

Alkane Biodegradation Genes from Chronically Polluted Subantarctic Coastal Sediments and Their Shifts in Response to Oil Exposure

Lilian M. Guibert, Claudia L. Loviso, Magalí S. Marcos, Marta G. Commendatore, Hebe M. Dionisi & Mariana Lozada

Microbial Ecology

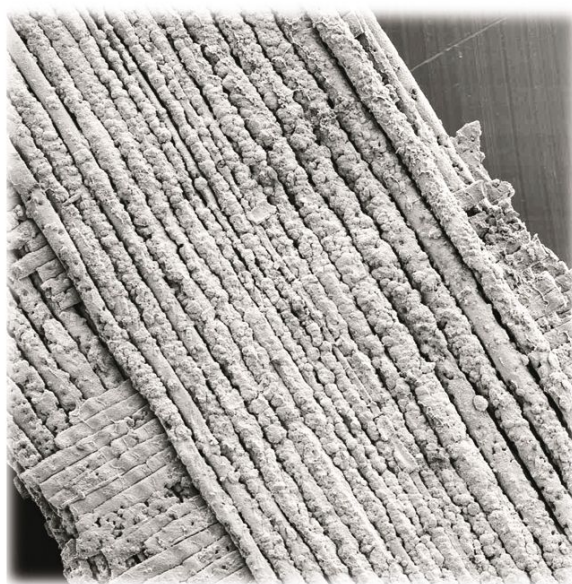
ISSN 0095-3628

Microb Ecol

DOI 10.1007/s00248-012-0051-9

Microbial Ecology

Volume 63 Number 2
February 2012



 Springer

63(2) 239–470 • 248 ISSN 0095-3628

Available
online
www.springerlink.com

 Springer

Your article is protected by copyright and all rights are held exclusively by Springer Science+Business Media, LLC. This e-offprint is for personal use only and shall not be self-archived in electronic repositories. If you wish to self-archive your work, please use the accepted author's version for posting to your own website or your institution's repository. You may further deposit the accepted author's version on a funder's repository at a funder's request, provided it is not made publicly available until 12 months after publication.

Alkane Biodegradation Genes from Chronically Polluted Subantarctic Coastal Sediments and Their Shifts in Response to Oil Exposure

Lilian M. Guibert · Claudia L. Loviso ·
Magalí S. Marcos · Marta G. Commendatore ·
Hebe M. Dionisi · Mariana Lozada

Received: 3 February 2012 / Accepted: 27 March 2012
© Springer Science+Business Media, LLC 2012

Abstract Although sediments are the natural hydrocarbon sink in the marine environment, the ecology of hydrocarbon-degrading bacteria in sediments is poorly understood, especially in cold regions. We studied the diversity of alkane-degrading bacterial populations and their response to oil exposure in sediments of a chronically polluted Subantarctic coastal environment, by analyzing alkane monooxygenase (*alkB*) gene libraries. Sequences from the sediment clone libraries were affiliated with genes described in *Proteobacteria* and *Actinobacteria*, with 67 % amino acid identity in average to sequences from isolated microorganisms. The majority of the sequences were most closely related to uncultured microorganisms from cold marine sediments or soils from high latitude regions, highlighting the role of temperature in the structuring of this bacterial guild. The distribution of *alkB* sequences among samples of different sites and years, and selection after experimental oil exposure allowed us to identify ecologically relevant *alkB* genes in Subantarctic sediments, which could be used as biomarkers for alkane biodegradation in this environment. 16 S rRNA amplicon pyrosequencing indicated the abundance of several genera for which no *alkB* genes have yet been described (*Oleispira*, *Thalassospira*) or that have not been previously associated with oil biodegradation (*Spongiibacter*—formerly *Melitea*—, *Maribius*, *Robiginitomaculum*, *Bizionia* and *Gillisia*). These

genera constitute candidates for future work involving identification of hydrocarbon biodegradation pathway genes.

Introduction

Petroleum crude oil and its derivatives are complex chemical mixtures, which can contain up to thousands of molecules, typically 90 % hydrocarbons [1]. In the marine environment, coastal sediments are the natural sink of hydrocarbons [2]. In cold regions, in particular, stranded oil can persist in the shoreline, producing short- and long-term undesirable ecological effects [3–6]. Increasing pressure on high latitude ecosystems due to oil exploration and exploitation as well as oil-based activities requires the development of knowledge-based bioremediation tools, specific for cold marine environments [7]. The first step to achieve this goal is the identification of ecologically relevant oil-degrading bacteria and their biodegradative capabilities. As aerobic biodegradation processes play a key role in the sediment system [8], increasing our knowledge on aerobic hydrocarbon-degrading bacterial communities from cold marine sediments is fundamental.

Aerobic biodegradation of oil is a complex process involving many different microorganisms [9]. In sediments, the intrinsic biodiversity and heterogeneity of indigenous microbial communities adds a level of complexity to the identification of hydrocarbon-degrading bacterial populations and their biodegradation pathways. This hampers our understanding of the ecological mechanisms underlying the biodegradation process. Important advances have been made in this direction in temperate [10–14] and tropical [15–22] sediments. In contrast, oil-degrading bacterial

Electronic supplementary material The online version of this article (doi:10.1007/s00248-012-0051-9) contains supplementary material, which is available to authorized users.

