Accepted Manuscript

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PII: S0272-7714(17)30168-3

DOI: 10.1016/j.ecss.2017.10.015

Reference: YECSS 5651

To appear in: Estuarine, Coastal and Shelf Science

Received Date: 11 February 2017

Revised Date: 9 October 2017

Accepted Date: 17 October 2017

Please cite this article as: Marina, Tomá.I., Salinas, V., Cordone, G., Campana, G., Moreira, Marí.Eugenia., Deregibus, D., Torre, L., Sahade, R., Tatián, M., Barrera Oro, E., De Troch, M., Doyle, S., Quartino, Marí.Liliana., Saravia, L.A., Momo, F.R., The food web of Potter Cove (Antarctica): Complexity, structure and function, *Estuarine, Coastal and Shelf Science* (2017), doi: 10.1016/ j.ecss.2017.10.015.

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1 The food web of Potter Cove (Antarctica): complexity, structure and function

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22 Abstract

23 Knowledge of the food web structure and complexity are central to better understand ecosystem 24 functioning. A food-web approach includes both species and energy flows among them, 25 providing a natural framework for characterizing species' ecological roles and the mechanisms 26 through which biodiversity influences ecosystem dynamics. Here we present for the first time a 27 high-resolution food web for a marine ecosystem at Potter Cove (northern Antarctic Peninsula). 28 Eleven food web properties were analyzed in order to document network complexity, structure 29 and topology. We found a low linkage density (3.4), connectance (0.04) and omnivory 30 percentage (45), as well as a short path length (1.8) and a low clustering coefficient (0.08). 31 Furthermore, relating the structure of the food web to its dynamics, an exponential degree 32 distribution (in- and out-links) was found. This suggests that the Potter Cove food web may be 33 vulnerable if the most connected species became locally extinct. For two of the three more 34 connected functional groups, competition overlap graphs imply high trophic interaction between 35 demersal fish and niche specialization according to feeding strategies in amphipods. On the other 36 hand, the prey overlap graph shows also that multiple energy pathways of carbon flux exist 37 across benthic and pelagic habitats in the Potter Cove ecosystem. Although alternative food 38 sources might add robustness to the web, network properties (low linkage density, connectance 39 and omnivory) suggest fragility and potential trophic cascade effects. Key words: Ecological networks; Structure; Degree distribution; Marine ecosystem; 40

41 Antarctica.

Food web (FW) characterization is essential to understanding ecology as a way to describe and

43 **1. Introduction**

44

45 quantify the complexity of ecosystems by identifying the trophic interactions among species 46 (Bascompte 2009). The framework of ecological network analysis could also be used to quantify 47 the effects of the environment and how indirect effects of such interactions influence overall 48 ecosystem properties (Brose and Dunne 2009). 49 Since the early 2000s, ecological networks from marine systems have received more attention 50 answering an emphatical call of Raffaelli (2000) for more research on marine webs. In this sense, 51 indices derived from Ecological Network Analysis (ENA), a system-oriented methodology to 52 analyze within system interactions (Fath et al. 2007), have been used to investigate trophic interactions in marine ecosystems (Baird et al. 2007, Ulanowicz 2011, Wuff et al. 2012, 53 54 Heymans et al. 2014). Among marine webs, polar FWs recently began to be considered in the 55 frame of FW theory (e.g. Jacob et al. 2006, Bodini et al. 2009, de Santana et al. 2013). Moreover, 56 some conclusions on the effects of global warming on Arctic and Antarctic marine FWs have 57 been proposed (de Santana et al. 2013, Kortsch et al. 2015). 58 Potter Cove is an Antarctic fjord that suffers from the impact of the high rate of warming 59 occurring in Western Antarctic Peninsula (Quartino et al. 2013, Deregibus et al. 2016). The 60 abundant and rich epibenthic fauna has been changing under the influence of considerable 61 sediment inputs and other effects derived from ice melting (Pasotti et al. 2015a, Sahade et al. 62 2015). The way in which network properties can be modified under climate change is in general, 63 poorly known (Petchey et al. 2010, Walther 2010, Woodward et al. 2010). To understand the 64 community-level consequences of the rapid polar warming, Wirta et al. (2015) suggested that we 65 should turn from analyses of populations, population pairs, and isolated predator-prey couplings

