

1 ***Integrating shotgun metagenomics, 16S rRNA gene metabarcoding and culture***
2 ***approaches: a better outlook for functional profiling of a PAH-contaminated soil***

3 Festa Sabrina^{1*}, Granada Marina¹, Irazoqui José M.², Cuadros-Orellana Sara³, Quevedo Claudio³,
4 Coppotelli Bibiana M.¹, Morelli Irma S.^{1,4}

5 1-Centro de Investigación y Desarrollo en Fermentaciones Industriales, CINDEFI (UNLP, CCT-La Plata,
6 CONICET), La Plata, Argentina

7 2- Instituto de Investigación de la Cadena Láctea, IdICaL (INTA-CONICET). Rafaela, Santa Fe, Argentina

8 3- Facultad de Ciencias Agrarias y Forestales, Centro de Biotecnología de los Recursos Naturales,
9 Universidad Católica del Maule, Talca, Chile

10 4- CIC-PBA, La Plata, Argentina

11 *Corresponding author: Sabrina Festa
12 CINDEFI, Street 50 N°227, 1900 La Plata, Argentina
13 E-mail: sfesta@biotec.quimica.unlp.edu.ar
14 ORCID: 0000-0002-3683-9305

15 **Key words**

16 polycyclic aromatic hydrocarbon, shotgun metagenomics, *Mycolicibacterium*,
17 enrichment culture

18 **Abstract**

19 Understanding bacterial diversity and function is critical for designing bioremediation strategies.
20 This research aimed to assess chronically hydrocarbon contaminated soil bacterial diversity and
21 their aromatic compound degradation (ACD) potential by integrating shotgun metagenomic, 16S
22 rRNA gene metabarcoding and culture approaches. While soil metabarcoding showed
23 dominance of Proteobacteria, metagenomics indicated that 99,5% of the sequences were
24 taxonomically assigned to Streptomycetales order and that almost all genes related to ACD were
25 assigned to the latter. To inspect other phyla contribution to ACD, a functional prediction was
26 delved, and two culture approaches were used. PICRUSt2 revealed that ACD pathways were
27 mostly found in Alphaproteobacteria, Actinobacteria and Gammaproteobacteria classes. An
28 enrichment culture (r-EFP) was obtained with pyrene as sole carbon and energy source and a
29 bacterial strain (S19P6), identified as a member of *Mycolicibacterium* genus, was isolated. Both
30 cultures demonstrated the ability to degrade more than 90% of the supplemented pyrene after
31 21 days of incubation. 16S rRNA gene metabarcoding and shotgun metagenomics approaches
32 in r-EFP indicated predominance of Proteobacteria Phylum and the presence of genes
33 responsible for the degradation of ACD mostly assigned to the predominant phyla.
34 Complementing different methodologies enable the recognition of the metabolic potential of
35 soil Proteobacteria related to ACD.

36 **Environmental implication**

37 Polycyclic aromatic hydrocarbons (PAH) are chemical compounds that are hazardous and
38 persistent in the environment. Due to its toxicity, strategies for removing these pollutants are
39 required. Microbial processes are considered one of the most significant and influential in PAH
40 removal; as a consequence, it is crucial for designing bioremediation strategies to understand
41 microbial functional and phylogenetic diversity. This investigation makes a major contribution
42 to research regarding the bacterial processes involved in a chronically contaminated soil by
43 integrating different methodological approaches to try to explain an observed limitation in PAH
44 degradation when applying different bioremediation strategies on a PAH contaminated soil.

45

46 **Introduction**

47 Microbial processes are considered one of the most significant and influential in polycyclic
48 aromatic hydrocarbon (PAH) removal (Miller *et al.*, 2004). The understanding of microbial
49 functional and phylogenetic diversity within a contaminated niche has been recommended as
50 critical for risk assessments, microbial monitoring and for designing bioremediation strategies
51 as a consistent approach to clean up environmental pollutants (Gosai *et al.*, 2018). To face
52 anthropogenic contamination, exploring indigenous microbial communities adapted to these
53 environments over long periods of time is expected to provide significant information (Scoma *et al.*, 2017). This implies advanced methodologies like molecular tools able to provide in-depth
54 understanding about the aspects of microbial processes and survival under stressed
55 environment (Mishra *et al.*, 2021).

57 The diversity and metabolic capacity of soil microbiome can be inferred from 16SrRNA gene
58 metabarcoding and described by shotgun metagenomics. Gene marker approaches coupled to
59 functional prediction software's like PICRUSt rely on the correlation between phylogenetic trees
60 and clusters of genes shared between taxa (Langille *et al.*, 2013). Shotgun metagenomics, on the
61 other hand, provides a direct assessment of the functional attributes of the microbiome (Jovel
62 *et al.*, 2016). However, metagenomics centred approaches alone could not be sufficient to reveal
63 functions belonging to the prokaryotic rare biosphere because most low abundance prokaryotes
64 remain uncultured and consequently understudied (Pascoal, Magalhães and Costa, 2020).
65 Cultivation-based approaches can acquire microorganisms that are undetected by molecular
66 methods and furthermore allow for the verification of metabolic potential determined by the
67 metagenomic data (Prakash *et al.*, 2013). The use of both molecular and culture-based methods
68 gives the opportunity not only to collect information on the composition of microbial
69 communities but also to isolate microorganisms capable of developing a specific activity
70 (Besaury *et al.*, 2013). Isolating environmental microorganisms and studying their physiology
71 under controlled conditions are essential aspects of understanding their ecology (Bartelme *et al.*, 2020). Culture is needed to describe novel species as a function and as the *sine qua non*
72 condition for understanding how the microbial world functions (Alain and Querellou, 2009). That
73 is why an integrative approach should be considered while assessing environmental microbial
74 diversity (Suenaga *et al.*, 2014).

76 Related isolation techniques are enrichment cultures, in which environmental conditions are
77 tailored to favor certain genotypes (that cross-feed essential metabolites), resulting in self-
78 assembled microbiomes (Souza, Shitut and Kost, 2018; Lawson *et al.*, 2019). While natural
79 consortia are still difficult to scrutinize, enrichment cultures offer a compromise between

80 natural and synthetic communities (Ponomarova and Patil, 2015) and a good model to study the
81 ecological niche of key taxa and the interactions that can occur between bacterial populations
82 during a bioremediation process (Desai, Pathak and Madamwar, 2010).

83 This paper attempts to explore the genetic potential of the bacterial community of a chronically
84 PAH contaminated soil. This soil, called IPK, came from a petrochemical area near La Plata city,
85 Argentina. This soil belonged to a landfarming unit where an alkaline petrochemical sludge (with
86 high PAH concentration) was treated; this landfarming unit was closed almost twenty years ago.
87 When sampled, PAH of three and four rings were detected in IPK soil. Different treatments have
88 been studied in order to select a strategy that can achieve the elimination of residuals PAH.
89 Allochthons bioaugmentation and surfactant-enhanced bioremediation were assayed and a
90 limitation in the degradation of contaminants was observed (Festa *et al.*, 2016; Cecotti *et al.*,
91 2018). To try to understand the observed limitation a metagenomic approach was performed.
92 Furthermore, enrichment cultures could serve as a tool for identification of microbial degraders
93 and metabolic pathways when biological components involved in hydrocarbon degradation in
94 the environment are below detection limits of the techniques used. That is why, considering
95 that the comprehension of the composition and metabolic potential of a soil bacterial
96 community could be also done by exploring bacterial cultures of that soil, and in light of
97 autochthonous bioaugmentation as a possible future strategy for decontamination, a culture
98 analysis was carried out. Therefore, this study integrates approaches to makes a major
99 contribution to research regarding the bacterial processes involved in a bioremediation.

