, so the increase of biodiversity in the subtidal assemblage might be related to the presence of *U. pinnatifida* and less environmental physical stress.

Differences on assemblage's composition, together with the higher total abundance and the lower diversity of the intertidal with regard to subtidal zone is likely to be the result of a trade-off between the individual physiological tolerance with the relative contribution of positive and negative interaction operating. In this sense, it has been proposed that under harsh physical conditions, positive interactions should be the main forces structuring the assemblage while as physical conditions decrease competitive interactions should be dominant structuring forces (Bertness & Callaway, 1994). Previous experimental researches on Patagonian rocky shores have demonstrated that: (1) the thick mussel matrix protects associated animals from lethal, wind-driven desiccation stress; and (2) maintenance of the whole-community diversity depends exclusively on facilitation by foundation species, rather than keystone predation, under the intense climate stress registered (Bertness et al., 2006; Silliman et al., 2011).

Experimental research is required to have a clear picture of the mechanism underlying assemblage patterns on Punta Este rocky shores. However, it is likely that the differences registered between intertidal and subtidal assemblages could be explained in part by an increase in physical stress at the intertidal with low predation pressure that promotes positive interactions (amelioration of stress by founding species, such as the mytilid matrix), while in the subtidal the increase in consumers and decrease of physical stress could lead to negative interactions, such as predation, that could lead to associational defences.

Our results add to the information generated in the last ten years by the NaGISA project, a series of globally-distributed standard transects in more than 250 sites within 28 countries that produced several papers (Plos collections, 2011). Data about the benthic diversity and assemblage of Punta Este rocky shore could be useful for a long-term monitoring programme to detect community changes due to anthropogenic activities or environmental changes and to test the factors that structure this assemblage. In this sense the recently created South American Research Group in Coastal Ecosystems (SARCE) will continue monitoring marine diversity and biomass in more than 50 localities using a more dynamic and non-invasive protocol than that used in the present work, encouraging the study of the ecosystem function and human impact on South American rocky shores.

## ACKNOWLEDGEMENTS

We are grateful to A. Irigoyen who generously helped in the subtidal sampling, F. Dellatorre and E. Gomez Simes who kindly helped in the identification of decapods and P. Raffo who helped to identify some algae. V. Böekenhans, F. Arrighetti, and J. Giménez helped in the sorting of species. This work was part of the NaGISA project of the Census of Marine Life at Puerto Madryn site.

## FINANCIAL SUPPORT

We thank the Census of Marine Life and the Nagisa project for funding this work.

