

Variation of mammal diversity along a gradient separated by geographic barriers within the Andes of Perú

JAVIER AMARU-CASTELO^{1,2*}, EDGAR LUIS MARQUINA-MONTESINOS^{1,2}, CAROLINA HERRERA-HUAYHUA^{1,3}, AND SERGIO YANQUE-ACHATA^{1,2}

¹ Crees Foundation for Manu. Fundo Mascoitania s/n, C. P. 17800, Manu. Madre de Dios, Perú. Email: jamarucastelo@gmail.com (JA-C); emarquina@crees-manu.org (EM-M), sergioyanqueachata@gmail.com (SY-A), cherrera@unsa.edu.pe (CH-H).

² Facultad de Ciencias biológicas, Universidad Nacional de San Antonio Abad del Cusco. Av. De La Cultura 773, C. P. 08001. Cusco, Perú. Email: jamarucastelo@gmail.com (JA-C); emarquina@crees-manu.org (EM-M), sergioyanqueachata@gmail.com (SY-A).

³ Facultad de Ciencias biológicas, Universidad Nacional de San Agustín. Calle Santa Catalina 117, C. P.04000, Arequipa. Arequipa, Perú. Email: cherrera@unsa.edu.pe (CH-H).

*Corresponding author: <https://orcid.org/0000-0001-7843-3146>

The protection of many mammal species is restrained by anthropogenic pressures. For this reason, using camera traps is critical to learning about the characteristics of their populations and communities, especially when geographic barriers limit their dispersal. This study aimed to measure the variation in mammal diversity in three areas under different protection levels (Piñi Piñi, Manu Learning Centre, and Aguanos), separated by geographic barriers within the Manu Biosphere Reserve. Relative abundance indices, correspondence analysis, non-metric multidimensional scaling, diversity analysis using Hill numbers, similarity analysis, and Bray-Curtis beta diversity partitioning were measured with the recorded data. Overall, 193 individuals of 36 species were recorded, some showing area preference. *Didelphis marsupialis*, *Dicotyles tajacu*, and *Sylvilagus brasiliensis* prefer areas with a lower protection level. According to Hill's diversity indices, the most diverse area is the Manu Learning Centre. The three areas show variations in diversity due to changes in their composition (balanced variation) influenced by geographic barriers, such as Cerro Teparo Punta and the Alto Madre de Dios River.

Muchas especies de mamíferos sufren presiones antrópicas que dificultan su protección, por lo que el uso de cámaras trampa para conocer las características de sus poblaciones y comunidades es muy importante, mucho más cuando se tienen barreras geográficas que podrían limitar su dispersión. El objetivo de este trabajo fue medir la variación de la diversidad de mamíferos en 3 zonas con diferentes grados de protección (Piñi Piñi, Manu Learning Centre y Aguanos), que se encuentran separadas por barreras geográficas dentro de la Reserva de Biosfera del Manu, Perú. Se midieron índices de abundancia relativa, análisis de correspondencia, escalamiento multidimensional no métrico, análisis de la diversidad usando números de Hill, análisis de similitud y partición de la diversidad beta de Bray-Curtis. Se registraron 193 individuos de 36 especies. *Didelphis marsupialis*, *Dicotyles tajacu*, *Sylvilagus brasiliensis* tienen tendencia hacia áreas con menor nivel de protección. El área más diversa según los índices de diversidad de Hill corresponde a Manu Learning Centre. Las tres áreas presentan una variación de la diversidad debida a cambios en su composición (variación balanceada) influenciada por la presencia de barreras geográficas como el cerro Teparo Punta y el río Alto Madre de Dios.

Keywords: Balanced variation; camera trap; disturbance; Madre de Dios; Manu Learning Centre.

© 2023 Asociación Mexicana de Mastozoología, www.mastozoologiamexicana.org

Introduction

The Neotropics is characterized by a high mammal richness, with 1,617 of the 6,495 species known worldwide (Burgin *et al.* 2018). Perú is the second most diverse country in mammals in South America, with 573 species grouped into 223 genera (Pacheco *et al.* 2021). Twenty-one of the mammal species registered for Peru are endemic to the country and listed in some threat category. The Manu Biosphere Reserve, in southeast Perú, is home to 222 species, accounting for 39 % of the species recorded in Perú (Solari *et al.* 2006). The number of species recorded in Perú and this reserve may increase because many areas are still unexplored or with little sampling effort (Pacheco *et al.* 2009, 2021). Mammals are frequently used in conservation as key, flag, and umbrella species for various reasons, such as their central role in trophic webs, charisma, and broad distribution (Thornton *et al.* 2016; Figel *et al.* 2018).

The distribution range of a species stretches from a center of abundance to barriers that limit its dispersion (Grinnell 1914; Aliaga-Samanez *et al.* 2020). These can be intangible, such as inter- and intraspecific relationships and climatic factors, or tangible, such as rivers, mountains, and land-use changes (Wallace 1854; Grinnell 1914; Oswald *et al.* 2016; Aliaga-Samanez *et al.* 2020). The permeability of barriers varies depending on the characteristics of each species and may even change over time (Aliaga-Samanez *et al.* 2020). These barriers lead to variations in the composition of communities and their diversity in landscape units or surrounding landscapes (Grinnell 1914; Ayres and Clutton-Brock 1992; Gascon *et al.* 2000).

Some studies have addressed mammal diversity and its differences between landscape types or vegetation units (Pérez-Irineo and Santos-Moreno 2010; Aquino *et al.* 2012; Cruz-Jácome *et al.* 2015; Hernández-Pérez *et al.* 2015; Li *et al.*

2021). In recent years, the use of trap cameras for the study of mammals has increased significantly since they are considered an affordable, reliable, and non-invasive research tool that allows for recording cryptic and evasive species (Pérez-Irinea and Santos-Moreno 2010; Hernández-Pérez et al. 2015; Mosquera-Guerra et al. 2018). Camera traps facilitate data collection to calculate relative abundance, activity patterns, diversity, and spatial variation (Cruz-Jácome et al. 2015; Hernández-Pérez et al. 2015; Mosquera-Guerra et al. 2018). This information is important to define priority conservation areas with quantitative methods, such as those proposed by Chávez-Gonzalez et al. (2014), or to contribute additional information to improve existing proposals in priority areas (Monroy-Vilchis et al. 2011; Mosquera-Guerra et al. 2018).