100 **Material and methods**

101 **1 Soil characterization**

102 **1.1 Physicochemical properties**

103 A chronically PAH contaminated soil, called IPK, was collected from a petrochemical area near
104 La Plata city, Argentina. This soil belonged to a land farming unit where an alkaline petrochemical
105 sludge (with high PAH concentration) was treated. In order to identify the physicochemical
106 properties of IPK an analysis was carried out in the Laboratory of Soil Science at the University
107 of La Plata and showed a clay loam texture, pH of 8, 5.73% (w/w) organic carbon, 9,88% (w/w)
108 soil organic matter, 0,134 (w/w) total nitrogen, C/N ratio, 25 mg.kg⁻¹ available phosphorus and
109 Ca⁺², Mg⁺², Na⁺ and K⁺ 28,4, 10, 2,14, 1,1 (cmolc.kg⁻¹) respectively. All the different analysis
110 performed regarding IPK soil were carried out with sieved soil (decreasing pore size until in a 2-
111 mm mesh.

112 **1.2 Chemical hydrocarbon extraction and quantification and microbial enumeration**

113 Total PAH and aliphatic hydrocarbon in IPK soil were extracted according to Mora *et al.*, (2014)
114 with hexane:acetone (1:1), and the resulting extract was analyzed in a PerkinElmer auto system
115 gas chromatograph with a flame ionization detector (GC-FID) according to Del Panno *et al.*,
116 (2005). Microbial counts were determined as described in previous work (Festa, Coppotelli and
117 Morelli, 2016); in summary viable heterotrophic bacterial counts were determined in R2A agar
118 medium and incubated at 24±2 °C for 10 days. Also, the aromatic hydrocarbon-degrading
119 bacteria was determined microtiter plates, using liquid mineral medium (LMM) and incubated
120 24±2 °C for 21 days. Most Probable Number (MNP) method was used to enumerate PAH
121 degraders as described by Festa, Coppotelli and Morelli (2016).

122 **1.3 DNA extraction for qPCR, 16S rRNA gene metabarcoding and shotgun metagenomic**
123 **analysis**

124 Total soil DNA extraction was performed in triplicates using E.Z.N.A.® Soil DNA Kit (Omega Bio-
125 Tek, Inc., Norcross, GA, USA) following manufacturer's instructions. DNA was stored at -20°C
126 until analysis.

127 Different qPCR were carried out for IPK soil DNA in order to qualitatively and quantitative
128 characterize the genetic potential of the IPK microbial community, using the following *primers*:
129 1055F (5'-ATGGCTGTCGTCAGCT-3') and 1392R (5'-ACGGGCGGTGTGTAC-3') for 16SrRNA gene;
130 PAH-RHD α GN F (5'-GAGATGCATACCACGKGGTTGGA-3') and PAH-RH α GN R (5'-
131 AGCTGTTGTGGAAGAYWGTGCMGTT-3') designed for amplifying the α subunit of the PAH ring
132 hydroxylating dioxygenase gene of gram negative bacteria; and NidAF (5'-
133 TTCCCGAGTACGAGGGATAC-3') and NidAR (5'-TCACGTTGATGAACGACAAA-3') designed for
134 pyrene dioxigenase gene (NidA). Amplification program and used pairs of *primers* were designed
135 by Harms *et al.*, (2003), Cébron *et al.*, (2008) and Peng *et al.*, (2010) respectively.

136 Quantification of 16SrRNA and PAH-RHD α gene copy number was performed in triplicates in a
137 Rotor-Gene Q® (QIAGEN). For all qPCR assays three technical replicates on samples and negative
138 controls (PCR blanks with only the reaction mix and PCR blanks containing the mix and 1 μ l of
139 PCR-grade water) were used. Separate tubes in triplicate were used to measure threshold cycles
140 (Ct) and the identity and purity of the amplified product were checked by analyzing the melting
141 curve at the end of amplification. Reaction was carried out as described by Madueño *et al.*,
142 (2018). Results were expressed as gene copy number.g⁻¹ of dry soil.

143 **1.3.1 Metabarcoding of 16S rRNA gene**

144 16S rRNA metabarcoding analysis for IPK soil was carried out in triplicates using 16S rRNA
145 universal bacterial primers 515F (5' -GTGCCAGCMGCCGCGGTAA-3') and 806R (5'-
146 GGACTACHVGGGTWTCTAAT-3') to amplify V4 hypervariable region of 16S rRNA gene. Sample
147 sequencing was performed using the sequencing platform Illumina NovaSeq 6000 at Molecular
148 Research laboratory (MR DNA; Shallowater, TX) based on established and validated protocols
149 (<http://www.mrdnalab.com/>) as described in previous work (Festa *et al.*, 2016). Paired end 16S
150 rRNA reads were analyzed using QIIME2 2021.2 pipeline (Bolyen *et al.*, 2019), raw sequence data
151 was demultiplexed and quality filtered followed by denoising with DADA2 to obtain Amplicon
152 Sequence Variants (ASVs). A filtering approach was carried out to remove low-abundance
153 sequences (<0.1% of the average depth of the samples), mitochondrial/chloroplast 16s rRNA
154 sequences and reads that were not classified at the phylum level. After alpha rarefaction and
155 alpha diversity analysis, taxonomy composition was explored Silva 138 trained Naive Bayes
156 classifier 99% OTUs from 515F/806R region of sequences. Metabarcoding analysis was coupled
157 to PICRUSt2 software (v2.4.1) to predict the abundance of gene families KEGG (Kyoto
158 Encyclopedia of Genes and Genomes) orthologs (KOs) and higher-level pathways (MetaCyc) was
159 used (Douglas *et al.*, 2020). PICRUSt2 was executed following the pipeline established by the
160 software developers. Prediction accuracy was estimated by the Nearest Sequenced Taxon Index
161 (NSTI) scores. Bar plots, Bubble plots and Sankey Diagrams were generated using R v. 4.2.0 (R
162 Foundation for Statistical Computing) and ggplot2_3.3.6 (Wickham, 2016).

163 **1.3.2 Metagenomic analysis**

164 Shotgun metagenomics was performed in Corporación Corpogen (<https://www.corpogen.org/>),
165 using an Illumina Miseq sequencing platform. Raw reads were quality trimmed using

166 Trimmomatic (version 0.39; Bolger, Lohse and Usadel, 2014), to remove adapters and low
167 quality regions, and an estimation of the total coverage obtained was calculated using Nonpareil
168 (version 3.2; Rodriguez-R and Konstantinidis, 2014). An assembly was obtained using SPAdes
169 (version 3.11.1; Bankevich *et al.*, 2012) and contigs were binned using MaxBin2 (version 2.2.5;
170 Wu, Simmons and Singer, 2016). The quality of the bins obtained was checked using CheckM
171 (version 1.0.11; Parks *et al.*, 2015). Then, genes were predicted on the assembled contigs using
172 prodigal (version 2.6.3). Finally, these sequences were annotated using Interproscan (version
173 5.50.84; Blum *et al.*, 2021) and kofamscan (version 1.3.0; Aramaki *et al.*, 2020). All metabolic
174 pathways regarding the degradation of aromatic compounds, and enzymes described for each
175 step were obtained from MetaCyc (Caspi *et al.*, 2018). For those enzymes described in the
176 InterPro or KEGG database, we extracted their annotation. In order to reconstruct the metabolic
177 pathways, we compared the annotation from the MetaCyc enzymes with that of our predicted
178 proteins and used CytoScape (version 3.9.1) to construct the metabolic network. Also,
179 GhostKOALA annotation server (Kanehisa, Sato and Morishima, 2016) was used to find potential
180 functions related to the metabolic pathways of interest.

