



REPORT

# Surface-water/groundwater exchange in a sand dune lake in the Dry Pampean Plain, Argentina: stable isotopic evidence

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## Abstract

Understanding the hydrological functioning of the scarce freshwater bodies of semiarid regions is crucial, especially in those areas affected by anthropic activities involving land-use changes. In the dry western edge of the Argentina Pampean plains, a system of more than 100 shallow lakes of remarkable stability occurs. These lakes exhibit low salinity compared to those located in the more humid belt. This system has constituted the main water resource for humans from prehispanic times to the present. Stable isotopes were used to establish the seasonal surface-water/groundwater interactions and the hydrological conditions in a lake of the Dry Pampean Plain (DPP), i.e., Lake Los Pocitos, to understand the mechanism that guarantees such a resource. Results indicate that evaporation mainly controls the isotopic composition of lake water, overwhelming the effect of higher rainfall inputs during the wet (but also most evaporative) season. The  $\delta^{18}\text{O}$  mass balance model indicates greater groundwater inflow to the lake during the dry season ( $\sim 0.4 \text{ m month}^{-1}$ ) compared to the wet season ( $\sim 0.2 \text{ m month}^{-1}$ ). Lake level decreased in the wet season due to the lowest groundwater inflow and the greatest evaporation rate. Based on the proportion of water entering a lake that leaves through evaporation, Los Pocitos corresponds to a throughflow lake with a short water residence time ( $\sim 0.47$  years). These hydrologic conditions, along with freshwater inputs from a dune located at the western margin of the lake, determine the existence of this relatively stable and freshwater lake in the DPP where high evaporation rates are registered.

**Keywords** Groundwater/surface-water relations · Environmental tracers · Mass balance method · Water residence time · Argentina

## Introduction

Lakes constitute one of the most sensitive ecosystems to anthropic impacts (Kidmose et al. 2013; Tao et al. 2013; Wetzel 2001). Often interactions with groundwater are central to lake hydrology, biogeochemistry, and ecological functioning, with natural and anthropic perturbations on aquifers

relevant for lakes (Arnoux et al. 2017a; Shaw et al. 2013). This connection is crucial in dry sandy land regions where lakes are scarce and more reliant on groundwater supply and constitute important water resources for wildlife and humans (e.g., Bouchez et al. 2016; Harvey et al. 2007; Williams 1999). The estimation of groundwater inflow and outflow in lakes is often dismissed due to the difficulty of its quantification (e.g., Arnoux et al. 2017b). During the past decades, numerous studies have demonstrated that these interactions are of great importance not only for the water balance (e.g., Liao et al. 2018) but also for nutrient dynamics (Ala-aho et al. 2013; Brock et al. 1982; Lewandowski et al. 2015; Meinikmann et al. 2015; Oliveira Ommen et al. 2012). Thus, the understanding of groundwater-lake interactions is important for the effective management of water resources and aquatic ecosystems (e.g., Winter et al. 1998).

Different methods have been used to quantify groundwater/surface-water interactions. Among them, dissolved conservative constituents such as chloride and stable isotopes in

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the water molecule, have been widely used for this purpose (e.g., Bocanegra et al. 2013; Campodonico et al. 2019; Cao et al. 2018; Gibson et al. 2016; Gurrieri and Furniss 2004; Hofmann et al. 2008; Montalván et al. 2017; Petermann et al. 2018; Rosenberry et al. 2015; Shaw et al. 2017; Stets et al. 2010; Turner and Townley 2006). Stable isotopes are also valuable for estimating the water residence time (WRT) of lakes (Brock et al. 1982; Gat 2010; Gibson et al. 2016; Gibson et al. 2002; Gonfiantini 1986; Turner et al. 2010). Conceptually, based on the ratio of evaporation ( $E$ ) to total water inputs ( $I$ ), three hydrological conditions for lakes can be defined: desiccating ponds, terminal lakes, and throughflow lakes (e.g., Gibson et al. 2016). In the first hydrological setting, inflow occurs once or sporadically, while conversely, in terminal lakes, inflow is continuous and long-term evaporation balances inflow, thus no liquid outflow occurs. Finally, throughflow lakes are those with continuous inflow, which is balanced with the outflows represented by evaporation and lake water outflow. In this latter case, groundwater input is higher than lake surface evaporation (Gat 1995; Gibson et al. 2002).

The Pampean Plain (PP) in Argentina has gentle slopes and more than one hundred shallow lakes that developed on geomorphological depressions that originated mainly by wind deflation during the Quaternary (Tripaldi and Forman 2007) and shallow water tables. Besides, during the last few decades, a water-table rise has been registered in this area due to land-use change related to the replacement of native vegetation by crops (e.g., Jobbágy et al. 2021). The hydrology of these lakes depends on rainfall, groundwater discharge, potential evaporation, and the topographic position (e.g., Bocanegra et al. 2013; Drago and Quiros 1996; Romanelli et al. 2014). This region of Argentina has a marked rainfall gradient from east to west, whereby in the east, the Humid Pampean Plain (HPP) has an annual rainfall >1,000 mm, and the west region of the Dry Pampean Plain (DPP) presents an annual rainfall <700 mm (Iriondo et al. 2009). In semiarid zones like the DPP, lakes are natural groundwater outcropping areas. Lakes located in the study area, known as “Mercedinas” due to their proximity to the city of Villa Mercedes (San Luis Province), constitute the only surface-water bodies in this area, where water is scarce, and they provide an essential habitat for aquatic fauna and flocks of water birds. They also offer crucial water sources for livestock and have shaped human displacement, roads and settlements throughout the Holocene until recent times (Heider et al. 2019). Currently, these lakes are exposed to several anthropic impacts such as recreational activities, including fishing, and the modification of their surrounding landscape through cattle grazing of native grasslands and cultivation.

In closed-basin lakes such as those located in the DPP, a modification in the land use of the surrounding landscape

can lead to an increase in groundwater recharge and water level rise, as documented when dryland agriculture replaces native vegetation, or water degradation due to below-ground nutrient inputs from fertilized plots (e.g., Blarasin et al. 2020; Nosetto et al. 2015; Santoni et al. 2010). Likewise, rise of the water table can generate salt leaching from soils into surface-water and groundwater systems (e.g., Jayawickreme et al. 2011). Furthermore, in the current environmental context, both climate and land-use change can lead to water-table fluctuations, which influence the dynamic of these kinds of lakes (Córdoba et al. 2014; Guerra et al. 2015; Liu et al. 2013; Nosetto et al. 2015; Tao et al. 2013).