Priority conservation areas are threatened by the expansion of the agricultural frontier and road networks, incorporation of pollutants from areas outside their limits, and resource overexploitation due to the intensification of hunting and fishing (Osorio-Plenge et al. 2012; SERNANP 2019; Shepard et al. 2010). In and around priority conservation areas, many mammal species are used as a source of protein (bushmeat) for cultural reasons, subsistence, and growing economic needs (Aquino et al. 2007; Fa et al. 2013). In this context, 25 % of species consumed as bushmeat in South America are under some category of threat according to IUCN, so their vulnerability is intensified by overhunting (Aquino et al. 2007; Fa et al. 2013). One of the priority conservation areas in Peru is the Manu Biosphere Reserve (RBM, for its acronym in Spanish; SERNANP 2019).

The RBM comprises an area of 1,881,200 ha, including a core area (Manu National Park) and a buffer zone (SERNANP 2019). This reserve is considered a conservation hotspot due to its high biological diversity (Myers et al. 2000) as a result of the different climate types and broad altitudinal range within its area (Smith et al. 2008; Serrano-Rojas et al. 2022). The RBM is home to a wide variety of ecosystems, the most representative of which are the pajonal, the high tropical forest, and the low tropical forest (SERNANP 2019). These and other ecosystems are delimited by geographic barriers that restrain species distribution.

This study aimed to investigate the variation in mammal diversity at three sites within the RBM, each subject to different forms of protection: government-managed, privately-managed, and unprotected. The study also sought to document observations related to both tangible physical barriers, such as the Alto Madre de Dios River and Cerro Teparo Punta, and intangible barriers, including anthropogenic activities and their interactions with mammalian communities.

Materials and methods

Study area. The present work was carried out in the district and province of Manu, in the department of Madre de Dios within the RBM (Figure 1). Phototrapping was carried out at three sites with different degrees of conservation, sepa-

rated by geographic barriers between them, and at different distances from urban and rural settlements:

Piñi piñi (-12.770769 °S, -71.489761 °W): Region adjacent to the upper Piñi Piñi River at 618 masl in the Manu National Park. This site is under strict protection by the Peruvian government through the Ministry of the Environment. It is located 15 km from the native community of Santa Rosa de Huacaria (-12.886353 °S, -71.4407001 °W and 4.5 km from the Amalia indigenous people in initial contact settlement (-12.742430 °S, -71.524425 °W). Santa Rosa de Huacaria belongs to the Huachperi-Matsigenka tribe and the Amalia settlement to the Matsigenka tribe. The main activities of these communities are small-scale agriculture for self-consumption and bushmeat hunting (e. g., *Ateles chamek*, *Tayassu peccari*, *Cuniculus paca*, *Dasyprocta punctata*, *Alouatta seniculus*, and *Crax tuberosum*) (Da Silva et al. 2005). It is located 19 km from Pillcopata, a major town where there is constant trade and home to hunters who use rifles and similar guns for bushmeat hunting (e. g., *Tayassu peccari*, *Cuniculus paca*, *Dasyprocta punctata*, and *Crax tuberosum*). It is separated from the Manu Learning Centre biological field station and Aguanos by the Cerro Teparo Punta, a hill stretching from the Coñecongo up to 56 km northwestward, from where trap cameras have been installed. According to Servicio Nacional de Meteorología e Hidrología del Perú (SENAMHI 2020), the local climate is rainy with a dry and temperate winter (B(i)B').

The Manu Learning Centre Biological Field Station (MLC; -12.809389 °S, -71.396056 °W) is situated on the left bank of the Alto Madre de Dios River at an elevation of 524 meters above sea level (asl) in the Manu National Park buffer zone. The station is enveloped by a secondary forest that has been undergoing a self-recovery process for over 30 years. It serves as a hub for biodiversity monitoring and ecotourism activities, and is not supported by government protection but is managed as a private conservation area by the Crees Foundation. It is 19 km from the native Palotoa Teparo community (Matsigenka tribe) on the same left bank of the Alto Madre de Dios River, 2.5 km from the Aguanos village, and 6 km from the Salvacion village, which is the main commercial trade center home to hunters who use firearms. According to SENAMHI (2020), the local climate is rainy with high humidity the year round (B(r)B').

Aguanos (-12.800532 °S, -71.372436 °W): It is located on the right bank of the Alto Madre de Dios River at 470 m asl in the RBM buffer zone, with neither government nor private protection, administered under the local government of the Manu province. It is a hamlet dedicated mainly to growing bananas (*Musa paradisiaca*) and papaya (*Carica papaya*; Santiago-Corisepa et al. 2022). The Villa Salvación village center (-12.836485 °S, -71.361210 °W) is 5 km away. According to SENAMHI (2020), the local climate is rainy with high humidity the year round (B(r)B').

Sampling with trap cameras. At each study site, eight stations were established, distributed within a system of grids

of two quadrants, separated from each other by a minimum distance of 1 km. In each station, we placed a camera trap at a height between 30 and 40 cm from the ground and set to capture 15-second videos with 30-second intervals between captures. At each station, cameras were in operation for three months between June and September, corresponding to the dry season in Peru (with 90 hours of effort per station). Those records separated by more than 1 hour were considered independent records for the analyses (Oliveira et al. 2020).

Identification and taxonomic criteria. Species identification was performed by comparison with previous records in the Crees Foundation database and using the descriptions by Emmons and Francois (1990). Species were listed, and scientific names were updated considering the proposal of Pacheco et al. (2021).

Data analysis. Potential differences between the three protection levels were investigated using a similarity analysis (ANOSIM) and a non-metric multidimensional scaling (NMDS). ANOSIM is a non-parametric test that uses permutations to calculate differences between groups; in the study, the groups are the eight sampling stations for each forest type (Legendre and Legendre 1983). The NMDS is an ordination method to detect differences between groups using a distance measure, in this case, the Bray-Curtis distance (Legendre and Legendre 1983). The radius of each circle surrounding each point is inversely proportional to the distance to the geometric centroid of the eight sampling points of each forest type.

To determine which of the three forest types is more diverse, we performed Hill's alpha diversity and evenness indices, and rarefaction analyses. Hill's diversity and evenness indices show a better diversity approximation than conventional diversity indices (Hill 1973). Hill's evenness index was calculated by dividing Hill's number of order 0 ($N(0)$) by the number of order 1 ($N(1)$).