181 **2. Evaluating pyrene degrading communities**

182 **2.1 Obtaining a pyrene degrading enrichment culture**

183 A pyrene-degrading bacterial enrichment culture was obtained from IPK soil by five sequential
184 transfers in LMM with 1000 mg.l⁻¹ of pyrene as sole carbon and energy source and 0,05 g.l⁻¹ yeast
185 extract. Biomass transfer:culture medium was 1:10 ratio and each transfer was incubated 7 days
186 at 28°C and 150 rpm. Last bacterial suspension was finally conserved in 40% of glycerol at -80
187 °C. When need it for an assay the enrichment culture was reactivated in LMM supplemented
188 with 1000 mg.l⁻¹ of pyrene and 0,05 g.l⁻¹ yeast extract for 7 days at 28 °C and 150 rpm. The
189 reactivated enrichment culture was named r-EFP

190 **2.1.1 Pyrene degradation kinetics**

191 The pyrene degradation by the enrichment culture was determined in LMM with 100 mg.l⁻¹ of
192 pyrene as a sole carbon and energy source and 0,05 g.l⁻¹ yeast extract. The cultures were
193 inoculated with the reactivated r-EFP and incubated at 28°C and 180 rpm (in triplicates) during
194 21 days of incubation. An abiotic control was incubated in the same conditions.

195 At different incubation times remaining pyrene was measured in the r-EFP cultures. A chemical
196 extraction was carried out three times with ethyl acetate as previously described (Festa,
197 Coppotelli and Morelli, 2013). The ethyl acetate extracts were analyzed by reversed-phase high-
198 pressure liquid chromatography (HPLC) using a Waters chromatograph with a Symmetry Waters
199 C18 column (15 cm 4.6 mm i.d., bead size 5 mm, pore size 100 Å) and a diode-array detector. A
200 linear gradient of 15 mM phosphoric acid in nanopure water solution and methanol (20:80 to
201 5:95 vol.vol⁻¹) over 15 min and a flow rate of 1 ml.min⁻¹ was used.

202 **2.1.2 DNA extraction for 16S rRNA gene metabarcoding and shotgun metagenomic analysis**

203 DNA extraction was carried out using E.Z.N.A.® Soil DNA Kit (Omega Bio-Tek, Inc., Norcross, GA,
204 USA) following manufacturer's instructions. DNA was stored at -20°C until analysis.

205 16S rRNA gene metabarcoding and shotgun metagenomics were performed in an Illumina
206 NovaSeq 6000 sequencing platform at Novogene (<https://en.novogene.com/>). Metabarcoding
207 analysis was carried out using 16S rRNA universal bacterial primers 515F (5' -
208 GTGCCAGCMGCCGCGGTAA-3') and 806R (5'-GGACTACHVGGGTWTCTAAT-3') to amplify V4

209 hypervariable region of 16S rRNA gene. Paired end sequence data analysis was performed as
210 described above in this manuscript for DNA IPK soil sample, using QIIME2 Pipeline (version
211 2021.2).

212 Shotgun metagenomic raw reads were quality trimmed using Trimmomatic (version 0.39). An
213 assembly was obtained using MegaHit (version 1.2.9; Li *et al.*, 2015) and the quality was checked
214 using QUAST (Gurevich *et al.*, 2013). Then, genes were predicted and annotated on the
215 assembled contigs using Prokka (Seemann, 2014). Contigs were binned using Binnacle (version
216 2.2.5; Muralidharan *et al.*, 2021) and the quality of the bins obtained was checked using CheckM
217 (version 1.0.11; Parks *et al.*, 2015). Taxonomical assignments of the metagenome assembled
218 genomes was carried out using PhyloPhlAn (version 3.0.36, Segata *et al.*, 2013) and CAT/BAT
219 (Bastiaan von Meijenfildt *et al.*, 2019). All metabolic pathways regarding the degradation of
220 aromatic compounds, and enzymes described for each step were obtained from MetaCyc (Caspi
221 *et al.*, 2018), KEGG AND AromaDeg database (Duarte *et al.*, 2014). GhostKOALA annotation
222 server (Kanehisa, Sato and Morishima, 2016) was used to find potential functions related to the
223 metabolic pathways of interest.

224 **2.2 Isolation of a pyrene-degrading bacterial strain**

225 A portion of IPK soil was suspended in an Erlenmeyer with physiological solution 0,85% w.v⁻¹.
226 Dilutions were carried out from the suspension and an aliquot was spread on agar plates
227 containing R3A (Vecchioli, Del Panno and Paineira, 1990) supplemented with a 1% agarose layer
228 with 0, 46 mg/ml of pyrene. After 28 days of incubation, bacterial colonies demonstrating a
229 transparent halo around the colony were isolated.

230 **2.1 Biodegradation studies and molecular characterization of S19P6 isolated strain**

231 Cultures with 100 ppm of pyrene as carbon and energy source were inoculated with S19P6 strain
232 and incubated at 28°C and 180 rpm (in triplicates). These cultures were used to quantify pyrene
233 degradation and to evaluate the growth curve of S19P6 using pyrene as sole carbon and energy
234 source at different incubation times. An abiotic control was incubated in the same conditions.

235 For pyrene degradation kinetics cultures were extracted three times with ethyl acetate and
236 analyzed as explained above by HPLC. Growth curve of S19P6 was monitored by counting
237 colony-forming units (CFU) after spreading 0.1 ml of an appropriate dilution on R2A medium and
238 incubating the plates in the dark for 7 days at 28°C (Festa, Coppotelli and Morelli, 2013).

239 In order to establish the identity of the isolated strain, DNA was extracted with boiling method
240 (Festa, Coppotelli and Morelli, 2013) and a PCR amplification using 16S rRNA gene targeted
241 primers was carried out. Resulting amplicon was sequenced by Sanger sequencing service in
242 MacroGen. Nucleotide sequences were compared to those in the National Center for
243 Biotechnology Information GenBank database by using BLAST program. To study the
244 phylogenetic relationship between the 16S rRNA gene sequences from the isolated strain and
245 selected sequences available in GenBank, a distance-based evolutionary tree was constructed
246 using Neighbor-Joining method and p-distance algorithm with the Molecular Evolutionary
247 Genetics Analysis package (MEGA version 4.0). The robustness of the phylogeny was tested by
248 bootstrap analysis with 500 iterations

249 Presence of genes related to key enzymes involved in the degradation pathway of pyrene in
250 S19P6 was evaluated by PCR using two sets of primers targeting *nidA* (as described above) and
251 *pahE* genes. For the latter, the used primers (*pahE4F* and *pahE4R*) were designed by Liang,
252 Huang and Wang, (2019). PCR conditions were carried out as described by those authors.

253 **Data deposition**

254 Metagenomic and amplicon sequence data are available at the NCBI Short-Read Archive under
255 project accession number PRJNA889761 for IPK soil and under project accession number
256 PRJNA906488 for r-EFP enrichment culture datasets. Full length 16S rRNA gene sequence of the
257 isolated strain is available in NCBI GenBank under accession number OP899840.

258 **Results**

259 **1. Soil characterization**

260

261 **1.1 Hydrocarbon quantification and microbial enumeration**

262 Chemical extraction was performed in order to describe PAH content of IPK soil. As it can be
263 seen from the data shown in table 1, IPK soil presented a total PAH hydrocarbon concentration
264 of 412.3 ± 24.8 mg.kg⁻¹ of dry soil. Notably, a predominance of three and four rings PAH was
265 observed with pyrene as major concentration hydrocarbon (Table 1), 119.3 ± 4.3 mg.kg⁻¹ of dry
266 soil, followed by acenaphtylene 79.1 ± 3.3 mg.kg⁻¹ of dry soil.

267 Regarding bacterial enumeration, an estimation of total heterotrophic cultivable bacteria, PAH-
268 degraders and specifically pyrene degrading bacteria, was assessed by culture dependent and
269 culture independent techniques.