66 to considering all the species interacting within communities. If species affected by perturbations 67 possess key functional roles in the FW, then the potential higher order, indirect effects of those 68 perturbations on the entire FW structure can be dramatic (Kortsch et al. 2015). Knowing that 69 climate change effects are already occurring in Potter Cove ecosystem and that ecosystems 70 respond to perturbations as a whole, a network approach could contribute to a better 71 understanding of changes in the ecosystem's synthetic properties like resilience or stability. A 72 representative roadmap of trophic interactions of Potter Cove will allow testing for the impact of 73 ongoing climate change effects (e.g. glacier retreat, loss of ice shelves, increment of 74 sedimentation input) which might be transmitted throughout the entire ecosystem. 75 Although FW studies use binary webs that indicate the presence of a trophic interaction but do 76 not provide any information on the frequency of the interaction or the rate of biomass flow through the interaction, overlap graphs (e.g. competition and common-enemy graphs), can 77 78 provide information about indirect interaction strength between predators and prey, respectively. 79 Indirect effects in predator and prey assemblages can also be studied by evaluating these graphs. 80 The strength of predator-predator and prey-prey indirect interactions is extremely difficult to 81 measure but, if they prove generally prevalent, they could be a major driver of community 82 dynamics and ecosystem functioning (Woodward et al. 2005). The analysis of the degree 83 distribution of links in the overlap graphs, omitted in most FW studies, might be very useful to 84 identify, based on the competition graph, generalist and specialist predators, and to evaluate 85 energy pathways in the common-enemy graph. 86 In the current work, we present the first, detailed analysis of the FW for the Potter Cove 87 ecosystem (South Shetland Islands, Antarctica). The objectives of this study were to: 1) analyze

the complexity and structure of the ecological network in the context of the most-studied marine

FWs; and 2) examine its degree distribution and overlap graphs in order to gain insight into theecosystem dynamics and functioning.

91 **2. Methods**

92 Potter Cove is a 4 km long and 2.5 km wide Antarctic fjord located at 25 de Mayo/King George 93 Island (62°14'S, 58°40'W, South Shetland Islands) (Fig. 1). A shallow sill (< 30 m) separates its 94 inner and outer areas. The inner cove is characterized by soft sediments and by a shallower depth 95 than the outer cove (< 50 m); in the outer cove the bottom is mainly rocky and with average 96 depths of 100 m. Potter Cove is adjacent to Maxwell Bay, which connects to the Bransfield 97 Strait. Water circulation in Potter Cove is strongly influenced by the general circulation of Maxwell Bay (Roese and Drabble 1998). A cyclonic circulation has been identified, with 98 99 efficient water renewal in the northern sector, where water from Maxwell Bay enters the Cove. 100 Freshwater input varies both seasonally and inter-annually and carries important amounts of 101 suspended sediments. Two main creeks discharge into the Cove, the Matias and the Potter 102 creeks. They exhibit different regimes, the first being snowy and lacustrine, the latter snowy and glacial (Varela 1998). Drainage ranged between 0.03 and 0.11 m³ s⁻¹ in the Matias Creek and 103 from 0.08 to 3.8 m³ s⁻¹ in Potter Creek (Varela 1998). Suspended sediment discharges ranged 104 between 0.04 and 15 kg m⁻³ (average = 0.14 kg m³), which correlate with air temperature. These 105 106 characteristics are consistent with data from other glaciomarine environments in Antarctic 107 coastal waters (Leventer and Dunbar 1985).



108 Fig. 1. Map of Potter Cove and its location on Isla 25 de Mayo/King George Island.

109 2.1. Food web assembly

110 We constructed the FW of Potter Cove ecosystem primarily based on information about species

111 living in that system and their feeding habits from studies within the framework of an

112 international research cooperation between Argentina and Germany initiated in 1994 and

113 ongoing for more than 20 years (Wiencke et al. 1998, 2008).

We collected information on feeding links by a thorough literature search (> 500 papers and reports revised). To assemble the network we only considered trophic interactions confirmed by gut content studies and/or field observation. Furthermore, direct observations of researchers from field sampling campaigns in the Cove (e.g. divers when collecting benthic samples) were also taken into account. Laboratory experimental studies, where feeding selectivity, palatability or behavior was tested, were not included in this study as we consider the trophic links proved from

- 120 experiments are not as robust as the ones gathered from the field data. Investigations using
- 121 biomarkers (i.e. stable isotopes and fatty acids) were not considered since trophic interactions are

122 established by sampling few individuals (n \approx 10-100) and studied prey-predator relationships are 123 usually between trophic species widely aggregated. Further details on the trophic links included 124 in the present study (references and methods used to confirm a link) are presented in the 125 electronic supplementary material (Appendix A). 126 Trophospecies, here defined as aggregated groups of taxa, were only considered when data on 127 specific biological species were not available (lack of data resolution) or when taxa shared the 128 same set of predators and prey within the FW (trophic similarity criteria). We have not 129 considered top vertebrate predators (e.g. penguins, seals, whales), as they only sporadically enter 130 the Cove to feed. In addition, pelagic fish (typically taken by Antarctic penguins and pinnipeds) 131 were not considered due to paucity of ocurrence (Barrera-Oro and Casaux 2008). 132 The diversity of the expertise of the authors contributing to the present study was a key factor in 133 generating the quality of the FW, and inherently improved the network representation of the 134 Potter Cove ecosystem.