To define whether the differences in the three protection level types are due to changes in composition, we constructed rank-abundance curves and performed a Bray-Curtis (Bray) beta diversity partitioning analysis. Beta diversity partitioning allows splitting the Bray-Curtis distance into a balanced variation resulting from changes in community composition and variation in gradients associated with the reduction in community richness and abundance (Jost 2007; Baselga 2013, 2017).

Preferences of some mammal species for a given degree of protection were explored through a correspondence analysis (CA). CA is an ordination method that reveals differences between objects and plots the descriptors associated with them using the Chi-square distance (Legendre and Legendre 1983). CA was carried out considering the type-I scaling, excluding species that were only recorded once. All the analyses and graphs were performed using the Python 3.10.9 programming language in the Spyder 5.4.2 IDE, using the packages NumPy 1.24, eCopy 0.1.2.2, Pandas 1.5.3, and Matplotlib 3.7.0.

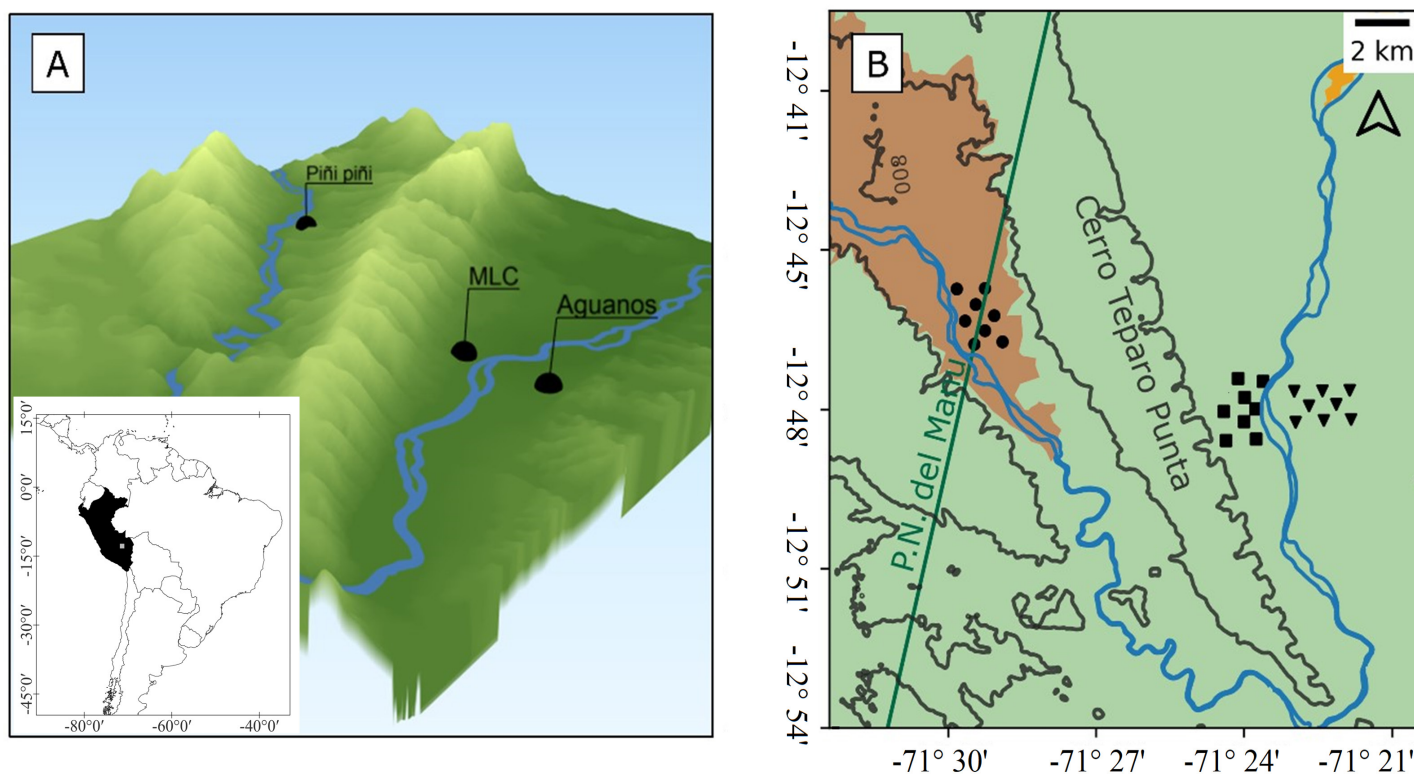


Figure 1. Study area. a) Three-dimensional elevation map showing the mountain formations, Madre de Dios River, Piñi Piñi River, and collection sites. b) Map of the layout of camera traps used in the study area showing mountain ranges, climate types according to SENAMHI (2020). Light green. weather A(r)A'; Light orange. weather B(r)B'; Line green. edge of Manu National Park; Black triangle. Aguanos; Black circle. Piñi piñi; Black square. MLC biological station.

Results

A total of 193 individuals of 36 species were recorded with a sampling effort of 2,160 h. The most abundant species was *Cuniculus paca*, with 21 individuals, followed by *Dasybus novemcinctus* and *Mazama americana*, with 19 individuals each. The families with the highest species richness were Didelphidae and Felidae, with five recorded species each. The most abundant families were Felidae, with 36 individuals; Cuniculidae, with 21; and Dasypodidae and Cervidae, with 19 individuals each. *Cuniculus paca*, *M. americana*, and *D. novemcinctus* are abundantly distributed in the three protection levels, being considered dominant species across the entire study area. Within the family Felidae, the dominant species in the three protection levels was *Leopardus pardalis* (Table 1).

At Aguanos, the most abundant species was *Dicotyles tajacu*, with eight individuals, followed by *C. paca* and *M. americana*, with seven individuals each. Six singleton species were observed, and *Galictis vittata*, *Potos flavus*, *Ateolocynus microtis*, *Saimiri boliviensis*, and *Caluromys lanatus* were recorded exclusively in this area (Figure 2 and Table 1). At MLC, the most abundant species were *C. paca* and *D. novemcinctus*, with eight individuals each, followed by *M. americana* and *Tapirus terrestris*, with seven individuals each. Nine singletons with a single record were observed, and *Microsciurus flaviventer*, *Sciurus ignitus*, *Callicebus urubambensis*, *Metachirus nudicaudatus*, *Procyon cancrivorus*, *Philander opossum*, *Sciurus spadiceus*, and *Chironectes minimus* were recorded exclusively in this area (Figure 2 and Table 1). Last, at Piñi Piñi, the most abundant species was *C. paca*, with six individuals; eight singleton species were observed, and *Sapajus apella*, *Dinomys branickii*, and *Lagothrix flavicauda* were observed only in this area (Figure 2 and Table 1).