Several studies have shown that groundwater constitutes an important water source in shallow lakes of the HPP (Drago and Quiros 1996; Fernández Cirelli and Miretzky 2004; Miretzky et al. 2001; Miretzky et al. 2000; Romanelli et al. 2014). However, this connectivity was not explored in the driest portion of the PP, and it may be particularly important in sand dune landscapes where an installed drainage network (permanent or ephemeral) does not exist. Moreover, in spite of the drier climate, lakes appear to be deeper and more stable than in the HPP, even during dry periods when most lakes across the plain desiccate (Vilanova et al. 2015). The main goal of this work is to establish surface-water/groundwater interactions, by means of stable isotopes, in one of the lakes of the sand dune system of the DPP known as Lake Los Pocitos. For this purpose, the  $\delta^{18}\text{O}$  mass-balance model was applied to quantify the groundwater discharge flux into the lake and the  $E/I$  ratio was used to establish the hydrological conditions of Lake Los Pocitos. Thus, this study will contribute to the understanding of the role of groundwater in sustaining the Lake Los Pocitos ecosystem, and provide an insight on the persistence of lakes with low total dissolved salts, particularly in an area of high evaporation rates. Furthermore, the information generated in this contribution will help to elucidate the hydrological functioning of water bodies located in other semiarid and arid regions such as the Nebraska Sand Hills in the USA (e.g., Rossman et al. 2019) and the Badain Jaran Desert in China (e.g., Liu et al. 2016).

## Lake Los Pocitos region

Semiarid dune fields and sandy mantles of late Quaternary age cover most of the DPP, in west-central Argentina (Iriondo and Kröhling 1995; Zárate and Tripaldi 2012). The regional topographic slope is northwest-southeast. Ceci and Coronado (1981) established, based on a few drilled wells and several vertical electrical soundings, that the crystalline basement is located at 200 m depth in the S and W from the city of Villa Mercedes. The thicknesses of the saturated sedimentary aquifer formations and their hydrological parameters are mostly unknown. The groundwater regional

flow follows the same direction of the regional topographic slope (NW-SE). The hydraulic gradient in the western part of the basin is 28% and in the southeastern is 30% (Ceci and Coronado 1981).

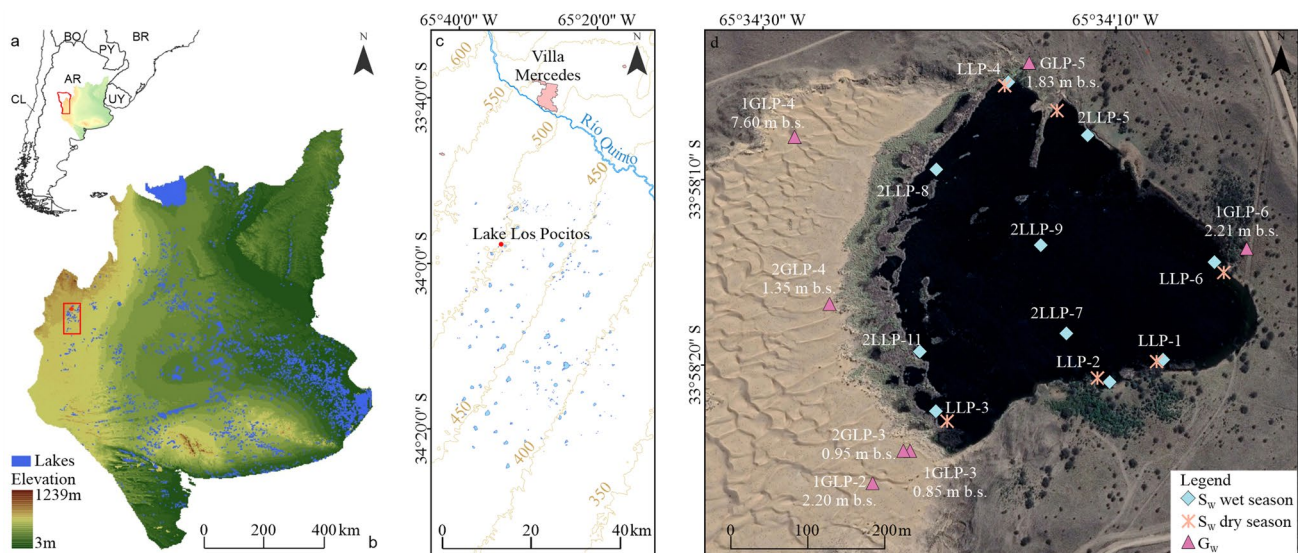
The only major permanent surface drainage is the Quinto River. However, the water table reaches the topographic surface in the region and floods the topographically lower areas, filling blowout depressions. In this sense, more than 200 seepage lakes are distributed over the sandy plain in deflation pans (Tripaldi et al. 2013; Vilanova et al. 2015). Lake Los Pocitos is one of these shallow inland water bodies (Fig. 1).

Lake Los Pocitos (33°58'12.77" S - 65°34'16.07" W) is a lake located south of the Quinto River, about 45 km away from the city of Villa Mercedes (Fig. 1c). It is located in the northwestern portion of the lake system. Its surface area is ~0.2 km<sup>2</sup>, the maximum water depth is about 11 m, with an average water depth of about 2 m. Since it has no surface-water drainage networks, it gains water mainly via rainfall and groundwater discharge, and loses water primarily through evaporation and groundwater outflow. Groundwater outflow was observed by means of the environmental tracer <sup>222</sup>Rn (Echegoyen et al. 2021). Los Pocitos waters are alkaline (mean pH 9.1), with low salinity (mean TDS 651 mg L<sup>-1</sup>), and of the HCO<sub>3</sub><sup>-</sup>-Na<sup>+</sup>-K<sup>+</sup> type (Echegoyen et al. 2021).

The Lake Los Pocitos area is in the denominated Western Pampean dunefields (Zarate and Tripaldi 2012). This region is characterized by a low relief plain that includes well vegetated and diverse stabilized dunes surrounded by

discontinuous aeolian sands mantles (Tripaldi and Forman 2007). The surrounding geology is dominated by aeolian sand deposits composed of lithic feldsarenites and feldspathic litharenites, where volcanic grains are by far the most common rock fragments (Tripaldi et al. 2010). Dunes show two main orientations in this area. Parabolic dunes (currently vegetated) suggest southeasterly paleowinds, whereas barchanoid ridges indicate a recent dune reactivation with winds coming from the northeast (Tripaldi et al. 2013). The compound parabolic dune at Lake Los Pocitos is a 3.5 m thick exposure of well-sorted fine sand, subdivided into two units according to the sedimentary structures (Tripaldi et al. 2013). The lower unit (~1 m thick) exhibits horizontal laminated sand deposited by wind ripple migration, and is covered by the upper unit (~2 m thick) which corresponds to a cross-laminated bed-set originated by dune migration (Tripaldi et al. 2013). Barchanoid ridges at the western margin of the lake evidence an aeolian reactivation where the dunes migrate in a NE-SW direction. These sandy sediments have relatively weak pedogenesis (Colazo 2012).