135 2.2. Network analysis

An interaction matrix of pairwise interactions was constructed; a value of 1 or 0 was assigned to each element a_{ij} of the matrix depending on whether the *j*-species preyed or not on the *i*-species. The FW is an oriented graph with *L* trophic links between *S* nodes or species. The FW graph was drawn from the interaction matrix using Visone software version 2.9.2 (Brandes and Wagner 2004).

Several network properties that are commonly used to describe complexity and structure in FWs
were calculated (Dunne et al. 2002b, de Santana et al. 2013): (1) number of species, *S*; (2) total
number of interactions or trophic links, *L*; (3) number of interactions per species or linkage

144 density, *L/S*; (4) connectance or trophic links divided by total number of possible interactions,

145 $C=L/S^2$; percentage of (5) top species (species with prey but without predators), (6) intermediate

- 146 species (species with prey and predators), (7) basal species (species with predators/consumers
- 147 but without prey); and (8) percentage of omnivores (species eating prey from more than one
- trophic level).
- 149 Trophic levels (TL) of species were calculated using the short-weighted TL formula of Williams
- and Martinez (2004). Short-weighted trophic level is defined as the average of the shortest *TL*

and prey-averaged *TL*. Shortest *TL* of a consumer in a food web is equal to 1 + the shortest chain

152 length from this consumer to any basal species (Williams and Martinez 2004). Prey averaged TL

153 is equal to 1 + the mean *TL* of all consumer's trophic resources, calculated as

$$TLj = 1 + \sum_{i=1}^{S} lij \frac{TLi}{nj}$$

where TL_j is the trophic level of species *j*; S is the total number of species in the food web; l_{ij} is the connection matrix with *S* rows and *S* columns, in which for column *j* and row *i*, l_{ij} is 1 if species *j* consumes species *i* and 0 if not; and n_j is the number of prey species in the diet of species *j*. Therefore, Short-weighted *TL* yields a minimum estimate of *TL* and assumes a value of 1.0 for basal species (Williams and Martinez 2004). We considered the mean *TL* of the web as the average of all species' *TL*.

160 Two secondary graphs, the competition graph and the common-enemy graph, were constructed. 161 The first one, also known as predator overlap graph, connects predators that share one or more 162 prey, while the latter is drawn by connecting prey species sharing one or more predators (Pimm 163 et al. 1991). Predator overlap graph results were discussed considering dietary data on each 164 predator species involved. To examine a plausible organization in predator and prey species, we 165 separately studied the degree distribution of links P(k) for each overlap graph. Links in predator

166	distribution represent the number of prey, while in prey distribution it depicts number of
167	predators. Graphs were plotted using Visone software (version 2.9.2).
168	We also studied the topology of the FW by measuring three more properties: (9) characteristic
169	path length (ChPath), or the average shortest path length between all pairs of species, (10)
170	clustering coefficient (CC), or the average fraction of pairs of species connected to the same
171	species that are also connected to each other, and (11) degree distribution, or the fraction of
172	trophic species $P(k)$ that have k or more trophic links (both predator and prey links) (Albert and
173	Barabási 2002). Trophic links were treated as undirected when calculating path length and
174	clustering because effects can propagate through the web in either direction, through changes to
175	both predator and prey species (Watts and Strogatz 1998).
176	Results of these properties and the ones aforementioned for Potter Cove FW were compared
177	among other marine webs that were chosen considering different criteria: size (S $>$ 25), temporal
178	era (fourth era, see Link et al. 2005) and quality data (i.e. FWs built upon stable isotopes were
179	excluded).
180	Degree distributions of the total FW and of the mentioned overlap graphs were examined and
181	fitted using nonlinear regression analysis (Xiao et al. 2011). Model selection was performed by
182	computing the Akaike Information Criterion corrected for small sample size (AICc) (Burnham

- 183 and Anderson 2002, Xiao et al. 2010). R package *nls* (Nonlinear Least Squares) was used to
- 184 make power-law and exponential fitting (R Core Team 2016).

185 **3. Results**

- 186 The Potter Cove FW (Fig. 2) includes 91 species, composed of 71 biological species, 17
- 187 trophospecies (i.e., merging two or more taxonomic species by trophic similarity) and 3 non-
- 188 living nodes (i.e. fresh detritus, aged detritus and necromass).