According to the correspondence analysis (Figure 3), some mammal species prefer a certain area. This preference is more noticeable in *Didelphis marsupialis*, *D. tajacu*, and *Sylvilagus brasiliensis*, which tend to prefer more open, disturbed, and unprotected areas (such as Aguanos); and *Nasua nasua*, which showed a trend toward protected and conserved areas within the Manu National Park (Piñi Piñi). These preference variations translate into statistically significant differences in the composition of mammal species between the three protection levels in the similarity analysis ($R = 0.159$; $P < 0.05$). Additionally, the above is supported by the non-metric multidimensional analysis (NMDS; Figure 3), showing that, although there is a high similarity between points closer to the centroid of each conservation level, there are also points that allow differentiating them.

The analysis of Hill's alpha diversity numbers (Table 2) revealed that the most diverse protection level is MLC, followed by Aguanos and Piñi Piñi. As for evenness ($E(1.0)$), the most even forest was Aguanos, followed by MLC and Piñi Piñi. The shift of order in the evenness index between MLC and Aguanos occurs because the former recorded more singleton species (9) than Aguanos. Since different numbers of total individuals were recorded in each zone (Aguanos, 73; MLC, 75; and Piñi Piñi, 45), the effort was equalized using the rarefaction index, with MLC attaining the highest index, followed by Aguanos and Piñi Piñi, with very similar index values (Table 2).

In the Bray-Curtis beta diversity partitioning (Table 3), the overall variation is mostly due to balanced variation (63.38%), indicating that the species composition changes drastically from one protection level to another. When analyzed separately, this prevalence of beta diversity due to balanced variation (B-bal) is maintained when comparing

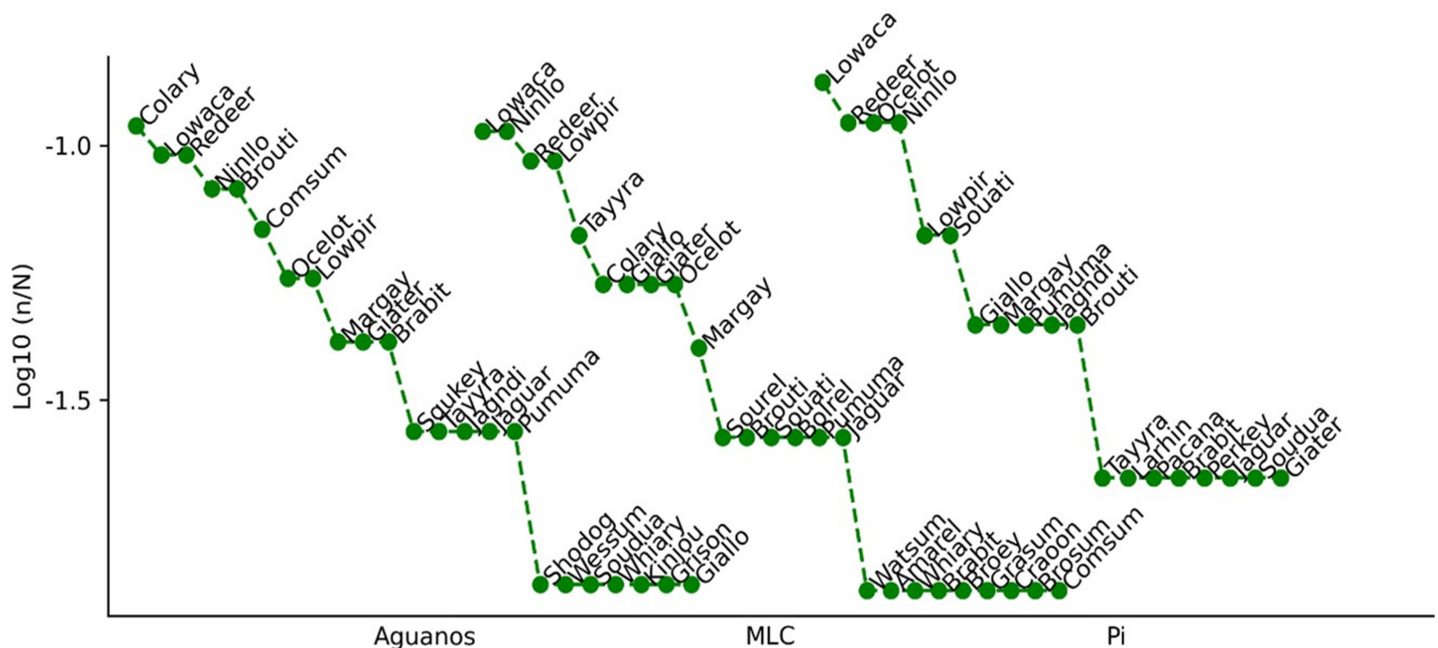


Figure 2. Rank-abundance curves of species recorded by forest type. Graph codes are specified in Table 1.