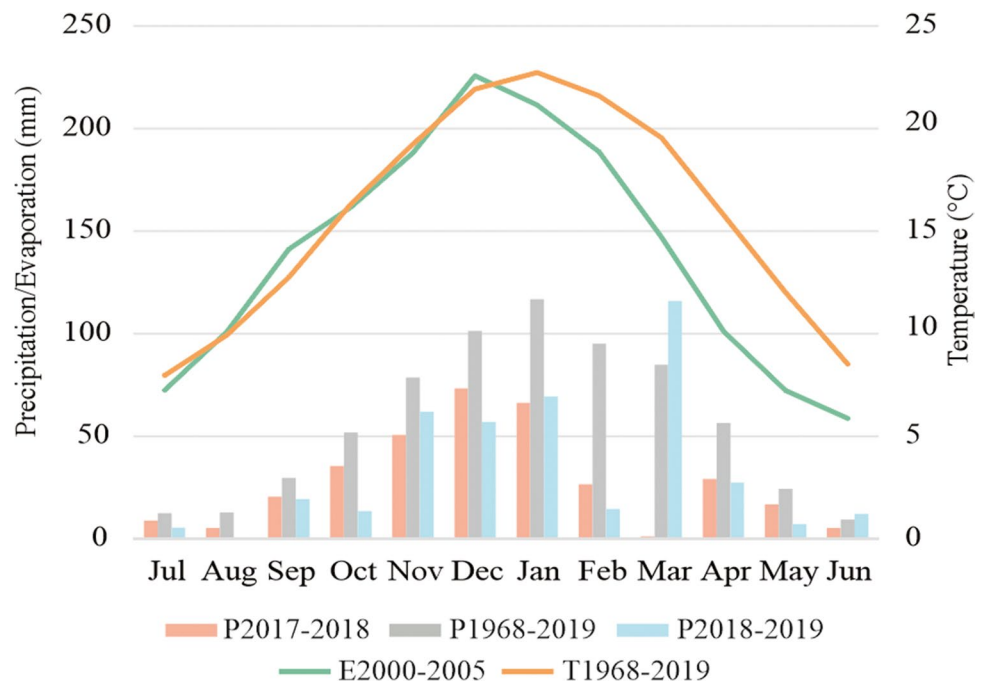
Natural grasslands prevail in the area (Colazo 2012) with different degrees of alterations caused by cattle grazing (Viglizzo and Jobbágy 2010). Native chañar (*Geoffroea decorticans*) and caldén (*Prosopis caldenia*) trees have progressively encroached on the area and are particularly abundant on the eastern (downslope) margin of the lake. The lake shoreline and shallow margins are occupied by marshes dominated by *Cortaderia*, *Schoenoplectus*, and *Typha* species.



**Fig. 1** **a** Location map of the San Luis Province and PP in Argentina. **b** Location of Mercedinas lakes and Lake Los Pocitos in the PP. **c** Mercedinas lakes, located south of the Quinto River and the city of Villa Mercedes (Lake Los Pocitos is located in the northwestern

portion of the lake system). **d** Location of sampling points of surface water (S<sub>w</sub>) and groundwater (G<sub>w</sub>) during dry (September 2017) and wet (April 2019) seasons

**Fig. 2** Monthly mean rainfall ( $P$ ) and temperature ( $T$ ) series at the city of Villa Mercedes (1969–2019 record period). Evaporation ( $E$ ) measured from evaporated ponds (2000–2005)



Climate in this region is semiarid, with marked seasons (Fig. 2). The mean annual rainfall is 679 mm (record period 1968–2018, INTA Villa Mercedes gaging station), and the mean annual temperature is 15.7 °C. Rainfall occurs mainly during the austral spring and summer, between October and March (>50 mm/month for the same record period), whereas the austral winter has the lowest (<50 mm/month) rainfall inputs. Because of the highest temperatures, evaporation rates peak simultaneously with rainfall, and are maximum during the austral summer (Marchesini et al. 2020). Figure 2 shows that evaporation exceeds rainfall all the year, with an annual evaporation rate measured from evaporated ponds of 1,541 mm year<sup>-1</sup> (for the record period 2000–2005 at the city of Villa Mercedes). The average relative humidity is 66%.

## Materials and methods

### Lake evaporation estimation ( $E$ )

Evaporation was estimated for each season (September 2017 and April 2019) by the Penman combination method (Rosenberry et al. 2007). Daily average air temperature, relative humidity, wind speed, and solar radiation data were obtained from the INTA Villa Mercedes gaging station. Water-surface temperature was measured directly in the field.

### Sampling and analysis

Surface-water samples were collected from Lake Los Pocitos at a depth of about 0.3 m in both dry ( $n = 6$ , September

2017) and wet ( $n = 10$ , April 2019) seasons. During the two sampling campaigns, groundwater samples ( $n = 5$  and  $n = 2$ , for dry and wet seasons, respectively) from the phreatic aquifer were obtained through hand-dug wells. The depth at which the water level was reached varied between 0.8 and 7.6 m below surface (bs; water samples were retrieved when the water table was reached), covering the entire perimeter of the lake (Fig. 1d).

Field measurements were performed following standardized methods (e.g., Rice et al. 2012); these determinations included temperature ( $T$ ), electrical conductivity (EC), and total dissolved solids (TDS). Unfiltered samples for stable isotope determinations were stored in 50 ml polyethylene bottles filled to the top with water, and capped without leaving any headspace. Stable isotope measurements were performed using cavity ring-down spectroscopy (Picarro L2120-i) at the GEA-IMASL-CONICET Institute. Results are expressed as  $\delta\%$  according to Eq. (1).

$$\delta\% = 1000 \frac{(R_S - R_{V-SMOW})}{R_{V-SMOW}} \quad (1)$$

where  $\delta$  is the isotopic deviation in %,  $R$  is the isotopic ratio (<sup>2</sup>H/<sup>1</sup>H or <sup>18</sup>O/<sup>16</sup>O),  $S$  is the sample, and  $V-SMOW$  is the reference material (Vienna Standard Mean Ocean Water; Gonfiantini 1978). Analytical uncertainties were  $\pm 0.1\%$  for  $\delta^{18}\text{O}$  and  $\pm 0.5\%$  for  $\delta^2\text{H}$ .

The lake perimeter was determined with a Garmin global positioning system (GPS) equipment and lake depths were measured using a Speedtech Depthmate portable sounder (readings are accurate from 0.6 to 79 m). The

lake area and volume were calculated using geographical information systems (ArcGIS).