270 Heterotrophic cultivable bacteria revealed a concentration of $4.15 \cdot 10^6 \pm 3.08 \cdot 10^5$ CFU. g dry
271 soil⁻¹ while pyrene and PAH degrading bacteria counts were four orders of magnitude lower,
272 $3.26 \cdot 10^2 \pm 1.58 \cdot 10^2$ and $4.09 \cdot 10^2 \pm 4.16 \cdot 10^1$ respectively. Quantification of 16S rRNA gene and
273 PAH-RDH α GN functional gene was performed by a qPCR assay. The obtained value for total
274 bacteria was $4.27 \cdot 10^{10} \pm 2.52 \cdot 10^{10}$ 16S rRNA copies. g dry soil⁻¹ and for the functional gene was
275 $2.16 \cdot 10^8 \pm 1.72 \cdot 10^8$ PAH-RDH α copies. g dry soil⁻¹; the difference between those values was not
276 as high as for the culture dependent approach.

277

278 **1.2 Diversity analysis and prediction of functional potential**

279

280 Taxonomic and diversity analysis was accomplished by 16Sr RNA gene metabarcoding. After
281 proper analysis, a total of 382,461 sequences were obtained for IPK triplicates samples and were
282 grouped in 369 amplicon sequence variants (ASVs). The rarefaction curves (Figure S1) and
283 Good's coverage (value of 1) illustrated that the effort was enough to cover most of the diversity
284 of the samples. As diversity measures, Hill-numbers, ⁰D, ¹D and ²D were 365, 110.77 and 31.21
285 respectively. The decrease observed between ⁰D and ¹D values suggested that IPK soil presents
286 a highly impacted and uneven bacterial community, with low richness and diversity (Table S1).

287 At phylum level a clear dominance of Proteobacteria (53.7% of relative abundance) was
288 observed, followed by Actinobacteria (26.3% of relative abundance) (Figure 1). Within
289 Proteobacteria, Immundisolibacterales was the most represented order in the samples (14.8%
290 of relative abundance) followed by Burkholderiales order (5.0% of relative abundance);
291 Actinobacteria was mainly represented by Microtrichales order (15.6% of relative abundance).

292 In order to assess the functional potential of IPK soil, a predictive functional profile of the
293 community was assessed using PICRUSt2. Mean Weighted Nearest Sequenced Taxon Index
294 (NSTI) value was 0.22 ± 0.01 . When looking to the predicted pathways associated to the
295 degradation of aromatic compounds (Figure 2a), it can be seen that the ortho-cleavage pathway

296 of protocatechuate (P3) was the most significantly ($p < 0,01$ ANOVA-Tukey) represented,
297 followed by the meta- and orto- cleavage of catechol pathways (P4 and P5 respectively) and the
298 superpathway of salicylate degradation (P8). Additionally, a prediction of the contribution of
299 each taxa to the mentioned pathways was carried out. A clear predominance of
300 Alphaproteobacteria, Actinobacteria, Gammaproteobacteria and Acidimicrobiia classes was
301 observed, representing more than the 80% of the contribution to each pathway (Figure.2a.). In
302 particular, Alphaproteobacteria class was the major contributor to ortho-cleavage of
303 protocatechuate (P3). A selection of the most represented KO for each of those pathways were
304 carried out and taxa contribution to those KO was predicted and showed in Figure 2b. Regarding
305 functions related to the ortho-cleavage of Protocatechuate (green bars), a higher proportion of
306 K01607 (EC:4.1.1.44) and K01055 (3.1.1.24) was found in Microtrichales and Hyphomicrobiales
307 orders. Concerning meta-cleavage of catechol (yellow bars): K07104 (1.13.11.2) was mostly
308 found to be related to Hyphomicrobiales order, while K01821 (5.3.2.6) was equally found in
309 Hyphomicrobiales and Burkholderiales. However, Xanthomonadales and Dongiales were major
310 contributors to K02554 (4.2.1.80), indicating that Proteobacteria Phylum (blue bars) was the
311 major contributor to this pathway.

312

313 **1.3 Metagenomic analysis**

314 **1.3.1 Soil taxonomic profile**

315

316 From the sequencing of total DNA IPK soil, approximately 11 million of high quality raw paired
317 end sequences were obtained. Sample coverage in Nonpareil software was analyzed, the result
318 indicated that IPK sample was oversample, meaning full dataset coverage was achieved,
319 inferring the presence of low diversity community in IPK soil (Figure S2).

320 The taxonomic description of the soil is shown in figure 3a. About 97% of the cleaned reads were
321 taxonomically classified and, almost all were assigned to Actinobacteria phylum (99,81%),
322 specifically Streptomycetales order (99,54%). A vast minority of all sequences were assigned to
323 two other phyla (Figure 3a) Proteobacteria (0,11%), represented by Pseudomonadales (0,02%)
324 and Immundisolibacterales order (0.01%), and Firmicutes (0,05%) represented by Bacilliales
325 order (0,03%). After the assembly of reads a total of 5357 contigs were obtained, being 242
326 contigs larger than 10 kbp. Around 6300 genes were predicted in those contigs (Figure 3b),
327 where almost 99% of genes were assigned to *Streptomyces* genus. After a binning procedure a
328 bin of a size of almost 5 Mb (81,6% of completeness and 0,86% of contamination) composed of
329 349 contigs was recovered, it possesses 73% of GC content and was taxonomically assigned to
330 Streptomycetaceae family.

331 **1.3.2 Functional characterization of IPK soil community**

332 *Functional categories related to hydrocarbon degradation*

333 Different approaches were used to search for genes related to aromatic compound degradation
334 in IPK soil metagenome.

335 Based on GhostKOALA annotation, a total of 12 degradation-related genes were found, 10
336 assigned to Actinobacteria phylum, mainly to Streptomycetales order, and 2 to Proteobacteria,
337 in particular to Burkholderiales and Hyphomicrobiales orders (Table 2). Meta-cleavage of
338 protocatechuate was the most represented pathway (with four sequences assigned to this

339 pathway), followed by meta-cleavage of catechol (with three KO sequences assigned to this
340 pathway), all represented by Actinobacteria related sequences. K14584 (related to naphthalene
341 degradation) was assigned to Burkholderiales order and K01055 (related to ortho-cleavage of
342 catechol and protocatechuate degradation) to Hyphomicrobiales.

343 As only a few aromatic compound degradation related genes were found using GhostKOALA
344 approach, a comprehensive analysis of the enzymatic domains present in IPK soil metagenome
345 was carried out using MetaCyc database. As a result, a metabolic network relating the aromatic
346 compound degradation pathways found is shown in Figure 4. All the enzymatic domains
347 associated to those pathways were found to be assigned to Streptomycetales order. In the
348 constructed network, the width of the lines connecting two compounds refers to the percentage
349 of completeness of the pathway involved (referring to what percentage of the enzymatic
350 domains were found in relation to the total), and the color refers to what type of dioxygenase
351 was found in the metagenome to be participating in that pathway.

352 With this approach, almost all genes codifying for enzymes of naphthalene degradation upper
353 pathway were found (78% of completeness, from naphthalene to salicylate). Concerning the
354 initial step of the pathway, the presence of dioxygenase enzyme belonging to Rieske non-heme
355 iron oxygenases family, capable of participating in this and in other several pathways (benzene,
356 p-cumate, biphenyl, phthalate and salicylate degradation) was found (green lines). Also, the
357 domain of a key enzyme responsible for the fifth step of the upper pathway, a trans-o-
358 hydroxybenzylidenepyruvate hydratase-aldolase (EC: 4.1.2.45, K14585) was recovered from the
359 metagenome.