Table 1. Abundance and richness of mammals collected in the study. Cod, name code; Ag, Aguanos; Pi, Piñi Piñi; M, Biological Station Manu Learning Centre; To, total, CA, conservation level according to SERFOR (2018). IUCN, threat category according to IUCN (2022); LC, Least Concern; EN, Endangered; VU, Vulnerable; NT, Near Threatened

Familia	Especie	Nombre	Cod	Ag.	Pi	M	To	CA	IUCN
Sciuridae	<i>Microsciurus flaviventer</i>	Amazon Dwarf Squirrel	Amarel	0	0	1	1	DI	LC
Sciuridae	<i>Hadroskiurus ignitus</i>	Bolivian Squirrel	Bolrel	0	0	2	2	-	LC
Leporidae	<i>Sylvilagus brasiliensis</i>	Brazilian Rabbit	Brabit	3	1	1	5	-	EN
Pitheciidae	<i>Callicebus urubambensis</i>	Brown Titi Monkey	Broey	0	0	1	1	-	-
Didelphidae	<i>Metachirus myosuroides</i>	Brown Four-Eyed Opossum	Brosom	0	0	1	1	-	-
Dasyproctidae	<i>Dasyprocta variegata</i>	Brown Agouti	Brouti	6	2	2	10	-	LC
Tayassuidae	<i>Dicotyles tajacu</i>	Collared Peccary	Colary	8	0	4	12	-	-
Didelphidae	<i>Didelphis marsupialis</i>	Common Opossum	Comsum	5	0	1	6	-	LC
Procyonidae	<i>Procyon cancrivorus</i>	Crab-Eating Raccoon	Craoon	0	0	1	1	-	LC
Chlamyphoridae	<i>Priodontes maximus</i>	Giant Armadillo	Giallo	1	2	4	7	V	VU
Myrmecophagidae	<i>Myrmecophaga tridactyla</i>	Giant Anteater	Giater	3	1	4	8	V	VU
Didelphidae	<i>Philander sp.</i>	Gray Four-Eyed Opossum	Grasum	0	0	1	1	-	-
Mustelidae	<i>Galictis vittata</i>	Grison	Grison	1	0	0	1	-	LC
Felidae	<i>Puma yagouaroundi</i>	Jaguarundi	Jagndi	2	2	0	4	-	LC
Felidae	<i>Panthera onca</i>	Jaguar	Jaguar	2	1	2	5	CA	NT
Procyonidae	<i>Potos flavus</i>	Kinkajou	Kinjou	1	0	0	1	-	LC
Cebidae	<i>Cebus apella</i>	Large-Headed Capuchin	Larhin	0	1	0	1	-	LC
Cuniculidae	<i>Cuniculus paca</i>	Lowland Paca	Lowaca	7	6	8	21	-	LC
Tapiridae	<i>Tapirus terrestris</i>	Lowland Tapir	Lowpir	4	3	7	14	CA	VU
Felidae	<i>Leopardus wiedii</i>	Margay	Margay	3	2	3	8	DI	NT
Dasyproctidae	<i>Dasyprocta novemcinctus</i>	Nine-Banded Armadillo	Ninllo	6	5	8	19	-	LC
Felidae	<i>Leopardus pardalis</i>	Ocelot	Ocelot	4	5	4	13	-	LC
Dinomysidae	<i>Dinomys branickii</i>	Pacarana	Pacana	0	1	0	1	V	LC
Atelidae	<i>Lagothrix lagothricha</i>	Peruvian Woolly Monkey	Perkey	0	1	0	1	EP	VU
Felidae	<i>Puma concolor</i>	Puma	Pumuma	2	2	2	6	CA	LC
Cervidae	<i>Mazama americana</i>	Red-Brocket Deer	Redeer	7	5	7	19	DI	DD
Canidae	<i>Atelocynus microtis</i>	Short-Eared Dog	Shodog	1	0	0	1	V	NT
Procyonidae	<i>Nasua nasua</i>	South American Coati	Souati	0	3	2	5	-	LC
Myrmecophagidae	<i>Tamandua tetradactyla</i>	Southern Tamandua	Soudua	1	1	0	2	-	LC
Sciuridae	<i>Hadroskiurus spadiceus</i>	Southern Amazon Red Squirrel	Sourel	0	0	2	2	-	LC
Cebidae	<i>Saimiri boliviensis</i>	Squirrel monkey	Squkey	2	0	0	2	-	LC
Mustelidae	<i>Eira barbara</i>	Tayra	Tayyra	2	1	5	8	-	LC
Didelphidae	<i>Chironectes minimus</i>	Water Opossum	Watsum	0	0	1	1	-	LC
Didelphidae	<i>Caluromys lanatus</i>	Western Wolly Opossum	Wessum	1	0	0	1	-	LC
Tayassuidae	<i>Tayassu pecari</i>	White-Lipped Peccary	Whiary	1	0	1	2	CA	VU
Total				73	45	75	193		

MLC with Aguanos (76.14 %) and Aguanos with Piñi Piñi (57.41 %), but changes to a variation by gradients when comparing MLC with Piñi Piñi (59.35 %). Although the variation is mostly balanced, composition changes are also due to gradient variation (B-gra).

Discussion

The protection level of an area determines the presence of human settlements, the activities allowed, and the intensity of resource extraction (Kuamara et al. 2004; Blom et al. 2005; Trisurat et al. 2005; Rabanal et al. 2010). All these factors influence mammalian communities in different regions and at different scales (Kuamara et al. 2004; Blom et al. 2005; Trisurat et al. 2005; Rabanal et al. 2010). The present study is one of the few works relating the compo-

sition of mammal communities to the protection level of areas within a Biosphere Reserve.

In general, the mammal community varies across the three protection levels as some species exhibit preferences for a given level; as a result, beta diversity is primarily due to balanced variation. This variation may be due to various factors, such as geographic barriers that delimit the distribution of species (Gascon et al. 2000; Maciel-Mata et al. 2015; Oswald et al. 2016), climatic conditions that determine seasonality and its presence (e. g., Cândido-Rocha et al. 2006), and anthropic activities such as hunting (Blom et al. 2005; Aquino et al. 2007; Fa et al. 2013).

Tangible geographic barriers, such as Cerro Teparo Punta and Alto Madre de Dios River, restrain the dispersal capacity of some populations, isolating them. As a result, these may