### Isotopic mass balance

Groundwater inflow to Lake Los Pocitos was determined by employing a mass balance method based on  $\delta^{18}\text{O}$  concentrations (Gibson et al. 2016; Gonfiantini 1986). Several studies have employed a similar approach to determine groundwater inflow into lakes. For instance, some recent works are those of Shaw et al. (2017) who used this model to determine groundwater inflow in a temperate seepage lake in Georgetown Lake in Montana (Canada), and Arnoux et al. (2017a), who applied this method in kettle lakes in Quebec (Canada). Recently, Campodonico et al. (2019) used the environmental tracer  $\delta^{18}\text{O}$  to quantify groundwater inflow into Laguna del Plata (Mar Chiquita system, Argentina). Steady-state mass balance equations assume that lakes are well mixed and maintain a long-term constant volume (Gat 2010; Gibson and Edwards 2002). Lake levels were reconstructed for Lake Los Pocitos using the available Sentinel-2 imagery from June 2016 to September 2020 (Fig. 3). For this period, a lake area of  $0.196 \pm 0.002 \text{ km}^2$  was calculated. The relatively constant lake area during this period, along with the depth of Lake Los Pocitos allows one to consider it a steady-state system for modeling purposes.

In absence of surface drainage networks, the water balance at Lake Los Pocitos is controlled by groundwater discharge into the lake, groundwater outflow from the lake, rainfall, and evaporation. Thus, its general water mass balance can be stated as follows:

$$\frac{dV}{dt} = G_i + P - G_o - E \quad (2)$$

where  $V$  is the water volume in the lake,  $t$  is time,  $G_i$  is groundwater inflow,  $P$  is precipitation,  $G_o$  is lake groundwater outflow and  $E$  is evaporation.

A lake isotopic balance can be performed by measuring or calculating the isotope compositions ( $\delta$ ) for all the components of the water balance equation. When combining each term of Eq. (2) with  $\delta^{18}\text{O}$  values, the following equation is obtained:

$$\frac{dV\delta_L}{dt} = G_i\delta_{G_i} + P\delta_P - G_o\delta_{G_o} - E\delta_E \quad (3)$$

Since water outflow from the lake does not cause isotopic fractionation, one can assume that isotopic compositions from groundwater outflow and the lake water are the same ( $\delta_{G_o} = \delta_L$ ; Sacks et al. 2014). Rearranging and combining Eqs. (2) and (3), one can calculate the groundwater inflow as follows:

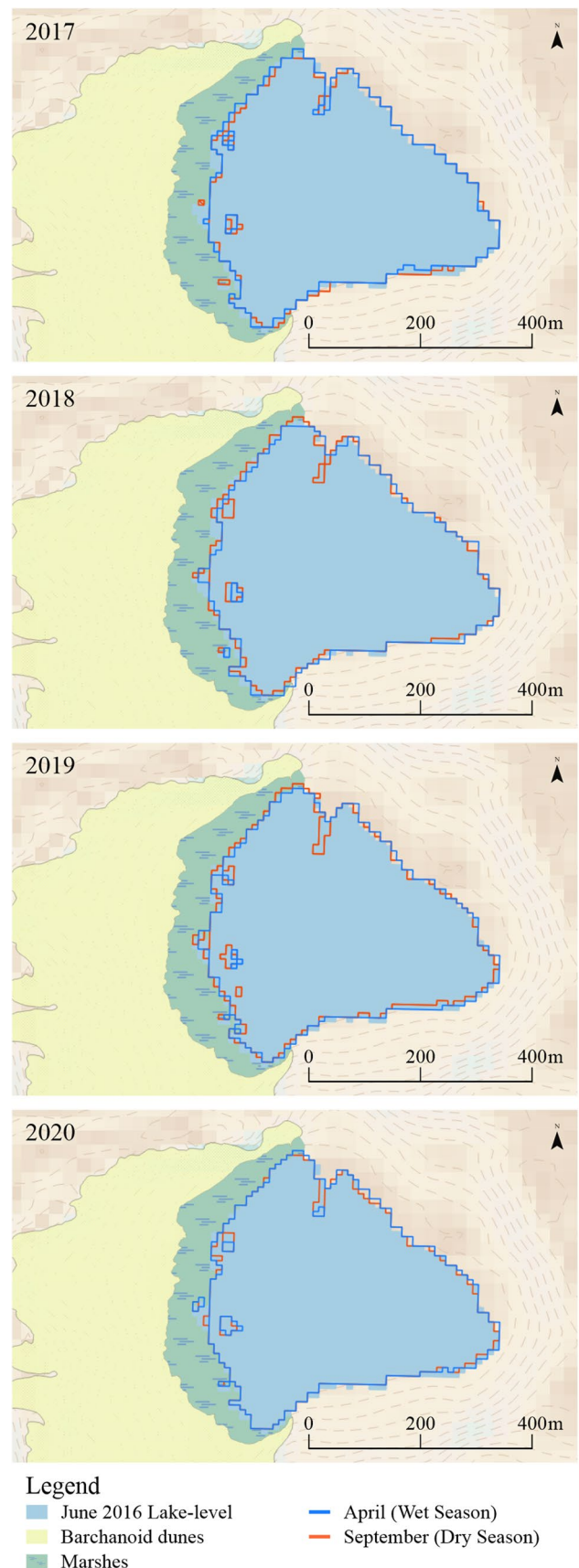


Fig. 3 Lake-level variation from June 2016 to September 2020

$$G_i = \frac{P(\delta_L - \delta_p) + E(\delta_E - \delta_L)}{\delta_{Gi} - \delta_L} \quad (4)$$

Most terms in Eq. (4) can be measured directly, with the exception of the isotopic composition of evaporated lake water ( $\delta_E$ ), which is difficult to measure. Thus,  $\delta_E$  can be estimated using the linear resistance model provided by Craig and Gordon (1965), which describes  $\delta_E$  (Eq. 5) as a function of the lake isotopic composition ( $\delta_L$ ), the relative humidity ( $h$ ), the isotopic composition of local atmospheric water vapor ( $\delta_A$ ), and fractionation factors.

$$\delta_E = \frac{\alpha^* \delta_L - h \delta_A - \epsilon}{1 - h + 10^{-3} \Delta \epsilon} \quad (5)$$

All fractionation factors were calculated using the experimental equations defined by Horita and Wesolowski (1994). The equilibrium isotope fractionation factor at the temperature of the air-water interface for  $^{18}\text{O}$  (i.e.,  $\alpha^*$ ) is given by:

$$\alpha^* = \frac{1}{\alpha_{L/V}} \quad (6)$$

where  $\alpha_{L/V}$  is defined by Gonfiantini (1986) as the ratio in liquid versus vapor, and can be estimated with Eq. (7):

$$\alpha_{L/V}(^{18}\text{O}) = \exp\left(\frac{-7.685}{10^3} + \frac{6.7123}{T} - \frac{1666.4}{T^2} + \frac{350410}{T^3}\right) \quad (7)$$

where  $T$  is the temperature in degrees K.

The kinetic fractionation factor ( $\Delta \epsilon$ ) depends on humidity, and can be calculated using Eq. (8).

$$\Delta \epsilon = K(1 - h) \quad (8)$$

where  $K(^{18}\text{O}) = 14.2\%$ , which is a value determined from wind tunnel experiments (Gonfiantini 1986).