360 Regarding central intermediates, almost the complete set of enzymes (between 80% and 100%)
361 was found for phthalate degradation, salicylate and catechol degradation (meta-cleavage
362 pathway). However, for the latter, the gene coding for the enzyme responsible for the meta
363 cleavage was not found in IPK metagenome (catechol 2,3-dioxygenase). Despite 57% of the
364 enzymes responsible for catechol ortho-cleavage pathway were found, a key enzyme, the
365 intradiol dioxygenase responsible for the cleavage of catechol was present in the metagenome
366 (catechol 1,2-dioxygenase, pink line). A high percentage (between 60 and 79%) of completeness
367 was observed for protocatechuate degradation (meta-cleavage pathway), protocatechuate
368 degradation (para-cleavage pathway) plus 2-hydroxypenta-2,4-dienoate degradation, gentisate
369 degradation and salicylate degradation plus gentisate degradation II.

370 Within protocatechuate metabolic pathways, protocatechuate 2,3-dioxygenase (responsible for
371 the initial step in the para-cleavage pathway) was found. Nor protocatechuate 4,5-dioxygenase
372 neither protocatechuate 3,4-dioxygenase genes were observed in the metagenome.

373 As dioxygenase responsible for the initial attack of pyrene did not have functional gene number
374 and therefore it was not possible to predict its presence in IPK soil metagenome, a PCR was
375 carried out with primers directed to a gene codifying for the PAH dioxygenase involved in pyrene
376 initial attack (*nidA*). A positive result was obtained, indicating that some members of IPK
377 bacterial community could be able to initiate PAH/pyrene degradation (Figure S3).

378

379 **2. Evaluating pyrene degrading communities**

380 **2.1 Pyrene degrading enrichment culture r-EFP**

381 In order to use culture approach as another tool for identification of microbial degraders, a
382 pyrene degrading bacterial enrichment culture, called r-EFP was obtained from IPK soil using

383 pyrene as sole carbon and energy source. The pyrene degradation kinetic determined during 21
384 days of incubation in r-EFP cultures is shown in figure 5. Although pyrene elimination was not
385 observed during the first 7 days of incubation, a significant decrease of the supplemented
386 pyrene was detected after 15 days of incubation (50%). Furthermore, at day 21 of incubation a
387 significantly higher degradation percentage (94.9%) was measured in r-EFP cultures.

388 **2.1.1 r-EFP diversity and functional potential**

389 The bacterial composition of the enriched culture r-EFP was assessed by 16S rRNA
390 metabarcoding. A total of 272080 paired end sequences were obtained after a denoising and
391 filtering procedure of the raw data. Rarefaction curves (Figure S4) demonstrated that the
392 sequencing depth was enough to cover r-EFP culture diversity. A total of 67 amplicon ASV were
393 obtained. Further taxonomic classification of these ASV showed preponderance of two bacterial
394 phyla, Proteobacteria (87.6% of relative abundance) and Bacteroidota (11.9% of relative
395 abundance). As can be seen in Figure 6, at order level Pseudomonadales (49.1%) were
396 predominant, followed by Enterobacterales (23.1%), Sphingobacterales (10.9%),
397 Hyphomicrobiales (8.8%), Xanthomonadales (3.8%) and Burkholderiales (1.8%). Orders between
398 0.1% and 0.05% of relative abundance were Corynebacterales, Micrococcales, Chitinophagales
399 and Sphingomonadales.

400 Insight into bacterial diversity and catabolic potential analysis was accomplished by a shotgun
401 metagenome approach of r-EFP culture. A total of 75250806 paired end raw sequences were
402 obtained and after a quality filtering, trimming and assembly process a total of 90707 contigs
403 were generated, 395 larger than 50000 bp. After annotation a total of 203798 coding DNA
404 sequences (CDSs) were obtained. Taxonomical assignments of the reads (Figure 6) revealed a
405 clear predominance of Pseudomonadales (48.8%), Hyphomicrobiales (5.1%), Sphingobacterales
406 (4.7%), Xanthomonadales (4.7%) Enterobacterales (3.7%), Sphingomonadales (2.6%),
407 Micrococcales (2.4%) and Burkholderiales (1.8%) orders were found in more than 1% of relative
408 abundance. Instead, for Chitinophagales and Corynebacterales orders a minor number of
409 sequences were found, representing 0.4% and 0.3% of relative abundance respectively.

410 After a binning process, eight high-quality metagenome assembled genomes were recovered
411 (completeness \geq 90% and contamination $<$ 5%). Table 3 provides an overview of each MAG
412 features such as completeness (%), contamination (%), size (bp), number of contigs, CDS and GC
413 content (%). GC content ranged from 40,37% to 70,06%. The larger MAG size was 6531845 bp
414 with 100% completeness, whereas the shortest one was 3174027 bp with 95,77% completeness.
415 Taxonomic classification (Table 3) was carried out using a variety of approaches and softwares.
416 Four MAGs were assigned to Proteobacteria (MAG3, MAG5, MAG9, MAG17), three to
417 Bacteroidetes (MAG7, MAG8, MAG14) and one to Actinobacteria (MAG12). Regarding
418 Proteobacteria phylum, MAG3 was classified as Gammaproteobacteria class, MAG5, MAG17 to
419 Xanthomonadaceae family (MAG5, MAG17) and MAG9 to Hyphomicrobiales order. Regarding
420 Bacteroidetes phylum, MAG7 was classified as belonging to Chitinophagia class and MAG8 and
421 MAG14 to Sphingobacteria class, although MAG14 was further classified to Sphingobacteriaceae
422 family. MAG12 was classified to Actinomycetia class and 16S rRNA gene sequence showed a 99%
423 similarity with Micrococcales order.

424 Degradation potential of aromatic compounds in the bacterial culture enrichment was explored
425 by examining the presence of sequences coding for proteins responsible for each step in
426 peripheral and central intermediate metabolic pathways and evaluating which member of the

427 enrichment could contribute to these steps. Overall, approximately 502 annotated genes
428 sequences related to these pathways were found. Integrating GhostKOALA annotation results,
429 KEGG and MetaCyc database a reconstruction of nine aromatic compounds aerobic degradation
430 described pathways was carried out (Figure 7 and 8 and table S2). Counting total sequences
431 annotated for those pathways demonstrated that orto-cleavage of protocatechuate was the
432 most represented pathway, followed by catechol degradation pathway via meta-cleavage and
433 orto-cleavage (Figure 7a). Three main bacterial classes, Alpha-, Beta- and Gammaproteobacteria
434 were found to be the mayor contributors to those pathways (Figure 7b). Among those classes,
435 four bacterial orders contributed to more than the 73% of all sequences, these are
436 Burkholderiales, Hyphomicrobiales, Pseudomonadales and Sphingomonadales (Figure 7c). In
437 figure 8, it is shown that r-EFP could be able to perform three steps of naphthalene degradation
438 to salicylic acid, mainly associated to Sphingomonadales and Burkholderiales orders. 1-Hydroxy-
439 2-naphthoic acid central intermediate of phenanthrene and pyrene degradation could follow
440 two degradation pathways, (1) one leading to salicylic acid, with further conversion to catechol
441 or gentisate, and the (2) other to phthalic acid, leading to the formation of protocatechuate.
442 Regarding the pathway leading to salicylic acid, gene sequences implicated in the conversion of
443 1-hydroxy-2-naphthoic to naphthalene-1,2-diol (K00480, EC:1.14.13.1, coding for a salicylate
444 hydroxylase) were mainly assigned to Hyphomicrobiales and Burkholderiales orders. Salicylic
445 acid conversion to catechol or gentisate and further gentisate degradation functions were
446 mostly found in Burkholderiales and Hyphomicrobiales affiliated gene sequences.
447 For catechol degradation through meta- and orto- cleavage annotated sequences included more
448 bacterial orders than all the other analyzed pathways, being Burkholderiales, Hyphomicrobiales,
449 and Pseudomonadales orders the predominant ones for both cleavage mechanisms and
450 additionally Sphingomonadales order for meta-cleavage. Most represented functions involved
451 in catechol degradation though meta-cleavage were K01821 and K02554, coding for a 2-
452 hydroxyruconate tautomerase (EC:5.3.2.6) and 2-oxopent-4-enoate hydratase (EC:4.2.1.80)
453 respectively, and though orto-cleavage was K01055 a beta-ketoadipate enol-lactone hydrolase
454 (EC:3.1.1.24). Concerning the metabolism of 1-hydroxy-2-naphthoic acid to phthalic acid, no
455 genes were found in r-EFP metagenome. However, genes involved in further degradation of
456 phthalic acid to protocatechuate and the latter to TCA intermediates were annotated.
457 Conversion to protocatechuate was mainly detected from Burkholderiales, Sphingomonadales
458 and Hyphomicrobiales orders. Genes involved in downstream conversion for meta- and orto-
459 cleavage protocatechuate were annotated and mainly found to be affiliated to
460 Hyphomicrobiales and Burkholderiales for the first pathway and also Pseudomonadales and
461 Bacillales for the latter. Most represented genes in these pathways were K10218 (coding for a
462 4-carboxy-4-hydroxy-2-oxoadipate aldolase, E.C: 4.1.3.17) and K01607 (coding for 4-
463 carboxyruconolactone decarboxylase, E.C: 4.1.1.44).
464 In order to assess the presence of dioxygenases coding genes, AromaDeg database was used
465 and as a result, different cluster of oxygenase families, correlating with the native substrates
466 oxidized by their members, were obtained. Rieske non heme iron oxygenases and extradiol
467 dioxygenases were found in r-EFP metagenome (Table 4), being the two most represented the
468 vicinal oxygen chelate superfamily and the LigB extradiol. Concerning Rieske non heme iron
469 oxygenases, (sequences of the biphenyl/naphthalene, phthalate, salicylate and benzoate family
470 were found to be assigned predominately to Alpha- and Beta-proteobacteria. Referring to
471 extradiol dioxygenases, sequences were mostly assigned to Alpha- and Gammaproteobacteria.