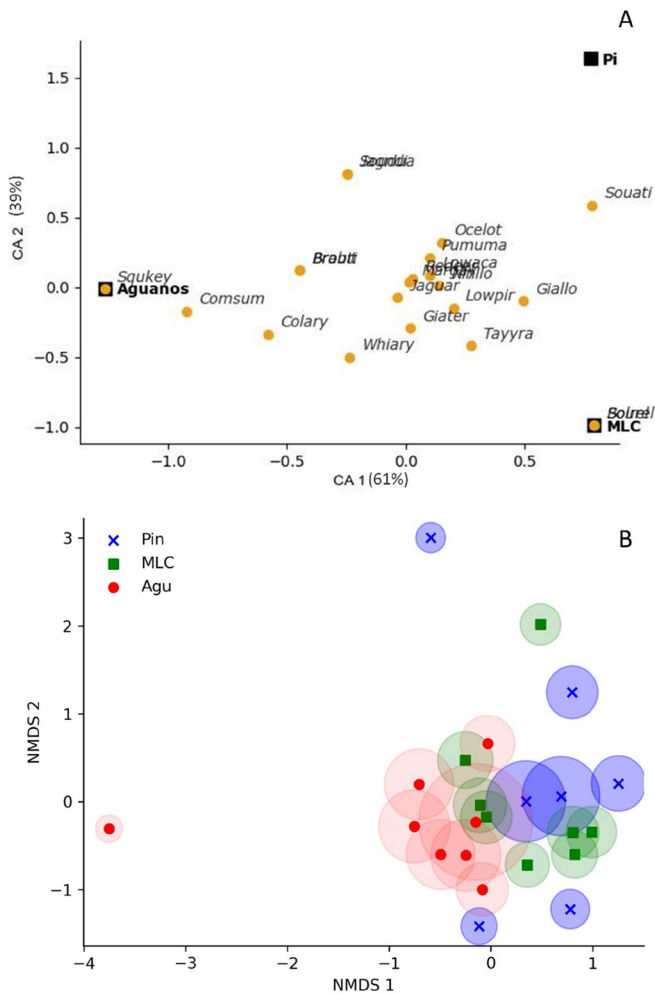


Figure 3. a) Correspondence analysis showing the three collection zones and associated species. b) Non-metric multidimensional scaling of the three forest types, showing camera traps.

produce potential new lineages that respond differently and independently to the local environment (Oswald et al. 2016). Mountain ranges are central to species composition and endemism patterns in Neotropical forests (Oswald et al. 2016). In this way, Cerro Teparo Punta is a barrier for some mammal populations that cannot move across the peak of this relief form. As regards the Alto Madre de Dios River, more studies should be carried out to determine whether it is a barrier restraining the distribution of mammals because not all rivers limit the movement of species, and their permeability depends on the dispersal capacity of each species (Grinnell 1914; Gascon et al. 2000).

Another factor that may explain the variation in the three sampling areas is intangible barriers such as anthropogenic activities, such as hunting, whose preference for some species and strategies differ between regions (Kumar et al. 2004; Aquino et al. 2007; Endo et al. 2010; Fa et al. 2013). This variation in hunting preference decreases the abundance of the most hunted species and favors the abundance of the least hunted ones in areas surrounding populated centers (Endo et al. 2010). In the present work, *Didelphis marsupialis*, *Sylvilagus brasiliensis*, and *Dicotyles tajacu* tend to be present in disturbed areas, maybe cause

Table 2. Analysis of Hill's alpha diversity, rarefaction, and evenness. N(0), zero Hill's number; N(1), First Hill's number; N(2), Second Hill's number; E(1,0), Hill's evenness.

	N (0)	N (1)	N (2)	E (1,0)	rarefy
Aguanos	23.000	18.216	15.446	0.792	19.388
MLC	25.000	19.159	15.756	0.766	20.298
Piñi Piñi	19.000	15.345	12.898	0.808	19.000

they are not usually hunted by local or indigenous populations in the surrounding areas (Endo et al. 2010; Fa et al. 2013; Farfan-Flores et al. 2023). The species most hunted and consumed by Matsigenka indigenous communities in the study area are *Ateles chamek*, *Lagothrix lagotricha*, and *Tayassu pecari*, which may explain the low frequency of these species in the present study (Endo et al. 2010; Farfan-Flores et al. 2023).

All species recorded in the study were previously reported for Peru by Solari et al. (2006) and Pacheco et al. (2021), so they are common for the Manu Biosphere Reserve. A large part of the recorded species are considered under some threat category. According to Servicio Nacional Forestal y de Fauna Silvestre (SERFOR 2018), 12 (34 %) species are listed under a threat category: four as Vulnerable, one as Endangered, four as Nearly Threatened, and three with insufficient data. According to International Union for Conservation of Nature and Natural Resources (IUCN 2022), 31 species (88 %) are considered under a threat category: 21 species as Least Concern (LC), five as Vulnerable (VU), one as Endangered (EN), three as Nearly Threatened (NT), and one as Data Deficient (DD; Table 1).

Cuniculus paca, *Mazama americana*, and *Dasyypus novemcinctus* were the most recorded species in the present work. The three species are widely distributed in Peru and have been recorded in the Pacific rainforest, equatorial dry tropical forest, yungas, and low tropical forest (Pacheco et al. 2021). These species can be considered very abundant in the Peruvian Amazon (Aquino et al. 2012). Aquino et al. (2007) mention them among the species facing heavy hunting pressure, so monitoring strategies are needed to ensure their conservation in the RBM. In addition to these three species, many mammals under a threat category are consumed as a protein source by local and native populations, affecting their diversity and total biomass (Endo et al. 2010; Fa et al. 2013). Therefore, the effect of hunting and the hunting methods should be considered within and around the current and proposed priority conservation areas.

Table 3. Bray-Curtis beta diversity partitioning analysis. B-bal, balanced partition of Bray-Curtis index; B-gra, gradient partition of Bray-Curtis index; B-total, Bray-Curtis index.

L1	L2	B-bal	%B-bal	B-gra	%B-gra	B-total
MLC	Aguanos	0.3263	76.1404	0.1023	23.8596	0.4286
MLC	Piñi piñi	0.2000	40.6452	0.2921	59.3548	0.4921
Aguanos	Piñi piñi	0.2333	57.4073	0.1731	42.5926	0.4065
MLC-Agu-Piñi		0.3314	63.3859	0.1914	36.6141	0.5228

The present work recorded five feline species, accounting for 62.5 % of the feline species reported for Peru (Pacheco et al. 2021). The presence of carnivores such as felines is essential for defining conservation criteria in each area because they play a central role in the ecosystem by limiting the number of herbivores; indeed, they are generally used in conservation strategies (Miller et al. 2001; Figel et al. 2018; Thornton et al. 2016). The most protected areas generally have a greater relative abundance of felines than the least conserved (Pardo Vargas et al. 2016). This is consistent with our work, where 26.6 % of felines were recorded at Piñi Piñi and only 14.6 % at MLC. *Puma concolor* and *Panthera onca* inhabit the entire study area, mainly the RBM, so they should be considered in local conservation strategies as umbrella species, replicating previous models (Solari et al. 2006; Figel et al. 2018; Thornton et al. 2016).