The total fractionation factor ( $\epsilon$ ) is expressed in % and is estimated with Eq. (9):

$$\epsilon = \epsilon^* + \Delta \epsilon \quad (9)$$

where  $\epsilon^*$  is the equilibrium fractionation factor which depends on temperature and it is estimated with Eq. (10):

$$\epsilon^* = (1 - \alpha^*) + \Delta \epsilon \quad (10)$$

Finally, the isotopic composition of ambient atmospheric water vapor ( $\delta_A$ ) was calculated by means of Eq. (11) assuming equilibrium with precipitation (Gibson et al. 2008):

$$\delta_A = \frac{\delta_p - \epsilon^*}{1 + 10^{-3} \epsilon^*} \quad (11)$$

## Hydrologic conditions and lake water residence time

Stable isotopes can also be used to classify the hydrologic conditions of lakes such as desiccating water bodies, terminal lakes, or throughflow lakes (Gibson et al. 2002; Gonfiantini 1986). In well-mixed and steady-state conditions, the general water mass balance and the isotopic mass balance equations can be rearranged from Eqs. (1) and (2), and the fraction of total water inputs lost by evaporation ( $E/I$ ) can be calculated as follows:

$$\frac{E}{I} = \frac{\delta_L - \delta_I}{\delta^* - \delta_L} m \quad (12)$$

where  $\delta_L$  represents water discharged from the lake,  $\delta_I$  represents inflowing water,  $\delta^*$  is the limiting isotope enrichment factor, and  $m$  is the enrichment slope. Values of  $E/I$  between 0 and 1 reflect varying degrees of throughflow lakes,  $E/I = 1$  correspond to terminal lakes, and  $E/I > 1$  represent desiccating water bodies (Gat 1995; Gibson et al. 2016). These last types of lakes are transient systems ( $dV/dt < 0$ ) and therefore cannot be considered to be in a hydrological steady state (Gibson et al. 2002).

The limiting isotope enrichment factor ( $\delta^*$ ) is estimated with Eq. (13).

$$\delta^* = \frac{h \delta_A + \epsilon}{h - 10^{-3} \epsilon} \quad (13)$$

The enrichment slope ( $m$ ) can be calculated with Eq. (14).

$$m = \frac{h - 10^{-3} \epsilon}{1 - h + 10^{-3} \Delta \epsilon} \quad (14)$$

Water residence time ( $\tau$ ) is a determinant parameter of the ecological health of lakes, and it is defined as the average time that a water molecule spends in the system. For instance, WRT has also been related to the N removal efficiency by the ecosystem (e.g., Finlay et al. 2013).

The lake WRT can be calculated from the estimation of  $E/I$  and lake volume ( $V$ ), according to the equation proposed by Gibson et al. (2002):

$$\tau = \left(\frac{E}{I}\right) \left(\frac{V}{E}\right) \quad (15)$$

## Results and discussion

### Lake evaporation estimation ( $E$ )

Evaporation ( $E$ ) measured from evaporated ponds located at INTA Villa Mercedes gaging station was  $1,541 \text{ mm year}^{-1}$ , for the 2000-2005 period. These data are not available

for the sampling period, therefore the Penman method was used to calculate  $E$ . The obtained results were 79 and 176 mm month<sup>-1</sup> for the dry and wet seasons, respectively. The difference between the evaporation values estimated by means of the Penman method and those measured directly at INTA ponds is <10% for the 2000-2005 period.

### Isotopic composition

Stable isotopic compositions can be visualized in the conventional  $\delta^2\text{H}$  (%) vs.  $\delta^{18}\text{O}$  (%) plot (Fig. 4), jointly with the local meteoric water line (LMWL) of the city of Río Cuarto (located 150 km NE of Lake Los Pocitos), i.e.:  $\delta^2\text{H}$  (%) =  $8.3 \delta^{18}\text{O}$  (%) + 15.2 (Cabrera et al. 2019). The obtained results indicate that groundwaters were more depleted in heavy isotopes ( $\delta^{18}\text{O}$  = -5.9 to -3.2%;  $\delta^2\text{H}$  = -35 to -19%; Table 1) than surface waters ( $\delta^{18}\text{O}$  = -1.6 to 1.4%;  $\delta^2\text{H}$  = -10 to 3%; Table 1); besides, lake samples showed different isotopic compositions during both seasons. During the dry season, in September, the average stable isotopic composition of lake water was -1.3% for  $\delta^{18}\text{O}$  and -8% for  $\delta^2\text{H}$ , while conversely, during the wet season, in April, lake waters were more enriched in heavy isotopes, and showed mean values of 0.9 and 0% for  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$ , respectively.

All surface-water samples plot below the LMWL, evidencing evaporation processes and defining a local evaporation line (LEL), which can be explained by the equation:  $\delta^2\text{H}$  (%) =  $4.8 \delta^{18}\text{O}$  (%) - 3.6 ( $R^2 = 0.98$ ; Fig. 4).

As stated in the preceding, groundwater samples correspond to the shallowest portion of the phreatic aquifer. The groundwater samples were divided into shallow (<1.5 mbs) and deep (>1.5 mbs) waters according to their isotopic composition. Shallow groundwater presents  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  values of -4.3 to -3.2% and -27 to -19%, respectively. Except for the sample located at the east side of the lake (1GLP-6),  $\delta^{18}\text{O}$  values in deep groundwater around Lake Los Pocitos varied from -5.9 to -5.1%, and  $\delta^2\text{H}$  values varied between -35 and -27% (Table 1). The LEL origin coincides with the isotopic signatures of groundwater samples, suggesting that they constitute an important source of water from which surface water is evaporated. Three samples plot near the LMWL but over the LEL, and correspond to groundwater samples that were taken on the dune at shallow depths (<1.5 mbs). In arid environments with sandy soils where the water table is high, direct evaporation from a bare soil surface is an important factor to consider (Gat 1996). Groundwater sample 1GLP-6 on the eastern margin of the lake shows a similar isotopic composition to lake waters, reflecting the effects of capillary evaporation. The remaining groundwater samples showed a similar isotopic signature to rainfall, as they plot near the LMWL. These groundwater samples are deeper (>1.5 mbs) and indicate that groundwater around Lake Los

Pocitos has a dominant meteoric origin, whereas shallow groundwater is affected by evaporation processes (Fig. 4). This effect of evaporative losses on shallow groundwater (less than 1 mbs) was also observed in other points of the PP (Poca et al. 2020). Particularly, a deep groundwater sample (1-GLP-2) shows a more depleted isotopic composition, probably due to the influence of recharge coming from the western area according to the general behavior of Pampean rainfalls.

The isotopic composition of recharge water/rainfall into the lake ( $\delta$  recharge) is commonly estimated by the intersection of the LEL and the LMWL ( $\delta^{18}\text{O}$  = -5.4% and  $\delta^2\text{H}$  = -30%; e.g., Gat 1996; Gibson et al. 1993; Krabbenhoft et al. 1994). These isotopic values for rainfall are also similar to the mean values of the deepest groundwater sample ( $\delta^{18}\text{O}$  = -5.4% and  $\delta^2\text{H}$  = -30%). This is consistent with the fact that the isotopic composition of deep groundwater not affected by evaporation or mixing processes represents the average isotopic composition of rainfall in temperate climates (Fontes 1980).