L. M. Guibert · C. L. Loviso · M. S. Marcos ·
M. G. Commendatore · H. M. Dionisi · M. Lozada (✉)
Centro Nacional Patagónico (CENPAT - CONICET),
Blvd. Brown 2915,
U9120ACD, Puerto Madryn, Chubut, Argentina
e-mail: lozada@cenpat.edu.ar

populations from cold sediments have only started to be described [23–26].

Since saturated hydrocarbons (*n*-alkanes, branched alkanes and cycloalkanes) constitute the largest fraction of crude oil by mass, their biodegradation is a major process during hydrocarbon removal from the environment [27]. Furthermore, there is evidence that alkanes could serve as cosubstrates for the cometabolic removal of more recalcitrant oil compounds like resins and asphaltenes [28, 29], as well as of other pollutants such as methyl tert-butyl ethers from gasoline [30]. Several microorganisms are known to use alkanes as carbon sources. In the marine environment, obligate hydrocarbonoclastic bacteria (OHCB), namely, *Alcanivorax* [31], *Thalassolituus* [32], *Oleiphilus* [33] and *Oleispira* [34], have been proposed as being ecologically important for the aerobic removal of alkanes [35]. Alkane-activating enzymes belong to different families, being integral membrane non-heme iron monooxygenases (AlkB) the best characterized [36–38]. The *alkB* genes have been used as functional biomarkers for the characterization of aerobic alkane-degrading bacterial populations in environmental samples [24, 25, 39–41] and in bioremediation experiments [42–44].

Ushuaia Bay, located within the Beagle Channel in the southernmost tip of South America, has been chronically exposed to various pollutants, among which anthropogenic hydrocarbons are of major relevance [26, 45–49]. Our previous work focused on microbiological and ecological aspects of polycyclic aromatic hydrocarbon biodegradation in sediments of this Subantarctic environment [26, 47, 50, 51]. The goal of this study was to identify key bacterial populations associated with aerobic alkane biodegradation processes in these Subantarctic coastal sediments. To achieve this goal, we analysed *alkB* genes from sediment samples obtained in various years, as well as shifts in this gene produced as a result of experimental oil exposure. To analyse these experiments from a phylogenetic perspective, we also performed 16 S rRNA gene amplicon pyrosequencing.

Methods

Sample Collection and Processing

Surficial sediments (0–3 cm) were collected using acrylic cores with an inner diameter of 4.4 cm. For intertidal sediment samples, sampling was performed along the low tide line at seven to ten random points in each sampling location, and the collected sediments were mixed thoroughly. The subtidal sample was collected at 11 m depth by scuba diving using the same acrylic cores, and in this case, the 0–5-cm fraction was used. All samples were stored at 4°C during transport to the laboratory and further stored at –80°C for

molecular analyses and at –20°C for hydrocarbon content analysis.

Construction of Sediment-in-Seawater Slurries

The sample used for the experiment (OR08) was retrieved on 18 December 2008, which corresponds to late spring. The sediment temperature at the time of sampling was 8.2°C. Sediment-in-seawater slurries were built by mixing approximately one part of fresh sediment with two parts of 0.45-μm filtered natural seawater, as follows. A sediment:water mix was prepared, decanted and further divided into three 500-ml flasks containing 40 g of sediment and 80 ml of seawater, with the following additions: (1) 0.46-ml light petroleum crude oil from Austral basin (Comodoro Rivadavia, Argentina) in 115 ml of sediment:seawater mix (0.4 % v/v, expOR08-O), (2) 0.4 % (v/v) light petroleum crude oil with the addition of 5-mM ammonium chloride and 0.25-mM sodium phosphate (expOR08-ON) and (3) a control with neither crude oil nor nutrient addition (expOR08-c). The three systems were held at 15°C with constant agitation (150 rpm) for 20 days in the dark. Samples were obtained for molecular and hydrocarbon content analyses and stored as mentioned previously.

Hydrocarbon Analysis

The procedure was followed according to [49, 52]. Briefly, samples were extracted with dichloromethane in a Soxhlet device. The extract was treated with activated copper and evaporated at 40°C with highly pure N₂ until a volume of 0.2 ml was reached. The extract was loaded into an activated alumina column in order to recover the aliphatic and aromatic fractions. The aliphatic fraction was then subjected to high-resolution gas chromatography on a Thermo Trace gas chromatograph with a TriPlus Autosampler (Thermo Electron Corporation, Waltham, MA, USA), equipped with a J & W DB5 fused silica column (30×0.25×25), split/splitless capillary injection system, and a flame ionization detector (FID). For *n*-alkanes from *n*C20 to *n*C28 range, recovery assays of spiked pristine samples resulted in 95±12 %. The percent of relative deviation (RDP) for individual aliphatic hydrocarbons varied from 0.4 % to 9 %, and the detection limit (LOD) was between 5 and 10 ng g^{–1}. The identification of resolved aliphatic hydrocarbons was achieved by comparing retention times with the corresponding standards (Chem Service Inc, West Chester, PA, USA). Total resolved aliphatics (TRA) and unresolved complex mixture (UCM) were calculated using the mean response factors of *n*-alkanes.