472 Functional annotation was carried out on each MAGs and the presence of genes related to
473 aromatic compound degradation was evaluated. Table S3 shows that all MAGs possess genes
474 involved in at least one of the analyzed pathways, however no genes coding for enzymes related
475 to gentisate degradation were found. Additionally, PCR gene amplification was carried out with
476 two set of primers, *nidA* and PAH-RHD α . For both sets, positive product amplification of the
477 expected size was observed (Figure S5).

478

479 **2.2 Isolation and characterization of pyrene-degrading bacterial strain**

480 In order to isolate PAH degrading strains from IPK soil, a bacterial count on a pyrene-R3A plate
481 was carried out and screened after 28 days to find a solubilizing/degradation halo zone around
482 colonies. This halo was found on a small, bright and yellow colony. S19P6 strain was isolated.

483 Pyrene elimination by S19P6 strain was measured in liquid mineral medium culture
484 supplemented with 0,05 g.l⁻¹ yeast extract and 100 ppm of pyrene. Figure 9a shows that after
485 15 days of incubation S19P6 strain was able to eliminate more than 85% of the supplemented
486 pyrene as sole carbon and energy source and complete pyrene elimination was observed after
487 28 days of incubation. No pyrene elimination was observed in the abiotic control. In addition,
488 growth curve of strain S19P6 was performed during pyrene degradation (Figure 9b). An
489 exponential phase was observed during the first 7 days of incubation reaching three orders of
490 magnitude higher than the beginning of the assay, followed by a stationary phase at the end of
491 the incubation period.

492

493 **2.2.1 Taxonomic and functional potential characterization of the isolated strain S19P6**

494 16S rRNA gene sequence analysis of strain S19P6 was carried out, indicating that strain had 99%
495 similarity to sequence assigned to *Mycolicibacterium* genus and was phylogenetically related to
496 members of that genus (Figure S6). Furthermore, 16S rRNA gene sequence was compared with
497 ASV sequences from metabarcoding analysis of IPK soil and r-EFP culture enrichment and with
498 16S rRNA sequences annotated in r-EFP shotgun metagenome. As a result, 100% similarity with
499 an ASV sequence from IPK soil and culture enrichment r-EFP (Table 5) was found.

500 To assess pyrene degradation potential of S19P6 strain a PCR was carried out with two sets of
501 primers, one targeting pyrene dioxygenase gene *nidA* and the second one targeting the *pahE*
502 codifying gene. Positive amplification (product of the expected size) was found in S19P6 for both
503 primers sets (Table 5).

504 **Discussion**

505 The impact of long-term PAH contamination on IPK soil bacterial community was assessed
506 through culture-dependent and independent complementary approaches, ranging from
507 enrichment culture and pure culture studies focusing on its metabolic potential, to community-
508 scale research trying to determine the link between the diversity and functional capabilities of
509 PAH-degraders. To identify key functions and microorganisms involved in the degradation of
510 environmental pollutants is essential to understand the ecological mechanisms of
511 environmental recovery (Gomes *et al.*, 2010; Gannes and Hickey, 2017).

512 The introduction of xenobiotic compounds generates selective loads to the microbial
513 communities in develop mechanism for degradation of such compounds (Gosai *et al.*, 2018). In

514 biodegradation studies, the term adaptation refers to a phenomenon that can occur both in
515 mixed microbial communities and in individual microbial lineages. The mechanisms that would
516 allow adaptation, proposed by van der Meer *et al.*, (1992), are: the induction and / or repression
517 of specific enzymes and / or the enrichment of the microbial population capable of transforming
518 the compound of interest. It was demonstrated by molecular and culture approaches that IPK
519 bacterial community is a highly impacted community. By 16S rRNA metabarcoding analysis a low
520 diversity community was observed with two predominant phyla Proteobacteria and
521 Actinobacteria (Figure 1 and Table S1). Previous studies evaluating diversity of PAH
522 contaminated soils also reported the predominance of the same phylum found in this study
523 (Haleyur *et al.*, 2019; Lu *et al.*, 2019); hydrocarbons tend to suppress certain sensitive groups to
524 select primarily for subgroups of the Actinobacteria and Proteobacteria (Stefani *et al.*, 2015).
525 Regarding predominant orders, Immundisolibacterales (14,80% of relative abundance) and
526 Microtrichales (15.30% of relative abundance) were found (Figure 1). Corteselli, Aitken and
527 Singleton, (2017) reported that a member of Immundisolibacterales order was able to grow on
528 pyrene, phenanthrene, anthracene, benz[a]anthracene and fluorene as a sole carbon and
529 energy source under mesophilic temperature, neutral pH and low salinity conditions. In contrast,
530 no link to PAH degradation was found in literature for Microtrichales order. Shotgun
531 metagenomics results showed that IPK community present a clear predominance of
532 Streptomycetales order (Figure 3a) with almost all annotated genes assigned to this order
533 (Figure 3b). In accordance with the present results, previous studies have reported that
534 Streptomycetales order is widely found in hydrocarbon polluted, with a substantial role in soil
535 ecology and being encountered to be the most abundant isolated genus (Anandan,
536 Dharumadurai and Manogaran, 2016; Baoune *et al.*, 2019).