### Groundwater inflow into Lake Los Pocitos: stable isotope mass balance

The groundwater inflow into Lake Los Pocitos was estimated using Eq. (4) during the dry and wet seasons. The rainfall value corresponds to the monthly rainfall registered at INTA Villa Mercedes gaging station for September 2017 and March 2019, representing the dry and wet seasons respectively ( $P_{\text{Dry}} = 28.5$  mm and  $P_{\text{Wet}} = 116$  mm). In addition, as stated in the preceding, the evaporation rate from Lake Los Pocitos was estimated for the dry and wet seasons using the Penman method ( $E_{\text{Dry}} = 79.0$  mm and  $E_{\text{Wet}} = 176.0$  mm).

The  $\delta^{18}\text{O}$  values for lake waters ( $\delta_L$ ) are -1.3 and 0.9% for the dry and wet seasons, respectively. Since Cabrera et al. (2019) observed a seasonal effect in the isotopic composition of rainfall at Río Cuarto city, two different  $\delta_p$  values were used for mass balance calculations. During the dry season, rainfall is more depleted ( $\delta_p = -5.7\%$ ), whereas in the wet period rainfall is more enriched in heavy isotopes, with a mean  $\delta^{18}\text{O}$  value of -4.5% (Cabrera et al. 2019). The  $\delta^{18}\text{O}$  value of groundwater ( $\delta_{\text{Gi}} = -4.6\%$ ) used for calculation purposes corresponds to the mean isotopic composition of samples collected around the perimeter of the lake, except for sample 1GLP-6 which clearly shows the effects of capillary evaporation. Both  $\delta^{18}\text{O}$  values (i.e.,  $\delta_p$  and  $\delta_{\text{Gi}}$ ) were multiplied by the lake surface area (0.2 km<sup>2</sup>).

Finally, the  $\delta^{18}\text{O}$  value of evaporating moisture ( $\delta_E$ ) was calculated for both seasons using the multilayered model (Eq. 5) developed by Craig and Gordon (1965) and considering the water salinity. Since  $\delta_E$  also depends on humidity, mean monthly humidity values were obtained from INTA

**Table 1** Physico-chemical parameters and isotopic composition of Los Pocitos Lake and groundwater samples. The groundwater sample depth is also included. *TDS* total dissolved solids; *EC* electrical conductivity; *T* temperature

Sample name	Depth (mbs)	<i>T</i> (°C)	EC (μS cm <sup>-1</sup> )	TDS (mg L <sup>-1</sup> )	δ <sup>18</sup> O (%)	δ <sup>2</sup> H (%)
1LLP-1		15.5	1,219	611	-1.1	-7
1LLP-2		15.8	1,218	611	-1.2	-8
1LLP-3		15.5	1,234	618	-1.0	-7
1LLP-4		16.3	1,169	587	-1.6	-10
1LLP-5		16.6	1,216	609	-1.5	-9
1LLP-6		16.7	1,333	668	-1.2	-7
2LLP-1		24.5	1,230	616	0.9	-1
2LLP-2		23.8	1,218	613	1.4	2
2LLP-3		23.7	1,218	610	1.1	2
2LLP-4		24.5	1,280	639	0.1	-5
2LLP-5		19.7	1,169	1,194	0.1	-3
2LLP-6		20.2	1,249	625	1.2	2
2LLP-7		23.6	1,238	620	1.3	3
2LLP-8		21.6	1,164	583	1.0	1
2LLP-9		21.1	1,180	591	1.3	2
2LLP-11		21.5	1,236	618	1.0	1
1GLP-2	2.20	25.2	89	177	-5.9	-35
1GLP-3	0.85	24.8	511	256	-3.7	-19
1GLP-4	7.60	19.4	472	236	-5.1	-27
1GLP-5	1.83	19.8	1,014	507	-5.2	-29
1GLP-6	2.21	Nd	1,651	826	0.1	-2
2GLP-3	0.95	Nd	332	166	-4.3	-27
2GLP-4	1.35	Nd	283	140	-3.2	-20

Villa Mercedes gaging station for September 2017 and March 2019. For the dry season, the  $\delta_E$  is  $-18.5\%$ , whereas for the wet season it is  $-7.8\%$ .

Results obtained from the isotope mass balance using  $\delta^{18}\text{O}$  values confirm that groundwater discharge into the lake occurs during both seasons (Table 2). During the dry season, a groundwater inflow of  $\sim 0.4 \text{ m month}^{-1}$  to Lake Los Pocitos was estimated, whereas for the wet season a lower groundwater inflow was detected, in the order of  $\sim 0.2 \text{ m month}^{-1}$ .

The groundwater discharge to Lake Los Pocitos during the dry season was previously demonstrated by Echegoyen et al. (2021) using the environmental tracer  $^{222}\text{Rn}$ . This method revealed that Lake Los Pocitos exhibits a differential hydrogeological behavior along its perimeter. Groundwater inflow was detected at the northwestern sector of the lake, whereas surface water discharged into the aquifer at the southeastern sector of the lake. The different groundwater inflow values obtained with both methods can be explained by the fact that they indicate  $G_i$  rates at different timescales. While the stable isotope mass balance methods reflect the average conditions during months to several years (during the entire WRT), radon-based approaches reflect a maximum period of 20 days (Arnoux et al. 2017b; Petermann et al. 2018). Moreover, it was considered that this period could be even shorter due to degassing of  $^{222}\text{Rn}$  to the atmosphere