Individual *n*-alkane concentrations from *n*-C10 to *n*-C35, Pristane (Pr) and Phytane (Ph) isoprenoid levels, total resolved *n*-alkanes ($\sum nalk$), TRA, UCM and total aliphatic

hydrocarbons (TAH=TRA+UCM) were calculated for each sample. The UCM involves cycloalkanes, branched alkanes and other compounds unresolved by the capillary column which show as a hump shape below the resolved compounds. Hydrocarbon concentration values were expressed as $\mu\text{g/g}$ dry weight, after correcting for sediment moisture content.

Evaluation indices selected for this study were *n*-C17/Pristane (*n*C17/Pr), *n*-C18/Phytane (*n*C18/Ph) and TRA/UCM, in order to evaluate the relative biodegradation of *n*-alkanes [52, 53]. In addition, the ratio between low molecular weight and high molecular weight hydrocarbons ($\Sigma\text{LMW}/\Sigma\text{HMW}$, the sum of alkanes $<n\text{C}20$ / the sum of alkanes $\geq n\text{C}21$) was calculated as an estimation of the relative biodegradation of lighter compounds in the experiments.

DNA Extraction

DNA was purified from 0.5 to 0.8 g wet weight sediment using the FastDNA[®] SPIN kit for soil (MP Biochemicals, Solon, OH, USA), according to the manufacturer's instructions with the following modifications: Samples were homogenized three times for 50 s at approximately 5,000 rpm (speed at high setting) with 1-min intervals using a mini bead-beater Biospec (Bartlesville, OK, USA). Two DNA extractions per sample were combined before further analysis.

PCR Amplification of *alkB* Genes

The *alkB* gene fragments were amplified using the *AlkB*484F (5' GGKCA YTTCTWCRTY GARCA 3')/*AlkB*824R (5' CCGTAGTGYTCRABRTARTT 3') primer set [54]. PCR reactions were performed in 25- μl volume containing 1 \times PCR buffer, 1.5-mM MgCl_2 , 0.2 mM of each dNTP, 0.5 μM of each primer and 1 U of T-PLUS DNA polymerase (Inbio-Highway, Tandil, Argentina). Template concentration was optimized for each sample (1–5 ng/reaction). Cycling conditions involved an initial 5-min denaturing step at 94°C, followed by 40 cycles of 30 s at 94°C, 30 s at 45°C and 30 s at 72°C, and a final 15-min elongation step at 72°C. Four separate reactions were run for each sample, pooled and purified from 1.5 % (w/v) agarose gels using Wizard SV Gel and PCR clean-up System kit (Promega, Madison, WI, USA).

Clone Library Construction, Screening and Sequencing

Prior to the cloning step, 3' A-overhangs were added to the purified *alkB* gene fragments by incubating 30 min at 72°C in the presence of PCR buffer, dATP, MgCl_2 and T-PLUS DNA polymerase in the same concentrations used in previous amplification reactions. Cloning was carried out using the TOPO-TA Cloning[®] Kit for Sequencing (Invitrogen, Carlsbad, CA, USA). Plasmids were purified with QIAprep

Spin Miniprep Kit (Qiagen, Valencia, CA, USA) and sequenced at MacroGen Inc. (Seoul, South Korea) or at INTA (Castelar, Buenos Aires, Argentina), using M13 primers from the cloning vector.

Pyrosequencing of 16S rRNA Gene Amplicons

The V4 hypervariable region of the 16 S rRNA gene was amplified from the same template DNA used for the construction of *alkB* gene libraries. The conserved bacterial primers 520 F (5'-AYTGGGYDTAAAGNG-3') and 802R (5'-TACCRGGGTHCTAATCC-3', 5'TACCAGAGTATCTAATTC-3', 5'-CTACDSRGGTMTCTAATC-3', and 5'-TACNVGGGTATCTAATCC-3') were obtained from the RDP Pyrosequencing Pipeline (<http://pyro.cme.msu.edu>). Each primer contained the Roche 454 sequencing A (forward) and B (reverse) adaptors, and the forward primer also contained a ten nucleotide "multiple identifier" sequence (MID) to sort the samples.

Four independent PCR reactions were performed for each sample in a 25- μl volume containing 2.5 μl of FastStart High Fidelity 10 \times Reaction Buffer (Roche Applied Science, Mannheim, Germany), 20 ng of template DNA, 0.4 μM of each primer, 1.25-U FastStart High Fidelity Enzyme Blend (Roche Applied Science) and 0.2-mM dNTPs. Cycling conditions involved an initial 5-min denaturing step at 95°C, followed by 30 cycles of 45 s at 95°C, 45 s at 57°C and 60 s at 72°C, and a final 4-min elongation step at 72°C. Purified PCR products were pooled in an equimolar ratio, immobilized onto DNA capture beads, amplified through emulsion-based clonal amplification and sequenced together in one PicoTiterPlate device on a Genome Sequencer FLX (Roche Applied Science) using Titanium Chemistry at INDEAR Genome Sequencing Facility (Rosario, Argentina).