Fig. 2. Graphic representation of Potter Cove FW with the trophic level (TL) on the vertical scaleand node width proportional to the total degree (in- and out-). Node colors are by functional

191 group. Network was plotted with Visone (version 2.9.2). See electronic supplementary material

192 (Appendix B) for exhaustive lists of trophic species, their trophic level, degree (in- and out-

- 193 links), functional and taxonomic group affiliation (e.g. algae, phytoplankton, zooplankton, fish,
- amphipods).
- 195 Algae (24 species) comprise red (13 spp.), brown (7 spp.) and green algae (4 spp.). The next
- trophic levels consist of 13 amphipod species, 3 isopod species, 4 sponge species (one

197 aggregated node: Stylocordyla borealis and Mycale acerata), 5 gastropod species, 2 bivalve

198 species, 7 echinoderm species, and 9 demersal fish species. See electronic supplementary

199 material (Appendix B) for exhaustive lists of taxa, their trophic level, degree (in- and out-links),

200 functional and taxonomic group affiliation (e.g. algae, phytoplankton, zooplankton, fish,

amphipods).

202 The first thing to note about Potter Cove FW is that most of the species (47%) were at

203 intermediate levels, implying that they act as predators and prey depending on the trophic

204 interaction they are involved in. Moreover, as shown in Fig. 2 some species are far more

205 connected (9 species with degree > 15) than others, according to the total number of trophic

206 interactions they have (e.g. fish and echinoderms).

207 The main properties of the network complexity for Potter Cove FW included 307 total

208 interactions and a linkage density of 3.4. As a consequence, a connectance of 0.04 was reported

209 (Table 1).

- 210 Table 1. Properties of network complexity and structure for Potter Cove FW. S = number of
- 211 trophic species, L/S = linkage density, C = connectance (L/S²), T = % top species, I = %

intermediate species, B = % basal species, Omn = percentage of omnivorous, TL = mean trophiclevel, ChPath = characteristic path length, CC = clustering coefficient.

Food web	S	L/S	С	Т	Ι	В	Omn	TL	ChPath	CC
Potter Cove	91	3.4	0.04	19	47	34	45	2.1	1.8	0.08

Although intermediate species outnumbered top and basal species, comprising more than half of the species in the FW, the basal species were also numerous (Table 1). In addition, almost half of the species were omnivorous (45%), similar to the percentage observed in intermediate species. The mean trophic level (TL) for Potter Cove FW was 2.1, which was supported by the relatively high proportion of basal species that tend to lower the average. 219 Network topological properties, characteristic path length (ChPath) and clustering coefficient

(CC) were 1.8 and 0.08, respectively.

- 221 The degree distribution for the Potter Cove FW (Fig. 3) showed that the exponential model best
- fitted the data, according to nonlinear regression and AICc analyses (Table 2). The three species
- 223 with the highest degree were: Notothenia coriiceps (fish, 48 links), Ophionotus victoriae

224 (echinoderm, 33 links) and *Gondogeneia antarctica* (amphipod, 20 links).



Fig. 3. Log-log degree distribution of links P(k) for Potter Cove FW. Two candidate models are shown. Best fit is the exponential model.

- 227 The competition graph derived from Potter Cove FW is highly connected. It includes 60 species
- and 478 indirect interactions (Fig. 4) and shows that several pairs of predators share many prey.
- For instance, all trophic species of sponges form a more connected group than with the rest of the
- 230 prey species. Furthermore, some species of echinoderms, amphipods and demersal fish are
- intensively competing for common food sources (see link width and color, Fig. 4).



Fig. 4. Competition graph for the Potter Cove FW. Node colors (as in Fig. 2): functional groups.Link width and colors: number of shared prey.

To study these potential species interactions, specific competition graphs for the latter two
functional groups were built (Fig. 5). The fish overlap graph includes 9 biological species and 28
competitive interactions. It is worthy to note that two species, *Notothenia coriiceps* and *Harpagifer antarcticus*, presented highly overlapping diets. Moreover, *N. coriiceps shares many of the same prey species, which may or may not involve any competition* (Fig. 5 a). On the other
hand, the amphipod overlap graph suggested low resource overlap among species. However, *Gondogeneia antarctica* and *Prostebbingia gracilis* have many prey in common (Fig. 5 b).



Fig. 5. Competition graphs for (a) demersal fish and (b) amphipod functional groups. Link widthand colors: number of shared prey (see Fig. 4).

The common-enemy graph shows a hyperconnected structure, where the majority of the species are connected. It contained 74 prey species and 1497 indirect interactions (Fig. 6, up-left). Most of the species are connected due to having only one predator in common. In order to elucidate groups of species having stronger indirect interactions, we eliminated links with value 1. This new graph (Fig. 6, large network) showed groups of species connected by strong interactions:

- 248 sponges (except for *Dendrilla antarctica*), benthic diatoms fresh detritus, benthic diatoms –
- 249 epiphytic diatoms, zooplankton phytoplankton, some species of amphipods (i.e. *Gondogeneia*
- 250 antarctica Paradexamine sp. Prostebbingia sp. Eurymera meticulosa), and several red and
- brown algae (Gigartina skottsbergii Desmarestia menziesii Iridaea cordata) (Fig. 6).