537 It can therefore be notice that taxonomic classification carried out with both molecular
538 approaches was divergent (Figure 1 and 3a). A possible explanation could be the difference in
539 16S rRNA gene copy number and genome size of the members of the community members.
540 Members of Proteobacteria are usually considered to be fast-growing bacteria (r-strategists)
541 with high 16S rRNA gene copy number, whereas Actinobacteria are mostly slow-growing
542 organisms (K-strategists) with low 16S rRNA gene copy number (Klappenbach, Dunbar and
543 Schmidt, 2000; Brzeszcz *et al.*, 2016; Pereira-Flores, Glöckner and Fernandez-Guerra, 2019).
544 Abundance estimation based on the 16S rRNA sequence counts tend to underestimate the
545 abundance of taxa with low 16S rRNA copy numbers and to overestimate taxa with high 16S
546 rRNA copy numbers (Větrovský and Baldrian, 2013). Microorganisms related to
547 Streptomycetales order have linearized chromosomes commonly over 8 Mbp with a high G+C
548 content (Law *et al.*, 2019), representing a challenge for PCR amplification(Baldrian, 2019).
549 Besides, a larger genome has greater possibilities of being sampled than a smaller genome, even
550 if they occur in equal abundances, as it will produce more fragments per genome (Angly *et al.*,
551 2009). A mayor influence in community analysis is the abundance of the members of the
552 community but also their genome sizes and 16S rRNA gene copy numbers of the abundant
553 microorganisms are important driving factors of the analysis. Evaluation using only 16S rRNA
554 gene fragment abundances should be done if community members share characteristics close
555 to the average genome size and 16S rRNA gene copy number (Frank and Sørensen, 2011). Other
556 potential source of variation in microbial community profiling could be the DNA extraction
557 approaches and the 16S rRNA gene variable region primers used for sequencing (Han *et al.*,
558 2020). In this study not only DNA extractions were carried out using the same commercial kit,

559 but also other authors could detect Streptomycetaceae family using the same pair of primers
560 employed for 16S rRNA gene sequencing (Ho, Di Lonardo and Bodelier, 2017; Kuang *et al.*, 2018;
561 BenIsrael *et al.*, 2020; Shimoi *et al.*, 2020).

562 Considering the above-mentioned limitation of both techniques, to further characterize IPK soil
563 community two culture approaches were performed, a culture enrichment and a strain isolation.
564 Microbial enrichment are natural microbiomes cultivated *ex situ* in a defined growth medium
565 under well-controlled conditions that, by means of sequential transfers with a selective
566 pressure, yield a specialized microbial consortium, where most niches are constructed by the
567 microorganisms in the original community (Jiménez, Mares and Salles, 2018; Estrela, Sánchez
568 and Rebolleda-Gómez, 2021). Because functional genes belonging to members of IPK bacterial
569 community other than Streptomycetales order were disguised by the latter and taking
570 advantage of the increased accessibility and tractability of these cultures compared to natural
571 communities (Blasche *et al.*, 2017), the potential functions related to aromatic compound
572 degradation of IPK soil community were also assessed by an enrichment culture and by an
573 isolated strain S19P6. Previously, Gomes *et al.*, (2010) took advantage of this kind of cultures to
574 study, by fingerprint analyses, potential functions present in an enriched community and
575 undetected in the source environment.

576 A pyrene degrading enrichment culture r-EFP was obtained from soil. After confirming pyrene
577 elimination (Figure 5), a 16SrRNA gene metabarcoding analysis, and a metagenomic approach
578 was carried out to assess diversity and functional potential of r-EFP. Regarding r-EFP microbial
579 taxa, contrary to IPK soil results obtained in this research, bacterial phylum and orders agreed
580 across sequencing methodologies (Figure 6), as reported by other authors (Jovel *et al.*, 2016;
581 Brumfield *et al.*, 2020). Both, 16SrRNA gene metabarcoding and shotgun metagenomic
582 approaches, indicated a clear predominance of Proteobacteria Phyla, being Pseudomonadales
583 order the one present in almost 50% of relative abundance, a vastly known PAH degrading
584 related order (Ma, Xu and Jia, 2013; Ghosh, Jasmine and Mukherji, 2014; Swati, Ghosh and
585 Thakur, 2019). Although bacterial orders were common in both approaches differences in
586 relative abundances were observed. Members of the other bacterial orders found in r-EFP
587 previously related to pyrene degradation by other authors were Sphingomonadales (Hesham *et al.*,
588 2014; Vila, Tauler and Grifoll, 2015), Burkholderiales (Al-Thukair and Malik, 2016),
589 Corynebacteriales (Wu *et al.*, 2019), Enterobacteriales (Gupta *et al.*, 2020), Micrococcales
590 (Gauthier *et al.*, 2003), Sphingobacteriales (Zhao *et al.*, 2013), Xanthomonadales (Mangwani *et al.*,
591 2014) and Hyphomicrobiales (Ortega-González *et al.*, 2015). With regards to *Chitinophagales*
592 orders, no isolated pyrene degrading microorganism related to this order was found. As
593 reported by Kwon, Kwon and Kim, (2019) only thirteen isolates from the phylum Bacteroidetes
594 were confirmed to degrade hydrocarbon, but none from the order mentioned above. When
595 focusing on potential functions related to aromatic compound degradation, members of r-EFP
596 enrichment culture held genes codifying for enzymes responsible for the degradation of
597 naphthalene, 1-hydroxy-2-naphthoic acid, salicylic acid, gentisate, phthalic acid, catechol and
598 protocatechuate (Figure 7 and 8). Predominant pathway in IPK soil predicted by PICRUSt2 was
599 concordant with the results obtained with r-EFP metagenome. Functional profiling with both
600 approaches demonstrated that protocatechuate degradation via ortho-cleavage was the most
601 represented pathway not only in r-EFP metagenome (Figure 7) but also in IPK soil PICRUSt2
602 prediction (Figure 2a). By both approaches K01607 (Figure 2b and 8) (4-carboxymuconolactone
603 decarboxylase) was found to be represented by members of Proteobacteria and Actinobacteria,

604 being Hyphomicrobiales order one of the mayor contributors. No genes related to this function
605 was found in IPK soil shotgun metagenome. Regardless not having found genes responsible for
606 the conversion of 1-hydroxy-2-naphthoic acid to phtalic acid, sequences related to the
607 degradation of the latter compound were found (Figure 8) so, degradation could occur also via
608 phthalic acid. Sequences related to dioxygenase coding genes belong primarily to Alpha-, Beta-
609 and Gammaproteobacteria classes (Figure 8). The presence of dioxygenases coding sequences
610 was demonstrated by shotgun metagenome analysis (Table 4) but also by PCR using specific
611 primer sets designed to amplify *nidA* gene for *Mycobacterium* genus (Debruyne, Chewning and
612 Saylor, 2007) and genes codifying for the alpha subunit of the PAH-ring hydroxylating
613 dioxygenases (Cébron *et al.*, 2008) (Figure S5 and S6). The positive amplification in *nidA* PCR
614 could be due to, despite being in very low relative abundance, Corynebacteriales order is present
615 in r-EFP (Figure 6) and so is a genus of this order, *Mycobacterium*, as demonstrated by the
616 similarity found in S19P6 isolated strain 16S rRNA gene with an ASV found in r-EFP (Table 5).
617 Furthermore, *nidA* gene could be in the genome of a gram-negative bacteria in r-EFP as
618 demonstrated by Klankeo *et al.*, (2007) who found the presence of *nidA* gene in a pyrene
619 degrading *Pseudoxanthomonas* sp. isolated from soil. As for PAH-RHD α PCR positive result, used
620 primer set was designed to amplify genes common to Gram negative PAH degrader (Cébron *et*
621 *al.*, 2008) including sequences belonging to Pseudomonadales, Sphingomonadales and
622 Burkholderiales order, all of them found in r-EFP (Figure 6). Binning of the r-EFP metagenome
623 reveled eight high quality metagenome assembled genomes (completeness \geq 90% and
624 contamination \leq 5%) (Table 3). Interestingly, none of them were taxonomically assigned to
625 Pseudomonadales predominant order but to orders present in minor abundance in the
626 enrichment culture; this result is contrary to a previous study which reconstructed high quality
627 MAGs from the most abundant OTUs identified in a biphenyl-degrading soil bacterial consortium
628 (Daniel Garrido-Sanz, Javier Manzano, Marta Martín, 2018). Sequences related to the analyzed
629 aromatic compound degradation annotated pathways were found in the recovered MAGs. Taxa
630 contribution to the initial pyrene degradation steps was not possible to determine as enzymes
631 involved have no functional gene numbers (Wang *et al.*, 2016). Even more, the observed
632 functional redundancy did not allow to assert which microorganisms present in r-EFP could
633 perform and ecological function or not (Rodríguez *et al.*, 2016), so isolating members of r-EFP
634 could be of great help to understand the ecological niche of each in the degradation of pyrene.
635 Furthermore, a pyrene degrading strain S19P6, identified as a member of *Mycolicibacterium*
636 genus (Table 5), was isolated from IPK soil demonstrating ability to grow and degrade pyrene as
637 sole carbon and energy source after 28 days of incubation (Figure 9a and b). Also, the presence
638 of *nidA* and *pahE* gene found by PCR in S19P6 strain further confirm the pyrene degradation
639 capacity (Table 5). *pahE* gene product catalyzes a significantly important step (the fifth) of the
640 aerobic PAH degradation pathway; its presence could be indicative that bacteria could grow
641 solely on PAH, not only add oxygen to this compound (Liang, Huang and Wang, 2019). *pahE* gene
642 has recently been postulated as the best functional marker to indicate PAH degradation (Liang
643 *et al.*, 2022). Among many bacterial isolates capable of degrading pyrene, the genus
644 *Mycobacterium*, which can utilize pyrene as a sole carbon and energy source, has been most
645 widely studied (Kim *et al.*, 2018; Yang *et al.*, 2021). Interestingly, neither the S19P6 isolated
646 strain nor the members of r-EFP were present between the most abundant taxa found by 16S
647 rRNA metabarcoding (Figure 1) nor shotgun metagenomics (Figure 3) in IPK soil. This result
648 further supports the existing gap between culture dependent and independent approaches