Bioinformatic and Statistical Analyses

Retrieved *alkB* gene nucleotide sequences were initially visualized and truncated to exclude primer and vector sequences using FinchTV 1.4.0 program (Geospiza Inc.). Nucleotide sequences were then translated with BioEdit software [55] and screened against the GenBank database using BLAST program (*tblastn* option) [56]. Deduced protein sequences (100 amino acid positions) were grouped into operational taxonomic units (OTUs) using mothur software v.1.22.2 [57]. Diversity estimators *Chao1* [58], ACE [59] and Good's coverage [60] were also calculated with mothur. For each OTU, one representative sequence from each library was kept for phylogenetic analysis. The sequences were aligned using ClustalX [61] tool within MEGA software version 5 [62]. In MEGA 5, a pairwise distance matrix was calculated based on Dayhoff substitution model with pairwise deletion of gaps

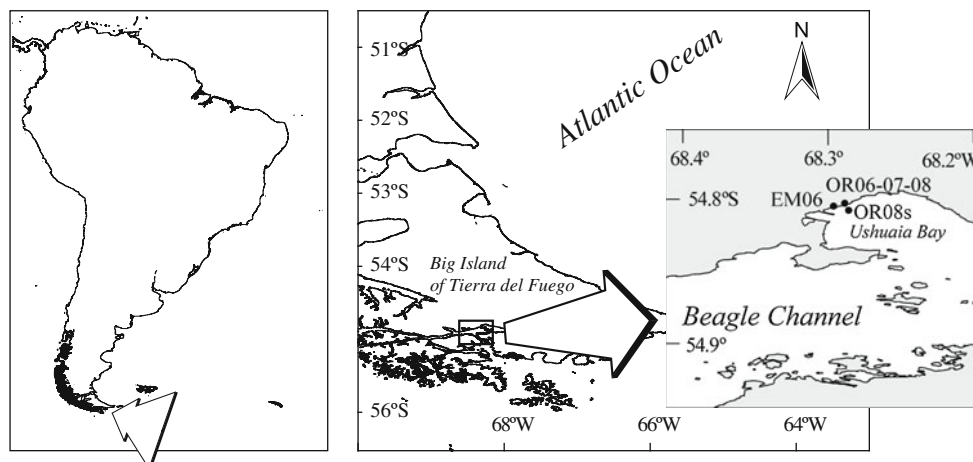
and missing data and was used to generate a phylogenetic tree by the neighbour-joining method with 1,000 bootstrap replicates.

Analyses of 16 S rRNA amplicons generated by 454 pyrosequencing were performed with mothur software v.1.22.2 [57] following the pipeline described by Schloss and Westcott [63], with minor modifications. Briefly, sequences were trimmed when the average quality score dropped below 25 within a 50-bp window, and sequences with at least one mismatch with the MID and/or two mismatches with the primer were eliminated. Trimmed sequences were aligned to the aligned SILVA reference database [64, 65]; the alignment was filtered to conserve 85 % of the sequences that overlapped in the same alignment space and used to perform OTU clustering. Before clustering, sequences were preclustered at 99 % to account for sequencing errors [66], and chimeras were removed by Chimera Slayer with mothur. OTUs were defined at a distance threshold of 0.03 (97 % identity cutoff). Alpha diversity estimators were calculated from OTU-based analyses using mothur. Alpha diversity metrics calculated included *Chao1* [58], Good's estimated coverage [60] and Simpson's Dominance index *D* [67]. Bray–Curtis similarity index [68] among samples was calculated based on shared OTUs at 97 % identity. Sequences were taxonomically identified up to the deepest possible phylogenetic level using the Ribosomal Database Project (RDP) [69] as reference with mothur, using a bootstrap cutoff of 50 % confidence. Sequences assigned to chloroplasts were excluded from the analysis. A table was built with relative frequencies of sequences at each taxonomic level for each sample.

Sequence Accession Numbers

The partial *alkB* sequences generated in the present study have been submitted to the GenBank database under accession numbers JQ437615–JQ437743. Partial 16 S rRNA sequences were deposited in the SRA database of NCBI under accession number SRA049611.

Figure 1 Sampling sites in Ushuaia Bay, Tierra del Fuego, Argentina. Samples are named as the site followed by the last two digits of the sampling year and followed by “s” for the subtidal sample. The map was constructed with Surfer version 8.02 (Golden Software, Inc.)



Results

Sampling Sites and Sample Hydrocarbon Content

Sampling sites are located in Ushuaia Bay, Argentina, in an area with high maritime traffic on the coast of Ushuaia city (Fig. 1). Intertidal sediment samples were retrieved at two sites 0.3 km apart: EM (54°48'19"S, 68°17'39"W) and OR (54°48'16"S, 68°17'23"W). The OR site is located next to a jetty used for loading and offloading of refined petroleum products. For details on sampling sites, see [47, 51].

Sediments contained total aliphatic hydrocarbons (TAH) in concentrations ranging from 19.7 to 347.2 µg/g dry weight sediment (dws, Online Resource 1). TAH content of OR samples from different years was fairly stable, while EM site showed the highest value (Online Resource 1). Values were in the range considered by UNEP/IOC/IAEA [70] as characteristic of polluted sediments from areas related to ports and industrial activities and in the category of moderately polluted sediments according to a previous survey carried out in the Patagonian coast [49]. All samples presented *n*-alkane homologous series, the isoprenoids pristane and phytane and unresolved complex mixture (UCM), characteristic of petrogenic origin. Biodegradation indices *n*C17/Pr and *n*C18/Ph ranged between 1.1 and 1.5 (Online Resource 1), suggesting ongoing biodegradation processes [52]. In addition, the ratio between total resolved aliphatics (TRA) and UCM (TRA/UCM) showed relatively low values (from 0.12 to 0.21, Online Resource 1) adding evidence to these results [49].