## REFERENCES

- Adami M.L., Tablado A. and Sodor M.A. (2008) Population dynamics of the intertidal mytilid *Brachidontes rodriguezii* (Bivalvia) on a rocky shore. *Thalassas* 24, 21–27.
- Bertness M.D. (1999) Rocky shores. In *The ecology of Atlantic rocky shorelines*. Sunderland: Sinauer Associates, pp. 177–247.
- Bertness M.D. and Callaway R. (1994) Positive interactions in communities. *Trends in Ecology & Evolution* 9, 191–193.
- Bertness M.D., Crain C.M., Silliman B.R., Bazterrica M.C., Reyna M.V., Hidalgo F. and Farina J.K. (2006) The community structure of western Atlantic Patagonian rocky shores. *Ecological Monographs*, 76, 439–460.
- Clarke K.R. (1993) Non-parametric multivariate analyses of changes in community structure. *Australian Journal of Ecology* 18, 117–143.
- Cruz-Mota J.J., Miloslavich P., Palomo G., Iken K., Konar B., Pohle G., Trott T., Benedetti-Cecchi L., Herrera A., Hernandez A., Sardi A., Bueno J., Castillo E., Klein E., Guerra-Castro E., Gobin J., Gomez D.I., Riosmena-Rodriguez R.I., Mead A., Bigatti G., Knowlton A.L. and Shirayama Y. (2010) Patterns of spatial variation of assemblages associated with intertidal rocky shores: a global perspective. *PLoS ONE* 5(12), e14354.
- Cuevas J.M., Martin J.P. and Bastida R. (2006) Benthic community changes in a patagonian intertidal: a forty years later comparison. *Thalassas* 22, 29–37.
- Darling E. and Côté I. (2008) Quantifying the evidence for ecological synergies. *Ecology Letters* 11, 1278–1286.
- Díaz P., López Gappa J.J. and Piriz M.L. (2002) Symptoms of eutrophication in intertidal macroalgal assemblages of Nuevo Gulf (Patagonia, Argentina). *Botanica Marina* 45, 267–273.
- Dittman D. and Robles C. (1991) Effect of algal epiphytes on the mussel *Mytilus californianus*. *Ecology*, 72, 286–296.
- Gelin A., Gravez V. and Edgar G.J. (2003) Assessment of Jessica oil spill impacts on intertidal invertebrate communities. *Marine Pollution Bulletin* 46, 1377–1384.
- Iken K. and Konar B. (2003) Natural Geography in nearshore areas (NaGISA): the nearshore component of the census of marine life. *Gayana* 67, 153–160.
- Irigoyen A.J., Trobbiani G., Sgarlatta M.P. and Raffo M.P. (2010) Effects of the alien algae *Undaria pinnatifida* (Phaeophyceae, Laminariales) on the diversity and abundance of benthic macrofauna in Golfo Nuevo (Patagonia, Argentina): potential implications for local food webs. *Biological Invasions* 1, 1521–1532.
- Jackson J.B.C., Kirby M.X., Berger W.H., Bjorndal K.A., Botsford L.W., Bourque B.J., Bradbury R., Cooke R., Erlandson J., Estes J.A., Hughes T.P., Kidwell S., Lange C.B., Lenihan H.S., Pandolfi J.M., Peterson C.H., Steneck R.S., Tegner M.J. and Warner R.R. (2001) Historical overfishing and the recent collapse of coastal ecosystems. *Science* 293, 629–637.
- Kelaher B.P., Castilla J.C., Prado L., York P., Schwindt E. and Bortolus A. (2007) Spatial variation in molluscan assemblages from corraline turf of Argentinean Patagonia. *Journal of Molluscan Studies* 73, 139–146.
- Labraga J.C. and De Davies E.C. (updated 2013) *Datos de la estación meteorológica del Centro Nacional Patagónico (CONICET), Puerto Madryn (42°46'S; 65°02'W), Chubut, Argentina*. Available at: <http://www.cenpat.edu.ar/fisicambien/climaPM.htm> (accessed 10 July 2013).
- Lohse D.P. (1993) The importance of secondary substratum in a rocky intertidal community. *Journal of Experimental Marine Biology and Ecology* 166, 1–17.
- López Gappa J.J., Tablado A. and Magaldi N.H. (1990) Influence of sewage pollution on a rocky intertidal community dominated by the mytilid *Brachidontes rodriguezii*. *Marine Ecology Progress Series* 63, 163–175.
- López Gappa J.J., Tablado A. and Magaldi N.H. (1993) Seasonal Changes in an intertidal community affected by sewage pollution. *Environmental Pollution* 82, 157–165.
- Miloslavich P., Klein E., Díaz J.M., Henández C.E., Bigatti G., Campos L., Artigas F., Castillo J., Penchaszadeh P., Eneill P.E., Carranza A., Retana M.V., Díaz de Astarloa J.M., Lewis M., Yorrio P., Piriz M.L., Rodríguez D., Yoneshigue-Valentín Y., Gamboa L. and Martín A. (2011) Marine biodiversity in the Atlantic and Pacific coasts of South America: knowledge and gaps. *PLoS ONE* 6(1), e14631.
- Miloslavich P., Cruz-Motta J.J., Klein E., Iken K., Weinberger V., Konar B., Trott T., Pohle G., Bigatti G., Benedetti-Cecchi L., Shirayama Y., Mead A., Palomo G., Ortiz M., Gobin J., Sardi A., Díaz J.M., Knowlton A., Wong M. and Peralta. A.C. (in press)