649 previously described by other researchers. Several reports have shown that not only the most
650 abundant microorganisms detected by culture independent techniques were not isolated by
651 culture approaches (Stefani *et al.*, 2015) but also that microorganisms captured by culturing
652 method are present in low abundance or absent in the sequencing analysis (Lee *et al.*, 2016;
653 Wang *et al.*, 2017). The enrichment and pure culture isolation strategies tend to favor the
654 growth of members with r-strategies (Alain and Querellou, 2009) and also the “conditional rare
655 taxa”. This kind of rare taxa could respond to natural or laboratory selective pressures during
656 culture enrichments becoming abundant by the changing conditions (Pascoal, Magalhães and
657 Costa, 2020). It was demonstrated that culture-based methodologies could complement and
658 increase knowledge obtained from independent-culture ones (Oita *et al.*, 2021) when
659 characterizing soil microbiomes.

660 **Conclusion**

661 Previous bioremediation strategies on IPK soil as allochthons bioaugmentation and surfactant-
662 enhanced bioremediation were assayed and a limitation in the degradation of contaminants was
663 observed (Festa *et al.*, 2016; Cecotti *et al.*, 2018). However, research carried out in this work
664 demonstrated that members of the bacterial community of the analyzed chronically PAH
665 contaminated soil have the genetic potential related to aromatic compound degradation. For
666 the abovementioned reason, other bioremediation strategies to trigger this non-active potential
667 need to be tested, such as the addition of easily accessible carbon source (biostimulation) or/and
668 autochthonous bioaugmentation with the obtained cultures.

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674 **Declaration of Competing Interest**

675 The authors declare that they have no known competing financial interests or personal
676 relationships that could have appeared to influence the work reported in this paper

677 **Figure legends**

678 **Figure 1.** Bacterial composition (relative abundance) at phylum and order level (>3% of relative
679 abundance) in IPK soil.

680 **Figure 2a.** Bacterial classes contribution to metabolic pathways related to the degradation of
681 aromatic compounds according to PICRUSt2 prediction. Results are shown as a mean of triplicate
682 independent samples. P1:Catechol degradation to β -ketoadipate; P2:Protocatechuate
683 degradation (meta-cleavage pathway); P3:Protocatechuate degradation (ortho-cleavage
684 pathway); P4:Catechol degradation (meta-cleavage pathway); P5:Catechol degradation (ortho-
685 cleavage pathway); P6:Catechol degradation to 2-oxopent-4-enoate II; P7:Catechol degradation
686 (meta-cleavage pathway); P9:Superpathway of salicylate degradation.

687 **Figure 2b.** Sankey diagram showing the absolute contribution of chosen PICRUSt2 predicted KO
688 by bacterial orders (orders with less contribution than 10% were grouped as “Other orders”).
689 Vertical nodes are proportional to size. Green node: functions related to Protocatechuate
690 degradation (ortho-cleavage pathway); yellow nodes: functions related to Catechol degradation
691 (meta-cleavage pathway); red nodes: orders belonging to Actinobacteria Phylum; blue nodes:
692 orders belonging to Proteobacteria Phylum; grey node: other orders.

693 **Figure 3a.** Relative abundance (%) of bacterial orders in IPK soil assessed by shotgun
694 metagenomic (orders with less than 0.01% of relative abundance were grouped as “Other
695 orders”).

696 **Figure 3b.** Barplot showing the number of genes annotated in IPK soil metagenome and its
697 taxonomic assignment.

698 **Figure 4.** Metabolic network plot showing the aromatic compounds degradation related
699 functions found in IPK shotgun metagenome. Line width: completeness of the pathway involved;
700 The line color indicates type of dioxygenase found in the metagenome to be participating in that
701 pathway (pink line: catechol 1,2-dioxygenase; light blue line: protocatechuate 2,3-dioxygenase;
702 green line: Rieske non-heme iron oxygenases family dioxygenase). Key pathway intermediates
703 are shown in red.

704 **Figure 5.** Pyrene concentration (mg.l^{-1}) in the r-EFP culture growing in LMM with pyrene as a
705 sole carbon and energy source during 21 days incubation. Results are means of triplicate
706 independent experiments. Bars represent standard deviations.

707 **Figure 6.** Bacterial composition (relative abundance) at order level revealed by 16S rRNA gene
708 sequencing ($>0.05\%$ of relative abundance) and shotgun metagenome ($>0.2\%$ of relative
709 abundance) analysis in r-EFP.

710 **Figure 7.** Heatmaps showing a) the total count of sequences related to aromatic compound
711 degradation pathways in the r-EFP shotgun metagenome and b) bacterial classes and c) orders
712 contributing to those pathways

713 **Figure 8.** Heatmap showing bacterial orders contribution to KEGG orthologs involved in aromatic
714 compound degradation found in r-EFP shotgun metagenome using GhostKOALA and AromaDeg
715 database.

716 **Figure 9.** a) Pyrene concentration (mg.l^{-1}) in the S19P6 cultures growing in LMM with pyrene as
717 a sole carbon and energy source during 28 days of incubation. b) Growth of S19P6 strain on
718 pyrene as an only carbon and energy source in LMM during 28 days of incubation. Results are
719 means of triplicate independent experiments. Bars represent standard deviations.

720 **Figure S1.** Rarefaction curves of IPK sequencing data indicating the observed features (y-axis) as
721 a function of sequencing depth (x-axis).

722 **Figure S2.** Estimated average coverage of the IPK metagenomic dataset using Nonpareil
723 Software.

724 **Figure S3.** Agarose gel electrophoresis of PCR products of *nidA* gene of IPK soil sample.

725 **Figure S4.** Rarefaction curves of r-EFP sequencing data indicating the observed features (y-axis)
726 as a function of sequencing depth (x-axis).

727 **Figure S5.** Agarose gel electrophoresis of PCR products of *nidA* gene in r-EFP (r-E).

728 **Figure S6.** Neighbor-joining tree based on a distance matrix analysis of 16S rDNA partial
729 sequences of S19P6 isolated strain and selected sequences available in GenBank was
730 constructed using the MEGA software version 4. The numbers at each node correspond to the
731 bootstrap per cent values. The space bar indicates 0.01 sequence variation.

732 **Supplementary material**

733 **References**

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