- Fig. 6. Common-enemy graph for Potter Cove FW. Original graph in left upper corner. Large
- 253 network shows prey species that share more than one predator. Node colors (as in Fig. 2):
- 254 functional groups. Link width and colors: number of shared predators.
- 255 Degree distribution of links in the competition and common-enemy graphs (Fig. 7) fit best to an
- exponential model (Table 2).



Fig. 7. Log-log degree distribution of links P(k) for (a) the competition and (b) common-enemy graphs. Best fit is the exponential model for both distributions.

Comparison between the Potter Cove FW and other marine webs showed that linkage density
(*L/S*) and connectance (*C*) were lower in the Potter Cove web. The proportions of top and basal
species were relatively high, whereas the percentage of omnivory was the second lowest among
all webs that were compared. While the characteristic path length in Potter Cove FW was similar
to the rest of the FWs, the clustering coefficient was one order of magnitude lower (Table 3).

Table 2. Model fit of exponential and power-law models for degree distributions of total FW (in-

and out-links), competition (only predators) and common-enemy (only prey) overlap graphs.

267 AICc and AIC $_{\Delta}$ are the Akaike corrected for small sample size and delta values for each

268 candidate model. * Indicates best-fit model.

	Model	AICc	AIC_{Δ}
Total FW	Exponential *	94.90	0.000
	Power-law	101.70	6.756
Competition graph	Exponential *	72.56	0.000
	Power-law	76.31	3.751
Common-enemy graph	Exponential *	76.16	0.000
	Power-law	82.00	5.839

269 Table 3. Comparison of network properties between Potter Cove and other marine FWs. S =

number of trophic species, L/S = linkage density, C = connectance (L/S²), T = % top species, I =

271 % intermediate species, B = % basal species, Omn = percentage of omnivorous, TL = mean

trophic level, ChPath = characteristic path length, CC = clustering coefficient. NA: not available

273 data

Food web	S	L/S	С	Т	Ι	В	Omn	TL	ChPath	CC	Source
Marine non-polar webs											
Car. reef (s)	50	11.1	0.22	0	94	6	86	2.9	1.6	0.36	Opitz (1996)
Benguela	29	7.0	0.24	0	93	7	76	3.2	1.6	0.30	Yodzis
											(1998)
NE US Shelf	79	17.8	0.22	4	94	3	62	3.1	1.6	0.31	Link (2002)
Carib. (l)	249	13.3	0.05	NA	NA	NA	NA	NA	1.9	0.16	Rezende et
											al. (2009)
Lough Hyne	350	14.7	0.04	NA	NA	NA	NA	NA	NA	NA	Riede et al.
											2010
Marine polar v	vebs										
Arctic I	140	6.8	0.05	40	56	14	80.7	2.3	NA	NA	Bodini et al.
											(2009)
Arctic II	159	8.6	0.05	NA	NA	NA	52	2.7	2.3	0.25	Kortsch et al.
											2015
Antarctic	586	6.8	0.01	23	21	56	41.1	2.2	3.0	0.14	de Santana et
	\mathbf{Y}										al. (2013)
Weddell Sea	491	33.2	0.07	6.5	80	13.5	67	2.5	NA	NA	Jacob (2005)
Potter Cove	91	3.4	0.04	19	47	34	45	2.1	1.8	0.08	This study

274 **4. Discussion**

4.1. Food web complexity and structure

276 Potter Cove FW properties of complexity and structure showed several singularities that make 277 the web unique in terms of species-richness, link configuration and topological characteristics. 278 Network complexity was mainly assessed by linkage density (L/S) and connectance (C). Both of 279 these properties were found to be relatively low in the Potter Cove web: L/S=3.4 and C=0.04. 280 Nevertheless, direct comparisons of linkage density and connectance values suggest that marine 281 FWs tend to resemble each other, and that they are fundamentally different from other kinds of 282 FWs, based on their high values (Dunne et al. 2004). Opposite to this hypothesis of marine FW similarity, de Santana et al. (2013) found that connectance in the Arctic marine FW was 5 times 283 284 larger than that of the Antarctic one (0.05 versus 0.01). Furthermore, within marine webs, polar 285 networks tend to display low values of linkage density (de Santana et al. 2013). In this sense, 286 complexity exhibited in the Potter Cove FW resembles closely to what is known so for FWs in 287 Polar regions.