Diversity of *alkB* Gene Sequences in Subantarctic Coastal Sediments

Using a highly degenerate primer set [54], we amplified *alkB* gene fragments from five Subantarctic sediment samples and used them to build PCR clone libraries. We analysed by sequencing a total of 202 randomly chosen clones

from the five libraries (Online Resource 1). All deduced AlkB sequences matched alkane monooxygenases from the GenBank database with significant identity and low *e* values [56, 71]. With one exception, they contained the terminal histidines of the conserved histidine box II, essential for alkane hydroxylase activity [71–73]. In addition, other residues conserved in most alkane monooxygenases [72] were also present. Deduced protein sequences were grouped into operational taxonomic units (OTUs) with a distance threshold of 0.20 (80 % sequence identity at the amino acid level), due to the high sequence variability of this gene [24, 25]. Online Resource 2 shows an alignment of a representative AlkB sequence from each OTU.

For diversity analysis, the sequences from the five clone libraries were pooled. A total of 30 OTUs were identified, and we estimated 95 % coverage of the overall *alkB* gene diversity in this environment, as far as it can be amplified with our primer set (Table 1). Richness estimated by ACE and *Chao1* indices was 39 and 48 OTUs, respectively (Table 1). OTUs contained sequences with average distances of up to 0.14, indicating that there was internal variability within each OTU. This was also evidenced by the increase in the number of OTUs when the cutoff distance was decreased (Table 1). Our estimations can thus be considered as a conservative measure of the *alkB* richness in this chronically polluted environment.

The chosen primer set was able to recover a high number of *alkB* genes with wide phylogenetic diversity (Fig. 2), with the exception of *alkB2* from *Alcanivorax borkumensis* SK2 (BAC98366) and related sequences, which were not covered by our primer set [54]. Subantarctic sediment clone libraries were mostly composed by AlkB sequences related to the ones described in *Proteobacteria* (165 sequences, 82 %). Ninety-nine of these sequences grouped into a large and diverse cluster related to AlkB1 of *A. borkumensis*, which also includes sequences from microorganisms such as *Pseudomonas*, *Thalassolituus*, *Marinobacter* and *Oleiphilus* (Fig. 2). The rest of the sequences were related to different types of genes, including sequences from *Alcanivorax dieselolei*, *Alcanivorax venustensis* and *Alcanivorax hongdengensis*, and to AlkM sequences from *Acinetobacter*. With the exception of OTU #6, which showed 85 % amino acid identity to *Marinobacter* sp. S17-4 (ACT31523), the OTUs related to AlkB from *Proteobacteria* showed identity to sequences from isolated microorganisms ranging from

51 % to 74 %. Moreover, sequences retrieved from the libraries were more closely related to AlkB sequences from uncultured microorganisms from cold marine sediments or soils from high-latitude regions (Online Resource 3).

Sequences related to AlkB from *Actinobacteria* accounted for the next most abundant group (33 sequences, Fig. 2). The most abundant OTU (#24) was, however, most closely related to an Arctic soil clone, with no near match from a cultured representative (Online Resource 3). The other three minority OTUs showed high identity values to AlkB sequences from *Rhodococcus* and *Arthrobacter* (Online Resource 3). In addition, there were four divergent sequences detected in the EM06 sample, which clustered with AlkB sequences from members of the *Bacteroidetes*, forming deep but highly supported branches in the phylogenetic tree (Fig. 2).

Oil Exposure and Biostimulation Experiment

In order to evaluate the response of the indigenous alkane-degrading bacterial populations to crude oil exposure, we set up laboratory scale experiments (sediment-in-seawater slurries) using OR08 sediment sample. The three conditions were (1) 0.4 % (v/v) crude oil (expOR08-O), (2) 0.4 % (v/v) crude oil with the addition of 5-mM ammonium chloride and 0.25-mM sodium phosphate (expOR08-ON) and (3) a control with neither crude oil nor nutrient addition (expOR08-c) (Online Resource 1).

Alkane Biodegradation

After 20 days, expOR08-O slurry evidenced a decrease in the relative concentration of *n*-alkanes lower than *n*C20, with respect to the crude oil gas chromatography profile (Online Resource 4). This reduction extended to the majority of the resolved alkanes in the expOR08-ON slurry, where the profile was composed mainly of UCM (Online Resource 4). Accordingly, the biodegradation indices *n*C17/Pr, *n*C18/Ph and TRA/UCM decreased in both oil-exposed slurries with respect to the added crude oil, being lower in expOR08-ON. The ratio $\sum\text{HMW}/\sum\text{LMW}$ followed the same trend (Online Resource 1). These results indicate that after only 20 days, a certain level of biodegradation occurred in the expOR08-O slurry and that this process was more efficient with the addition of nutrients.