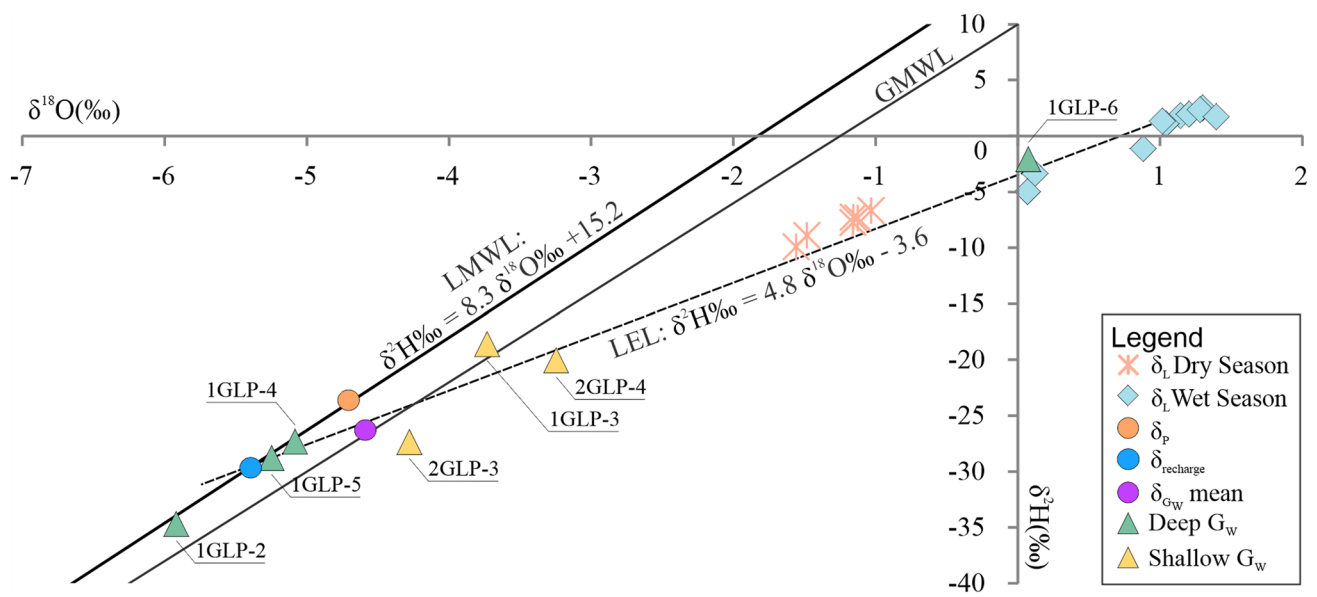
and wind-induced water mixing; thus, the authors consider that results obtained by means of both methods are consistent with different timescales of sensitivity.

Regarding the sources of error of the stable isotope mass balance, previous works (e.g., Krabbenholft et al. 1990; Shaw et al. 2017) have demonstrated that this method is more sensitive to the estimation of the isotopic composition of the evaporation flux, which depends on humidity and the isotopic value of atmospheric moisture. Thus, the model uncertainty may be increased due to the fact that these parameters are estimated by indirect methods. Despite these sources of error, the employed method is adequate as a first assessment of groundwater inflow quantification in these types of lakes where previous studies have not considered the groundwater component.

### Hydrologic conditions and lake water residence time

The proportion of the total water input that is lost through evaporation was calculated using Eq. (12). The  $E/I$  ratios of Lake Los Pocitos were 0.22 and 0.62 for dry and wet seasons, respectively, indicating that about 22% of lake water was lost through lake surface evaporation during the dry period, whereas during the wet period the lake water lost by evaporation increased to 62%. Values of  $E/I$  between 0 and





**Fig. 4** Isotopic composition ( $\delta^2\text{H}$  vs.  $\delta^{18}\text{O}$ ) of Lake Los Pocitos ( $\delta_L$ ) and associated groundwaters. The LMWL from the city of Rio Cuarto (Cabrera et al. 2019) and the mean isotopic composition of local rainfall ( $\delta_P$ ) are also included

approaching 1 (which corresponds to 0-100% respectively) reflect varying degrees of throughflow lakes (Gat 1995; Gibson et al. 2016; Turner et al. 2010), where inflow is continuous and balanced by outflow (Gibson et al. 2002).

In the PP, dunes, especially those without vegetation, act as preferential recharge areas where freshwater from rainfall can locally form important groundwater lenses in the upper part of the aquifer, surrounded by salty groundwater (Blarasin et al. 2014; Jobbágy et al. 2011). This occurs in the dunes of Los Pocitos area, where barchanoid dunes act as preferential recharge zones. Here, groundwater measurements show EC values  $<550 \mu\text{S cm}^{-1}$ . Regional studies performed on rainwater samples from Río Cuarto station (2006-2012) have shown EC values of  $\sim 54 \mu\text{S cm}^{-1}$  and that they correspond to the calcium sulfate or calcium carbonate type (Cabrera et al. 2013).

The throughflow condition of the lake, along with freshwater inputs from the dune, can explain the existence of a nonsalty lake like Los Pocitos (mean TDS  $651 \text{ mg L}^{-1}$ ) in a

semiarid region in which high evaporation rates would warrant a fast concentration of solutes in any stagnant water body.

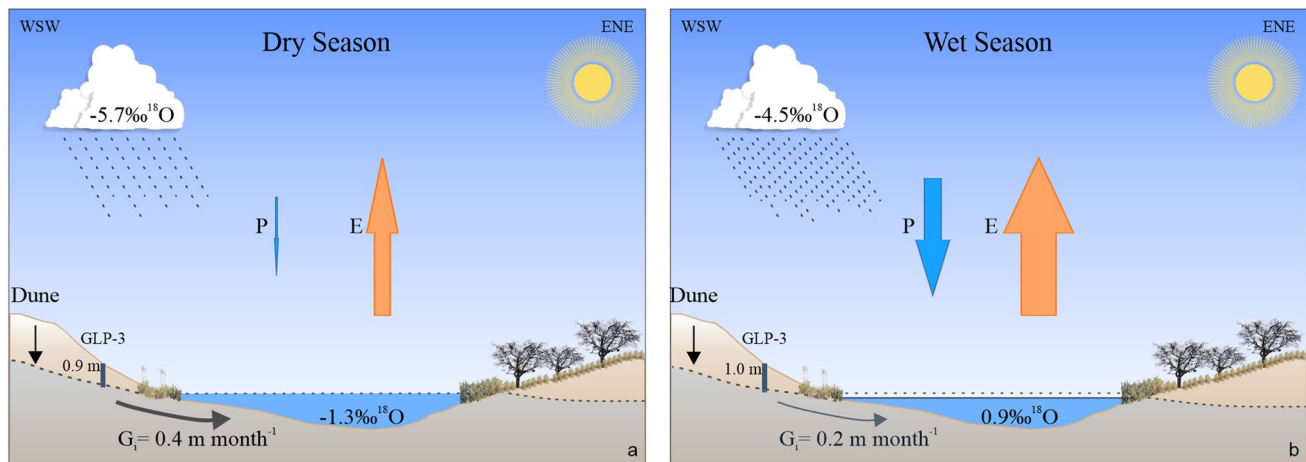
Figure 5 corresponds to a schematic diagram depicting the different components of the hydrological budget for this lake system. Evaporation exceeds rainfall during both seasons. Although groundwater inflow occurs during both seasons, it is lower during the wet season. Despite the fact that this is the season when the highest rainfall is recorded, due to lower groundwater inflow in the lake and a greater loss of water by the intense evaporation, a decrease of 10 cm in the lake level was observed.