288 Could low values of linkage density and connectance in Potter Cove network be a consequence 289 of methodological issues? Dunne et al. (2002b) suggested that both low- and high-connectance 290 FWs are unusual, and that extreme connectances may sometimes be artifacts of assembly 291 procedures. They exemplified this statement by showing that the lowest connectance webs they 292 studied ($C \approx 0.03$, Grassland and Scotch Broom), which is similar to Potter Cove FW 293 connectance value, are "source-webs". These are constructed by linking food chains upward 294 starting from one or a few basal species. The Potter Cove FW is a species-rich ecological 295 network and not a source-web since it was not constructed upward from one or two basal species 296 but it is characterized by > 30% basal species. Thus there is no evidence we know of which

297	suggests that our low values of linkage density and connectance were a consequence of the
298	assembly procedure of the network. In turn this implies that the assembly-connectance
299	relationship in FWs is not as strong as previously thought (Dunne et al. 2002b).
300	Whether ecological networks display low or high L/S and C values is crucial to gain insight in
301	the ecosystem's synthetic properties like robustness. Empirical analyses of FWs support the
302	notion that the robustness of a FW increases with its linkage density and connectance (De
303	Angelis 1975, Dunne et al. 2002a, Montoya and Solé 2003). Low values of L/S and C found in
304	Potter Cove FW, combined with ongoing climate change effects on benthic communities in the
305	area (Pasotti et al. 2015b, Sahade et al. 2015), suggest potential ecosystem fragility which need
306	to be addressed.
307	Furthermore, direct comparison of common FW properties, like percentages of top, intermediate
308	and basal species, indicates that the Potter Cove network has strong structural differences and
309	shows unique features compared to other marine ecosystems. Important dissimilarities were
310	found in top and basal species values as Potter Cove FW shows a higher number of these trophic
311	species. After comparing 19 FW properties, Dunne et al. (2004) concluded that the excessively
312	low percentage of basal taxa in marine FWs compared to other systems is clearly an artifact of
313	poor resolution of primary producers and consumer links to them. One of the methodological
314	strengths of Potter Cove FW is the high taxonomic resolution of the basal nodes. A good
315	taxonomic resolution of the lower trophic levels, such as the macroalgal community, is essential
316	to understand Potter Cove ecosystem functioning, since there seems to be a species-specific
317	selective consumption (Barrera-Oro and Casaux 1990, Iken et al. 1997, Iken et al. 1998).
318	Furthermore, algal species show a marked pattern of depth distribution and tridimensional
319	structure (Quartino et al. 2005, Huang et al. 2007). Macroalgae are one of the main primary

320 producers in Potter Cove, and probably support a large fraction of secondary production of the 321 benthos community (Quartino et al. 2008). Implications in ecosystem functioning and stability 322 are only possible to elucidate in FWs where species involved in energy and matter transfer 323 processes are well represented. 324 Proportions of intermediate species (I) and omnivory (Omn) in Potter Cove FW are relatively 325 low when compared to other marine webs, but close to values for Antarctic FW as reported by de 326 Santana et al. (2013). Levels of I and omnivory are usually correlated in FW studies, as the 327 majority of species acting as predators and prey also feed on more than one trophic level 328 (omnivorous). The importance of omnivory for the structure and dynamics of FWs is a long-329 standing controversy in ecology (Burns 1989, Polis 1991), and whether omnivory stabilizes or 330 destabilizes webs is not clear (Vandermeer 2006, Namba et al. 2008, Johnson et al. 2014). In 331 Antarctica a recent study suggests that omnivory is a beneficial trait as it allows for more 332 responsive and flexible utilization of food sources that may be temporally and spatially 333 constrained and unpredictable (Norkko et al. 2007). The omnivory reported here for Potter Cove 334 FW is the second lowest percentage among marine webs included in the present study, would suggest a low stability for Potter Cove FW. Additionally, this result generates testable 335 336 hypotheses about the probable stabilizing role of omnivory in large communities, since it was 337 proven that the risk of secondary extinctions after primary loss of species depends on the trophic 338 position of the extinct species (Borrvall et al. 2000) and the diversity of that trophic level 339 (insurance hypothesis, Yachi and Loreau 1999). 340 The mean trophic level for this FW (2.1) is also relatively low, which is the result of several 341 singularities of the Potter Cove ecological network. Firstly, as already mentioned, the number of

basal trophic species is high, exceeding 30% of number of species (diversity). What's more, the

343 maximum trophic level was 4.27, lower than most other FWs studied (Dunne et al. 2002b, 2004), 344 which implies that top and basal species are separated by few intermediate taxa. It is worthy to 345 clarify here that Antarctic top predators, e.g. marine mammals, might increase maximum trophic 346 level of the web but were not included as they are rarely reported in the Cove. Therefore, the 347 transfers of energy or nutrients from the base to the top of Potter Cove FW is small, so that the 348 number of times chemical energy is transformed from a consumer's diet into a consumer's 349 biomass along the FW is also small. Another reason why the mean trophic level is low is the fact 350 that most predators at intermediate levels (e.g. amphipods, isopods, bivalves, N. coriiceps) feed 351 predominantly on algae species and/or detritus, being mainly the product of dead and 352 decomposed macroalgae in Potter Cove (Iken et al. 1998, Huang et al. 2006, Quartino et al. 353 2008). The macroalgal detritus decomposes and is eaten by detritivores and suspensivores (e.g. 354 sponges, ascidians, bryozoans, cnidarians), supporting an important amount of the secondary 355 production (Tatián et al. 2004). The obtained low mean trophic level for Potter Cove FW clearly 356 shows what species-specific and/or community studies have suggested. These characteristics of 357 ecological communities have a high impact on ecosystem functioning, such as nutrient and 358 carbon cycling, and trophic cascades (Post 2002). 359 Short characteristic path length for Potter Cove FW (≈ two degrees of separation) is similar to 360 lengths found in other marine FWs. The length between pairs of species within marine webs is 361 low (≈ 1.6 links) compared to other types of FWs, with values ranging from 1.3 to 3.7 (Dunne et