Table 1 *alkB* gene sequence diversity in coastal sediments from Ushuaia Bay

Number of samples	Number of sequences	Cutoff distance ^a	<i>S</i> _{obs} ^b	<i>Chao1</i>	ACE	Coverage
5	202	0.20	30	48 (34–105)	39 (32–61)	0.95
		0.02	46	72 (55–119)	71 (56–109)	0.89

^a Distance threshold at the amino acid level used to define OTUs

^b Number of observed OTUs. For richness estimators, 95 % confidence intervals are shown in parentheses

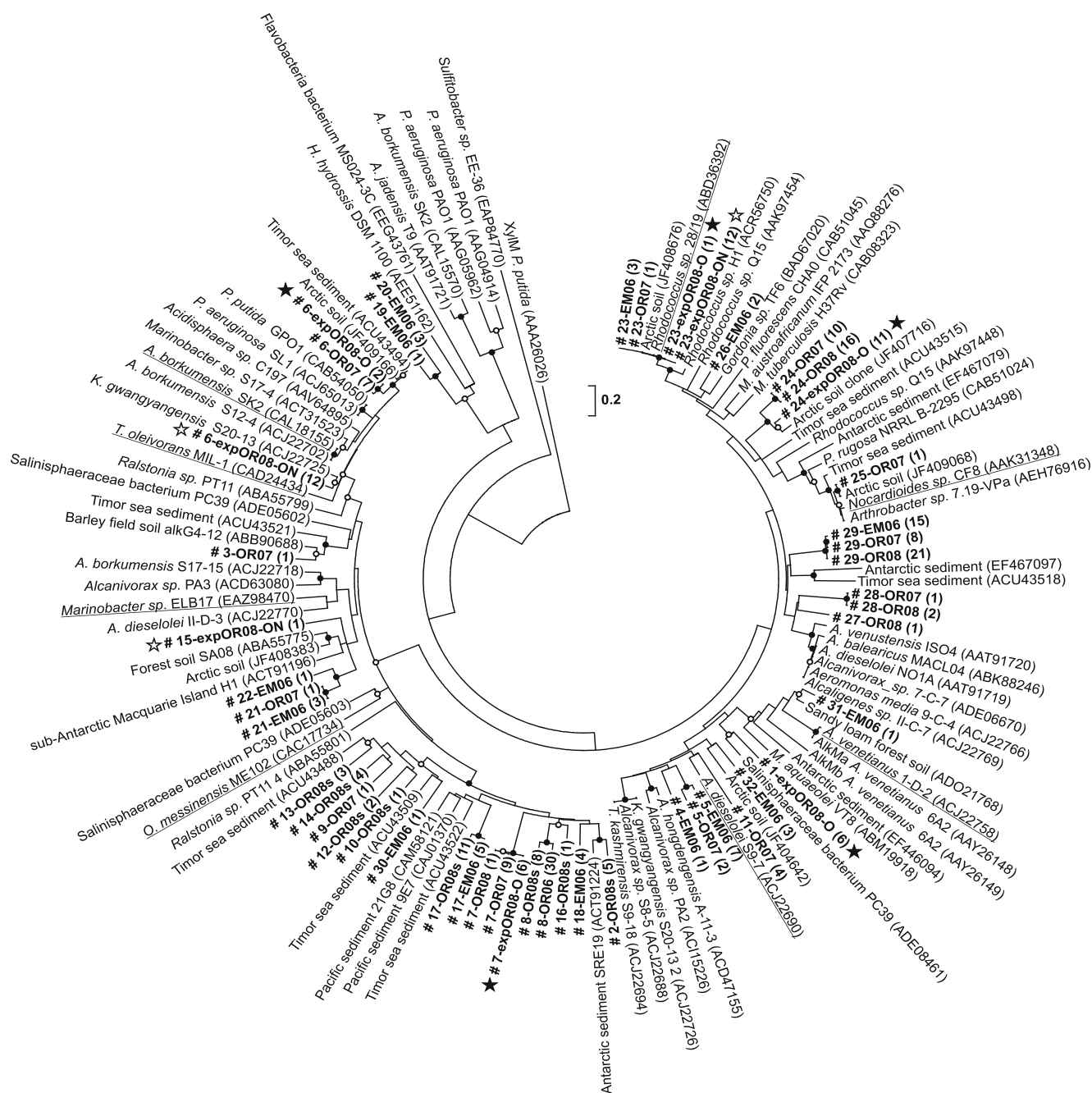


Figure 2 Phylogenetic tree of translated AlkB sequences from clones obtained in this study and representative sequences from the databases. Accession numbers from GenBank database are shown in parentheses. One representative sequence for each OTU formed at 0.20 cutoff distance from each library was used to build the tree. OTUs are shown in **bold type** as: **OTU number-library (number of sequences)**. Oil-exposed and oil-plus-nutrient experimental systems are indicated by

filled and open stars, respectively. The tree was constructed from deduced amino acid sequences using the neighbour-joining algorithm with MEGA 5 [62]. Bootstrap values ≥ 60 % and ≥ 90 % (1000 repetitions) are presented at nodes as open and filled circles, respectively. The scale bar represents 0.20 sequence divergence. The tree was rooted with xylene monooxygenase catalytic subunit (XylM) from *Pseudomonas putida* TOL plasmid (AAA26026)