From the records, a detailed review of the *Philander* species recorded in the Manu Learning Centre biological station is needed because this genus entails a controversial taxonomy, with eight described species, of which five are reported for Peru (Voss et al. 2018; Pacheco et al. 2021). Another recorded species for which the distribution should be studied to determine its threat category is *Callicebus urubambensis* (Figure 4), a species endemic to Peru treated

as a subspecies of *Callicebus brunneus* before its description as a separate species (Vermeer and Tello-Alvarado 2015).

Acknowledgments

The authors wish to thank Q. Meyer and J. C. Cardenas, founder and CEO, respectively, of Crees Manu, as well as its team, for facilitating our study of mammal biodiversity in secondary forests of the Manu Learning Centre Biological Field Station.

Likewise, we thank the Frankfurt Zoological Society, particularly O. Mujica, for providing some of the necessary materials to develop the project, and the anonymous reviewers of the article, whose comments improved the manuscript. Thanks also to SERNANP for its support in conducting the work, the research permit, and the technical support through the park ranger, E. Oruro. The work was carried out with the co-financing of the New England Biolabs Foundation and T and J Meyer Family of the project "Reconnecting forest for nature and people: Assessing the conservation and sustainable potential of the Manu-Amarakaeri Conservation Corridor". This manuscript was translated into English by M. E. Sánchez-Salazar through a Therya's full translation grant by the Asociación Mexicana de Mastozoología A. C.



Figure 4. Photography of *Callicebus urubambensis*.

Literature cited

- ALIAGA-SAMANEZ, A., *ET AL.* 2020. Modelling species distributions limited by geographical barriers: A case study with African and American primates. *Global Ecology and Biogeography* 29:444–453.
- AQUINO, R., C., *ET AL.* 2007. Evaluación del impacto de la caza en mamíferos de la cuenca del río Alto Itaya, Amazonía peruana. *Revista Peruana de Biología* 14:181–186.
- AQUINO, R., C. TUESTA, AND E. RENGIFO. 2012. Diversidad de mamíferos y sus preferencias por los tipos de hábitats en la cuenca del río Alto Itaya, Amazonia peruana. *Revista Peruana de Biología* 19:35–42.
- AYRES, J. M., AND T. H. CLUTTON-BROCK. 1992. River boundaries and species range size in Amazonian primates. *The American Naturalist* 140:531–537.
- BASELGA, A. 2013. Separating the two components of abundance-based dissimilarity: Balanced changes in abundance vs. abundance gradients. *Methods in Ecology and Evolution* 4:552–557.
- BASELGA, A. 2017. Partitioning abundance-based multiple-site dissimilarity into components: Balanced variation in abundance and abundance gradients. *Methods in Ecology and Evolution* 8:799–808.
- BLOM, A., R., *ET AL.* 2005. Factors influencing the distribution of large mammals within a protected central African forest. *Oryx* 39:381–388.
- BURGIN, C., *ET AL.* 2018. How many species of mammals are there? *Journal of Mammalogy* 99:1–14.
- CANDIDO-ROCHA, E., *ET AL.* 2006. Evaluación estacional de la riqueza y abundancia de especies de mamíferos en la Reserva Biológica Municipal “Mário Viana”, Mato Grosso, Brasil. *Revista de Biología Tropical* 54:879–888.
- CHÁVEZ-GONZÁLEZ, H., M. J. GONZÁLEZ-GUILLEN, AND P. HERNÁNDEZ. 2014. Metodologías para identificar áreas prioritarias para conservación de ecosistemas naturales. *Revista Mexicana de Ciencias Forestales* 6:8–23.
- CRUZ-JÁCOME, O., *ET AL.* 2015. Richness and relative abundance of medium and large mammals in a community of the Biosphere Reserve Tehuacán-Cuicatlán, Oaxaca, Mexico. *Therya* 6:435–448.
- DA SILVA, M., G. H. SHEPARD, AND D. W. YU. 2005. Conservation implications of primate hunting practices among the Matsigenka of Manu National Park. *Neotropical Primates* 13:31–32.
- EMMONS L. H., AND F. FRANCOIS. 1990. *Neotropical Rainforest Mammals: A Field Guide*. University of Chicago Press, Chicago, U.S.A.
- ENDO, W., *ET AL.* 2010. Game vertebrate densities in hunted and nonhunted forest sites in Manu National Park, Peru. *Biotropica* 42:251–261.
- FA, J. E., *ET AL.* 2013. Reflexiones sobre el impacto y manejo de la caza de mamíferos silvestres en los bosques tropicales. *Ecosistemas* 22:76–83.
- FARFAN-FLORES, J., *ET AL.* 2023. Cacería para autoconsumo durante la pandemia de COVID-19. *Revista Peruana de Biología* 30:e24901.
- FIGEL, J. J., *ET AL.* 2018. Threatened amphibians sheltered under the big cat’s umbrella: conservation of jaguars *Panthera onca* (Carnivora: Felidae) and endemic herpetofauna in Central America. *Revista de Biología Tropical* 66:1741–1753.
- GASCON, C., *ET AL.* 2000. Riverine barriers and the geographic distribution of Amazonian species. *PNAS* 97:13672–13677.
- GRINNELL, J. 1914. Barriers to distribution as regards birds and mammals. *The American Naturalist* 48:248–254.
- HERNANDEZ-PEREZ, E. L., *ET AL.* 2015. Fototrampeo de mamíferos terrestres medianos y grandes asociados a Petenes del Noroeste de la Península de Yucatán, México. *Therya* 6:559–574.
- HILL, M. O. 1973. Diversity and evenness: A unifying notation and its consequences. *Ecology* 54:427–432.
- INTERNATIONAL UNION FOR CONSERVATION OF NATURE (IUCN). 2022. IUCN Red List of Threatened Species. Version 2022-2. IUCN.
- JOST, L. 2007. Partitioning diversity into independent alpha and beta components. *Ecology* 88:2427–2439.
- KUMARA, H. N., AND M. SINGH. 2004. The influence of differing hunting practices on the relative abundance of mammals in two rainforest areas of the Western Ghats, India. *Oryx* 38:321–327.
- LEGENDRE, P., AND L. LEGENDRE. 1983. *Numerical Ecology*. Elsevier, Amsterdam, Países bajos.
- LI, W., *ET AL.* 2021. Habitat characteristics or protected area size: What is more important for the composition and diversity of mammals in nonprotected areas? *Ecology and Evolution* 11:7250–7263.
- MACIEL-MATA, C. A., *ET AL.* 2015. Geographical distribution of the species: A concept review. *Acta Universitaria* 25:3–19.
- MILLER, B., *ET AL.* 2001. The importance of large carnivores to healthy ecosystems. *Endangered Species UPDATE* 18:202–205.
- MONROY-VILCHIS, O., *ET AL.* 2011. Fototrampeo de mamíferos en la Sierra Nanchititla, México: Abundancia relativa y patrón de actividad. *Revista de Biología Tropical* 59:373–383.
- MOSQUERA-GUERRA, F., *ET AL.* 2018. Diversidad, abundancia relativa y patrones de actividad de los mamíferos medianos y grandes, asociados a los bosques riparios del río Bitá, Vichada, Colombia. *Biota Colombiana* 19:202–218.
- MYERS, N., *ET AL.* 2000. Biodiversity hotspots for conservation priorities. *Nature* 403:853–858.
- OHL-SCHACHERER, J., *ET AL.* 2007. The sustainability of subsistence hunting by Matsigenka native communities in Manu National Park, Peru. *Conservation Biology* 21:1174–1185.
- OLIVEIRA, T. G., *ET AL.* 2020. A refined population and conservation assessment of the elusive and endangered northern tiger cat (*Leopardus tigrinus*) in its key worldwide conservation area in Brazil. *Global Ecology and Conservation* 22:1–14.
- OSORES-PLENGE, F., *ET AL.* 2012. Minería informal e ilegal y contaminación con mercurio en Madre de Dios: Un problema de salud pública. *Acta Medica peruana* 29:38–42.
- OSWALD, J. A., *ET AL.* 2016. Historical climatic variability and geographical barriers as drivers of community composition in a biodiversity hotspot. *Journal of Biogeography* 43:123–133.
- PACHECO, V., *ET AL.* 2009. Diversidad y endemismo de los mamíferos del Perú. *Revista Peruana de Biología* 16:5–32.
- PACHECO, V., *ET AL.* 2021. Updated list of the diversity of mammals from Peru and a proposal for its updating. *Revista Peruana de Biología* 28:1–30.
- PARDO VARGAS, L. E., *ET AL.* 2016. Assessing species traits and landscape relationships of the mammalian carnivore community in a neotropical biological corridor. *Biodiversity and Conservation* 25:739–752.