al. 2002b). This suggests that most species in Potter Cove FW are potentially very closeneighbours, and that negative effects could spread rapidly and widely throughout the web

364 (Dunne et al. 2002a).

365 Additionally, the clustering coefficient in this web (0.08) was an order of magnitude lower than 366 those reported for other marine FWs (Link 2002, Dunne et al. 2004). A low coefficient indicates 367 that most species are similarly connected to each other, i.e. there are no densely sub-groups of 368 species interacting with one other. Particularly, the clustering coefficient result of Potter Cove 369 FW might be the consequence of hubs (i.e. species with high degree, > 20 links) connected with 370 most of the species across the web and not with a specific group of species. The most connected 371 species, N. coriiceps (demersal fish) and Ophionotus victoriae (brittle star), have the widest 372 ecological niches in our study, being generalists and omnivores. By feeding across several 373 trophic levels and transversely in the FW, these species have a strong effect on clustering. 374 Specifically, N. coriiceps probably represents a keystone species in the bentho-pelagic coupling 375 process promoting the transfer of matter and energy between habitats (Barrera-Oro and Casaux 376 2008). At the same time, these hub species might be essential for understanding the spread of 377 perturbations (i.e. biodiversity loss) through the entire FW network.

4.2. Degree distribution and overlap graphs: implications for ecosystem

379 functioning

Webs with low connectance ($C \approx 0.03$), such as Potter Cove FW, are more likely to display power law degree distributions (Dunnet et al. 2002a, Montoya and Solé 2002), consistent with the small-world phenomenon. These are webs combining high clustering, like regular lattices and short path length, like random graphs (Watts and Strogatz 1998). Therefore, the Potter Cove FW, with a low estimated connectance (C = 0.04), should display a power law degree distribution. However, it fits best to an exponential distribution according to the low clustering coefficient. The existence of a universal functional form in the degree distribution of FWs is still under

387 debate, though Stouffer et al. (2005) have shown that approximately exponential degree 388 distributions can be derived from two different models: nested-hierarchy and generalized 389 cascade. 390 The influence of the degree distribution on the vulnerability of complex networks against 391 random failures and intentional attacks has become well known since the work of Albert et al. 392 (2000). Considering this relationship between degree distribution and vulnerability, Potter Cove 393 FW would be highly fragile to the removal of the most connected species, but not as much as in 394 power law networks (Albert el al. 2000, Dunne et al. 2002a, Estrada 2007). Furthermore, de 395 Santana et al. (2013) suggested that less connected communities should be more sensitive to the 396 loss of basal species than complex communities because the consumers in simple communities 397 are dependent on only a few species and cannot survive their loss. Nevertheless, we hypothesize 398 that although Potter Cove FW shows low connectance, it will be robust against basal node 399 extinctions due to the high percentage of these trophic species. 400 In addition, degree distribution of links in the competition graph showed that most species have 401 limited diets, feeding exclusively on few prey, whereas few species feed on a large amount of 402 food-sources, usually being generalists. The graph suggests that several predator species have 403 high prey overlap and thus the potential to strongly interact and compete for common prey; this 404 is the case for sponges, demersal fish and amphipods. We focused the analysis on fish and 405 amphipods as they are known to play an important role on the Antarctic marine ecosystem 406 (Barrera-Oro and Casaux 1998, Momo et al. 1998, Barrera-Oro 2002, Huang et al. 2006). Fish 407 data reflects that there is dietary overlap between N. coriiceps and H. antarcticus on the one hand 408 and between Trematomus newnesi and N. rossii on the other hand. Most of the dietary 409 comparisons for demersal Antarctic fish communities have dealt with food overlap between fish