Changes in *alkB* Gene Diversity

Five *alkB* gene OTUs were observed in the library of expOR08-O slurry after 20 days of oil exposure. Four of them were also present in the libraries constructed from the

environmental samples (Fig. 2, Online Resource 3). The most abundant OTU (#24, 42 % of the total sequences) was related to AlkB sequences mostly described in members of *Actinobacteria*, although no closely related sequence from an isolate has yet been described. OTUs #1 and #7

were the next most abundant gene sequences in the library (23 % each). OTU #1 clustered strongly with an AlkB sequence from an unclassified *Gammaproteobacteria* from the *Salinisphaerales* (Fig. 2). On the other hand, OTU #7 clustered in a highly supported branch including only sequences from uncultured microorganisms, mostly from deep sea sediments. This branch is in the clade where *A. borkumensis* AlkB1 is located (Fig. 2). The two minority OTUs were OTU #6, clearly affiliated with uncultured clones from the Arctic and deep sea, and #23 with 100 % identity to *Rhodococcus* sp. 28/19 (ABD36392) (Fig. 2).

In the *alkB* clone library from expOR08-ON slurry, only two OTUs (#6 and #23) accounted for 96 % of the sequences. They were equally represented and were also observed, although at low frequencies, in the expOR08-O slurry (Fig. 2). However, in this case, amino acid sequences belonging to OTU #6 were unequivocally placed with AlkB from *Kordiimonas gwangyangensis* (ACJ22725). OTU #23 was 100 % identical to AlkB from *Rhodococcus* sp. 28/19 (ABD36392).

PCR amplifications from the control slurry (expOR08-c) resulted in low yields and in the cloning of a high proportion of non-target DNA fragments, probably due to the low target concentration in this experimental condition.

Changes in Bacterial Community Structure

We analysed a total of 44,380 pyrosequencing reads of the V4 hypervariable region of the 16 S rRNA gene, amplified from the three slurries (expOR08-O, expOR08-ON and expOR08-c) as well as the original sediment sample (OR08) (Table 2). Coverage values were at least 94 % in all cases. The OR08 sample, as well as the expOR08-c and expOR08-O slurries, showed high bacterial richness and low dominance values. The OR08 sample was the most diverse, with approximately 1,800 and 3,000 observed and estimated OTUs, respectively (Table 2). The expOR08-c slurry experienced a decrease in richness of about 38 % with

respect to the original sediment, which can be attributed to the passage to laboratory conditions (Table 2). However, this slurry was the most similar to the original sediment (Bray–Curtis similarity index based on 16 S rRNA OTUs=0.48). The expOR08-O slurry showed slightly lower richness values than expOR08-c (Table 2), with Bray–Curtis similarity indices of 0.47 with expOR08-c and 0.34 with OR08. On the other hand, expOR08-ON was the most simplified, with only 163 and 288 observed and estimated OTUs, respectively (Table 2). This was also the only sample which showed an asymptotic behaviour in its rarefaction curve (data not shown). The structure of the bacterial community of the expOR08-ON slurry was also very different from the sediment and other treatments (Bray–Curtis similarity indices from 0.03 to 0.07).

Various bacterial genera known to include hydrocarbon degraders [9, 35, 74] were relatively abundant in Ushuaia Bay sediment sample OR08 (Fig. 3). These included *Oleispira*, *Pseudoalteromonas*, *Nocardioides*, *Sulfitobacter* and *Marinobacter*. Only *Sulfitobacter* and *Nocardioides* were also present in expOR08-O in relatively high abundances, while other genera were strongly enriched (*Spongiibacter*, *Sphingopyxis*, *Maribius* and *Robiginitomaculum*). Altogether, the aforementioned genera accounted for 38.5 % of the bacterial community of the expOR08-O slurry, as estimated by 16 S rRNA gene analysis.

In the expOR08-ON slurry, sequences classified as belonging to *Alcanivorax* were highly enriched (45.9 %), followed by *Thalassospira*, *Sulfitobacter* and various Flavobacteriales (*Flavobacterium*, *Bizionia* and *Gillisia*) (Fig. 3).

Discussion

The results of this study indicate the existence of a high diversity of relatively divergent *alkB* genes in the sediments of the chronically polluted environment of Ushuaia Bay.

Table 2 Bacterial diversity in OR08 sediment sample and slurries, as estimated by pyrosequencing of 16S rRNA amplicons

Sample		Number of sequences ^b	S_{obs} ^c	Coverage	<i>Chao1</i>	D^d
Type	Name ^a					
Sediment	OR08	15,344	1,829	0.94	3,011 (2,817–3,243)	0.043 (0.041–0.045)
Experimental systems	expOR08-c	9,357	1,128	0.94	1,961 (1,788–2,179)	0.034 (0.032–0.036)
	expOR08-O	11,636	896	0.96	1,503 (1,366–1,680)	0.033 (0.032–0.035)
	expOR08-ON	8,043	163	0.99	288 (230–398)	0.247 (0.238–0.256)

^a expOR08-c control slurry (nor oil neither nutrients added), expOR08-O oil-exposed slurry, expOR08-ON oil-plus-nutrient amended slurry

^b Obtained after discarding low-quality reads, chimeras and sequences assigned to chloroplasts