Water residence time, which was calculated taking into account evaporation, precipitation, the absence of an installed drainage network that flows into the lake, and the lake volume, indicates how rapidly the water in the system is replaced. WRT of 0.28 and 0.67 years were estimated for Lake Los Pocitos for dry and wet seasons, respectively. Low values (i.e.,  $<1$  year) indicate a short WRT in the system, which generates a constant exchange of water. Moreover, the low WRT inhibits the interaction between water and rocks,

**Table 2** Input data for stable isotope mass balance and evaporation loss in Lake Los Pocitos

Season	Area (m <sup>2</sup> )	T (°C)	h	P (mm m <sup>-1</sup> )	E (mm m <sup>-1</sup> )	$\delta_L$ <sup>18</sup> O (‰)	$\delta_P$ <sup>18</sup> O (‰)	$\delta_{Gi}$ <sup>18</sup> O (‰)	$\delta_A$ (‰)	$\delta_E$ (‰)	$G_i$ (m month <sup>-1</sup> )	$G_o$ (mm day <sup>-1</sup> )
Dry	195,884	12.7	0.62	21	79	-1.3	-5.7	-4.6	-16	-19	0.4	12.8
Wet	195,185	17.9	0.74	116	176	0.9	-4.5	-4.6	-14	-7.8	0.2	5.5

T: air temperature; h: relative humidity; P: precipitation; E: evaporation;  $\delta_L$ :  $\delta^{18}\text{O}$  composition of lake waters;  $\delta_P$ :  $\delta^{18}\text{O}$  composition of precipitation;  $\delta_{Gi}$ :  $\delta^{18}\text{O}$  composition of groundwater;  $\delta_A$ :  $\delta^{18}\text{O}$  composition of atmospheric water vapor;  $\delta_E$ :  $\delta^{18}\text{O}$  composition of evaporated lake water;  $G_i$ : groundwater inflow



**Fig. 5** Schematic hydrological model for Lake Los Pocitos during **a** dry and **b** wet seasons. The arrow thickness is proportional to the component contribution. The  $\delta^{18}\text{O}$  values for rainfall ( $\delta_p$ ) are those reported by Cabrera et al. (2019) for Rio Cuarto station

reducing the intensity of weathering processes, which favors a low salt concentration. The WRT values calculated for Lake Los Pocitos were lower than those reported for other lakes in the HPP ( $\tau = 1.11$  years) by Quiroz Londoño et al. (2020). These differences may be attributed to the existence of different climatic regimes in the PP. For instance, within the USA, aridity is an important factor controlling  $E/I$  and  $\tau$ , with relative humidity and annual precipitation being the main drivers (Brooks et al. 2014).

Different studies (Brock et al. 1982; Wolfe et al. 2007) established that throughflow lakes and, to a lesser extent, those with relatively short WRT, show good biological conditions and low concentrations of nitrogen, phosphorus, and dissolved organic carbon. Thus, the hydrological conditions determined for Lake Los Pocitos could explain the good water quality observed in this water body.

Unfortunately, the environmental quality of water bodies in this region is at risk of deterioration in response to disturbances caused by human activities. Changes in the hydrologic balance as a result of the conversion of native vegetation to areas devoted to agriculture have been reported for instance in Australia, Southern High Plains in the USA, and the Sahel plains in North Africa (George et al. 1997; Leblanc et al. 2008; Scanlon et al. 2005). A widely observed pattern of rising water levels and soil salinity has been reported in such semiarid sedimentary plains. This does not only cause the loss of cultivable hectares, but also the deterioration of ecosystems that depend on groundwater (George et al. 1997; Williams 1999). A widespread water-table rise accompanied by incipient salinization of soils, phreatic groundwater, and surface waters has also been recorded in the DPP (Bogino and Jobbágy 2011; Jayawickreme et al. 2011; Santoni et al. 2010). Even the appearance of deep canyons and streams in the western edge of the DPP caused by sapping has been

related to land use change (i.e., replacement of native vegetation by crops) and the consequent water-table rise (Contreras et al. 2013; Jobbágy et al. 2021). Particularly, in the study area, the anthropic activities have not significantly impacted the water quality and lake levels which have remained relatively constant during the last few years. However, if human activities continue without control and remediation policies, it is probable that these modifications could generate a negative impact on the ecosystem in a short time.

The surface-water and groundwater degradation due to the use of fertilizers is another risk associated with agriculture expansion that has become an environmental problem in many regions of the world. In Argentina, N-containing fertilizers have been utilized inefficiently, and large quantities are frequently discharged into aquatic ecosystems causing eutrophication (e.g., Licursi et al. 2016). N-degraded groundwater and lakes have been recognized in the PP (Blarasin et al. 2020; Romanelli et al. 2020). Lakes receive water inputs from nitrate-contaminated groundwater, which can cause ecosystem deterioration. WRT is an important factor related to the nitrogen (N) self-purification capacity of aquatic systems (e.g., Finlay et al. 2013; Tong et al. 2019). For instance, a longer WRT results in higher N removal efficiency via denitrification or permanent burial in water bodies (Finlay et al. 2013).

The results obtained in this work give an idea of the hydrological functioning of water bodies emplaced in the DPP and their possible response to anthropic disturbances. Thus, establishing the surface-water/groundwater interactions, the type of lake, and the WRT in Lake Los Pocitos results in a relevant consideration of the changes in land use that have occurred in the DPP in the last few decades. For instance, the ecosystem N removal efficiency via denitrification or permanent burial would be affected by the short WRT in Lake Los Pocitos. The preservation of groundwater

quality is crucial in this region, where a high percentage of total inflow to surface-water bodies comes from this source.

## Concluding remarks

Lakes where the evaporation rate exceeds rainfall are usually desiccating ponds. However, Lake Los Pocitos has maintained its water level through the last decade, showing a relatively minor seasonal fluctuation. The  $\delta^{18}\text{O}$  mass balance model suggests a groundwater inflow of  $\sim 0.4$  and  $\sim 0.2$  m month<sup>-1</sup> to Lake Los Pocitos during the dry and the wet season, respectively. These results confirm that groundwater discharge into the lake occurs during both seasons, allowing a relatively constant water level throughout the year. During the wet season, in spite of the highest rainfall rate, the lowest groundwater inflow and the most intense evaporation causes a lake-level decrease of  $\sim 10$  cm compared to the dry season, when it increases again.

Based on the ratio of water inputs to the lake and surface evaporation, Lake Los Pocitos is classified as a throughflow lake with uninterrupted inflows that exceed loss by evaporation, and yield a WRT of approximately half a year. These hydrologic conditions, along with freshwater inputs from the dune, explain why, in spite of the aridity of the region, low water salinity and stable lake levels prevail.

This study constitutes a first approach to the quantification of the groundwater inflow for a large group of lakes located in the DPP, helping researchers to understand the lake sensitivity to changes in groundwater resulting from the ongoing climate and land-use changes and supporting decision-makers in the development of water management policies in these landscapes. In addition, the results indicate that these lakes offer an ideal opportunity to observe the impact of land-use changes in this semiarid region, which is part of one of the major agricultural regions in the world. Future research efforts should address the effects of land-use change on the hydrological cycle of the DPP.

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## Declarations

**Conflict of interest** On behalf of all authors, the corresponding author states that there is no conflict of interest.

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