- Large-scale spatial distribution patterns of gastropod assemblages in rocky shores. *PLoS ONE*.
- Olivier S., Paternoster I.K. and Bastida R.** (1966) Estudios biocenóticos en las costas de Chubut (Argentina). I. Zonación biocenológica de Puerto Pardelas (Golfo Nuevo). *Instituto de Biología Marina* 10, 1–74.
- Orensanz J.M., Schwindt E., Pastorino G., Bortolus A., Casas G., Darrigran G., Elias P.J., López Gappa J.J., Obenat S., Pascual M., Penchaszadeh P., Piriz M.L., Scarabino F., Spivak E.D. and Vallarino E.A.** (2002) No longer the pristine confines of the world ocean: a survey of exotic marine species in the southwestern Atlantic. *Biological Invasions* 4, 115–143.
- Paine R.T.** (1996) Food web complexity and species diversity. *American Naturalist* 100, 65–75.
- Paruelo J.M., Beltran A., Jobbagy, Sala O.E. and Golluscio R.A.** (1998) The climate of Patagonia: general patterns and controls on biotic processes. *Ecología Austral* 8, 85–101.
- Penchaszadeh P.E.** (1973) Ecología de la comunidad del mejillín (*Brachydontes rodriguezii* d'Orb.) en el mediolitoral rocoso de Mar del Plata (Argentina): el proceso de recolonización. *Physis, Sección A*, 32, 51–64.
- Ragnarsson S.A. and Raffaelli D.** (1999) Effects of the mussel *Mytilus edulis* L. on the invertebrate fauna of sediments. *Journal of Experimental Marine Biology and Ecology* 241, 31–43.
- Rechimont M.E.** (2011) *Patrón de distribución y abundancia de invertebrados bentónicos en costas del Golfo Nuevo*. In Facultad de Ciencias Naturales, Puerto Madryn: Universidad Nacional de la Patagonia. Degree thesis, 69 pp.
- Rigby P.R., Iken K. and Shirayama Y.** (2007) *Sampling biodiversity in coastal communities: NaGISA protocols for seagrass and macroalgal habitats*. Kyoto: Kyoto University Press.
- Rigby P.R., Kato T. and Shirayama Y.** (2000) *An introduction to the Natural Geography In Shore Areas (NaGISA) project*. Kyoto: Seto Marine Biological Laboratory, Kyoto University.
- Rivas A.L. and Beier E.J.** (1990) Temperature and salinity fields in the north patagonic Gulf. *Oceanológica Acta* 13, 15–20.
- Seed R. and Suchanek T.H.** (1992) Population and community ecology of *Mytilus*. In Gosling E.M. (ed.) *The mussel Mytilus*. B.V., Amsterdam: Elsevier Science Publishers, pp. 87–169.
- Silliman B.R., Bertness M.D., Altieri A.H., Griffin J.N., Bazterrica M.C., Hidalgo F.J., Crain C.M. and Reyna M.V.** (2011) Whole-community facilitation regulates biodiversity on Patagonian rocky shores. *PLoS ONE* 6, e24502.
- Smith J.R., Fong P. and Ambrose R.F.** (2006) Dramatic declines in mussel bed community diversity: response to climate change? *Ecology* 87, 1153–1161.
- Smith S.D.A. and Simpson R.D.** (1995) Effects of the 'Nella Dan' oil spill on the fauna of *Durvillaea antarctica* holdfasts. *Marine Ecology Progress Series* 121, 73–89.
- Sueiro M.C., Bortolus A. and Schwindt E.** (2011) Habitat complexity and community composition: relationships between different ecosystem engineers and the associated macroinvertebrate assemblages. *Helgoland Marine Research* 65, 467–477.
- Teso S.V., Bigatti G., Casas G., Piriz M.L. and Penchaszadeh P.** (2009) Do native grazers from Patagonia, Argentina consume the invasive kelp *Undaria pinnatifida*? *Revista del Museo Argentino de Ciencias Naturales* 11, 7–14.
- Torres A. and Caille G.** (2009) The hard-bottom intertidal communities before and after removal of an anthropogenic disturbance: a case study in the coast of Puerto Madryn (Patagonia, Argentina). *Revista de Biología y Oceanografía* 44, 517–521.
- Trovant B., Ruzzante D.E., Basso N.G., and Orensanz J.M.** (2013) Distinctness, phylogenetic relations and biogeography of intertidal mussels (*Brachidontes*, *Mytilidae*) from the south-western Atlantic. *Journal of the Marine Biological Association of the United Kingdom (Firstview)*, 1–13. doi:10.1017/S0025315413000477.
- Vallarino E.A.** (2002) *La comunidad bentónica intermareal de Brachidontes rodriguezii (D'Orb.) y su relación con el efluente cloacal de la ciudad de Mar del Plata (38°S)*. PhD thesis. Universidad de Mar del Plata.
- Vallarino E.A., Rivero M.S., Gravina M.C. and Elías R.** (2002) The community-level response to sewage impact in intertidal mytilid beds of the Southwestern Atlantic and the use of the Shannon index to assess pollution. *Revista de Biología Marina y Oceanografía* 37, 25–33.
- Van der Molen S., Márquez F., Idaszkin Y.L. and Adami M.** (2013) Use of shell-shape to discriminate between *Brachidontes rodriguezii* and *Brachidontes purpuratus* species (*Mytilidae*) in the transition zone of their distributions (south-western Atlantic). *Journal of the Marine Biological Association of the United Kingdom* 93, 803–808.
- and
- Underwood A.J.** (2000) Experimental ecology of rocky intertidal habitats: What are we learning? *Journal of Experimental Marine Biology and Ecology* 250, 51–76.

#### Correspondence should be addressed to:

G. Bigatti  
 Laboratorio de Reproducción y Biología Integrativa de Invertebrados Marinos (LARBIM)  
 CENPAT–CONICET, Bvd Brown 2915, U9120ACD, Puerto Madryn, Chubut, Argentina  
 email: gbigatti@cenpat.edu.ar