410 species pairs (Barrera-Oro 2003). Dietary overlap index ("S" index of Linton et al. 1981) 411 between N. coriiceps and N. rossii in Potter Cove as estimated by Barrera-Oro (2003) was 55%, 412 meaning that these species could compete for more than half of their food-sources. The same 413 study estimated the index for N. coriiceps – T. newnesi, being 18%, and N. coriiceps – H. 414 antarcticus, being 19%. Barrera-Oro (2003) concludes that there is no evidence of food 415 competition among the shallow cold-water fish communities in Potter Cove. Nevertheless, our 416 results show that N. coriiceps and H. antarcticus have many prey in common, with a high degree 417 of overlapping. However, due to the differences in mobility, habitat use and adult size between 418 these two species (total length: 45 and 13 cm respectively), competition is probably low (Casaux 419 1998, Barrera-Oro 2003). Although the first one is a generalist and the latter a specialist, both 420 species can be grouped in the same feeding category given that they are benthos feeders, eating 421 amphipods (e.g. Gondogeneia antarctica, Paradexamine sp., Prostebbingia sp., Eurymera 422 monticulosa), gastropods (e.g. Margarella antarctica, Nacella concinna, Eatoniella sp., 423 Neobuccinum eatoni), polychaetes (e.g. Nereidae), and krill in summer (Euphausia superba). 424 On the other hand, the competition graph for amphipods exhibited low dietary overlap among 425 species. It is worth mentioning that hyperiids and *Bovallia gigantea* are not connected, which 426 indicates that they do not share food sources with any other amphipods, nor between themselves. 427 Hyperiids and *B. gigantea* are both carnivores, though the latter mainly feeds on other species of 428 amphipods, such as E. monticulosa, Prostebbingia sp. and G. antarctica (Richard 1977). On the 429 contrary, hyperiids principally eat planktonic prey, such as copepods (Pakhomov and 430 Perissinotto 1996). The most important result of the overlap graph is that species are separated 431 according to their feeding strategies: herbivores (P. gracilis, G. antarctica, O. bidentata and 432 Prostebbingia sp. - left of the graph), detritivores (C. femoratus and Paradexamine sp. - middle

433 graph), and scavengers (W. obesa, H. kergueleni, O. plebs and P. integricauda – right of the 434 graph). This demonstrates the importance and utility of the analysis of competition graphs, in 435 order to better understand alternative energy pathways within apparent trophic guilds; analysis 436 that would be improved by adding information on each predator species (e.g. body size and 437 mass, niche specialization). 438 Common-enemy graph derived from Potter Cove FW showed a hyper-connected graph, which 439 implies that most prey species share at least one predator. The fact that the prey overlap graph of 440 this FW exhibited high connectivity and exponential distribution has implications for the 441 functioning of the ecosystem. High-connected prey in Potter Cove FW are: phytoplankton -442 zooplankton, benthic diatoms - epiphytic diatoms, and fresh detritus - benthic diatoms. The 443 latter shows that several sources of food and alternative energy pathways exist in the Potter Cove ecosystem: phytoplankton (Ahn et al. 1993), benthic microalgal production (Dayton et al. 1986, 444 445 Gilbert 1991), and horizontal advection of allochtonous food particles (Dunbar et al. 1989); 446 important sources of organic matter for marine organisms living in coastal Antarctic ecosystems. In conclusion, comparison of FW properties revealed a particular combination of characteristics 447

for the Potter Cove ecological network: middle size ($S \approx 100$) compared to other marine FWs, low linkage density and connectance (with no evidence of being an artifact of resolution or assembly procedure), low %-omnivory, short path length and low clustering coefficient. According to the overlap graphs and their degree distributions, and the consistency with field observations and investigations, we suggest these analyses are useful tools to gain insight into ecosystem functioning. What is more interesting, the common-enemy graph showed the existence of alternative energy pathways consistent with field investigations in the Cove. As also

- 455 suggested for East Antarctica FW (Gillies et al. 2012), carbon flow among the benthic fauna in
- 456 Potter Cove is complex, with multiple sources of carbon being utilized, which can be asserted
- 457 given the good basal resolution of our network.
- 458 From a network perspective, Potter Cove FW properties suggest fragility and potential trophic
- 459 cascade effects although multiple energy pathways might add robustness to the web. Our results
- 460 suggest that species with a high number of links (e.g. Notothenia corriceps, Ophionotus
- 461 *victoriae*, *Gondogeneia antarctica*) could be considered as keystone species for the robustness of
- 462 Potter Cove ecosystem.

463 Acknowledgments

- 464 This research was supported by Consejo Nacional de Investigaciones Científicas y Técnicas
- 465 (CONICET, Argentina), Universidad Nacional de General Sarmiento and Alfred Wegener
- 466 Institute for Polar and Marine Research (AWI, Germany). The work was partially funded by PIO
- 467 14420140100035CO CONICET Argentina and conducted in the frames of the EU research
- 468 network IMCONet funded by the Marie Curie Action IRSES (FP7 IRSES, Action No. 319718).
- 469 We thank Dave K.A. Barnes for constructive suggestions on language aspects, which helped us
- to improve the manuscript.

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684 Appendices

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The food web of Potter Cove (Antarctica): complexity, structure and function

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