- PÉREZ-IRINEO, G., AND A. SANTOS-MORENO. 2010. Diversidad de una comunidad de mamíferos carnívoros en una selva mediana del noroeste de Oaxaca, México. *Acta Zoológica Mexicana* 26:721–736.
- RABANAL, L. I., ET AL. 2010. Oil prospecting and its impact on large rainforest mammals in Loango National Park, Gabon. *Biological Conservation* 143:1017–1024.
- SANTIAGO-CORISEPA, K. V., ET AL. 2022. Input to the knowledge of arthropods associated to banana (*Musa paradisiaca*), Madre de Dios, Peru. *Agroindustrial Science* 12:169–174.
- SCRIVEN, S. A., ET AL. 2017. Barriers to dispersal of rain forest butterflies in tropical agricultural landscapes. *Biotropica* 49:206–216.
- SERVICIO NACIONAL DE METEOROLOGÍA E HIDROLOGÍA DEL PERÚ (SENAMHI). 2020. Mapa de clasificación climática nacional y resumen ejecutivo. Senahmi. Lima, Perú.
- SERVICIO NACIONAL FORESTAL Y DE FAUNA SILVESTRE (SERFOR). 2018. Fauna silvestre amenazada del Perú. Libro rojo. Serfor, Lima, Perú.
- SERVICIO NACIONAL DE ÁREAS NATURALES PROTEGIDAS (SERNANP). 2019. Plan Maestro del Parque Nacional del Manu, 2019 – 2023. Sernanp. Lima, Perú
- SERRANO-ROJAS, S. J., ET AL. 2022. Indigenous lands are better for amphibian biodiversity conservation than immigrant-managed agricultural lands: A case study from Manu Biosphere Reserve, Peru. *Tropical Conservation Science* 15:1–18.
- SHEPARD, ET AL. 2010. Trouble in paradise: Indigenous populations, anthropological policies, and biodiversity conservation in Manu National Park, Peru. *Journal of Sustainable Forestry* 29:252–301.
- SMITH, V. S., ET AL. 2008. Rodent louse diversity, phylogeny, and co-speciation in the Manu Biosphere Reserve, Peru. *Biological Journal of the Linnean Society* 95:598–610.
- SOLARI, ET AL. 2006. Mammals of the Manu Biosphere Reserve. Pp. 13–23, in *Mammals and birds of the Manu Biosphere Reserve, Peru* (Patterson, B. D., D. F. Stotz, and S. Solari, eds.). *Fieldiana Zoology*. Chicago, U.S.A.
- THORNTON, D., ET AL. 2016. Assessing the umbrella value of a range-wide conservation network for jaguars (*Panthera onca*). *Ecological Applications* 26:1112–1124.
- TRISURAT, Y., ET AL. 2014. Assessing potential effects of land use and climate change on mammal distributions in northern Thailand. *Wildlife Research* 41:522–536.
- VERMEER, J., AND J. C TELLO-ALVARADO. 2015. The distribution and taxonomy of titi monkeys (*Callicebus*) in central and southern Peru, with the description of a new species. *Primate Conservation*. 2015:9–29
- VOSS, R. S., ET AL. 2018. A revision of *Philander* (Marsupialia: Didelphidae), part 1: *P. quica*, *P. canus*, and a new species from Amazonia. *American Museum Novitates* 3891:1–72.
- WALLACE, A. R. 1854. On the monkeys of the Amazon. *Annals and Magazine of Natural History* 14:451–454.
- WILLIG, M. R., AND S. J. PRESLEY. 2016. Biodiversity and metacommunity structure of animals along altitudinal gradients in tropical montane forests. *Journal of Tropical Ecology* 32:421–436.

Associated editor: Lázaro Guevara

Submitted: April 28, 2023; Reviewed: June 23, 2023

Accepted: August 18, 2023; Published on line: September 15, 2023