^c Number of observed OTUs, defined at 0.03 distance threshold

^d Simpson's dominance index [66]. For richness and dominance estimators, 95 % confidence intervals are shown in parentheses

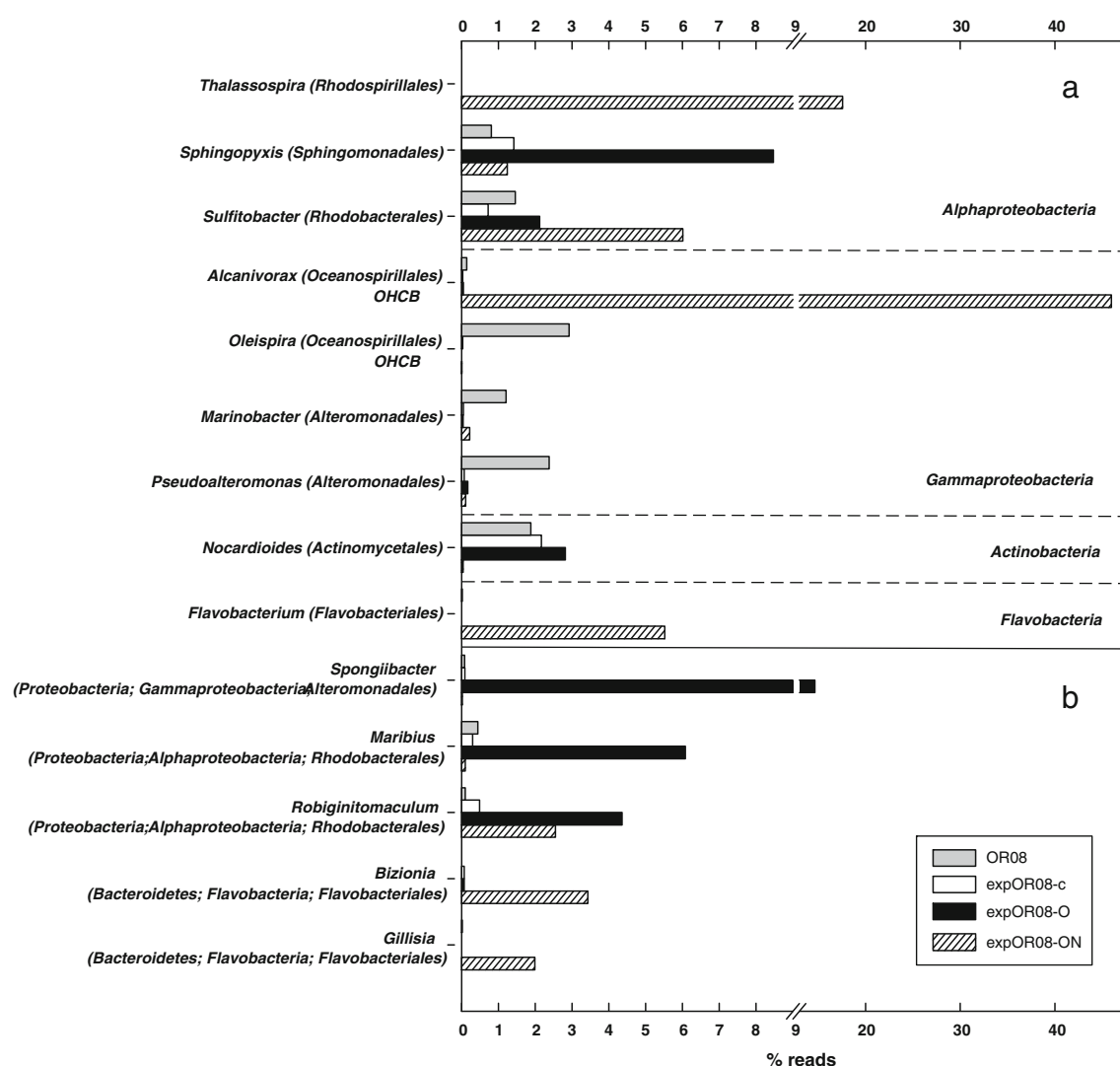


Figure 3 Relative frequency of selected bacterial genera from OR08 sediment sample and experimental systems, as estimated by 16 S rRNA amplicon pyrosequencing analysis. **a** Genera linked to hydrocarbon biodegradation in marine environments [9, 35, 74]. **b** Genera not previously linked to hydrocarbon biodegradation, but showing important shifts in frequency among treatments. Only genera with

frequency values above 1 % in at least one sample are shown. *expOR08-c* control slurry (nor oil neither nutrients added), *expOR08-O* oil-exposed slurry, *expOR08-ON* oil-plus-nutrient amended slurry, *OHC* obligate hydrocarbonoclastic bacteria. Analyses were performed with mothur v.1.22.2

Moreover, the most abundant *alkB* gene variants found in our study were detected across sampling sites and years, and various were further selected in oil enrichment experiments, suggesting the ecological relevance of bacterial populations carrying these genes in Subantarctic sediments. Chronic hydrocarbon pollution at this site [46, 49] has been key for the establishment of an active and diverse bacterial alkane-degrading guild in these sediments. In general, the closest matches of *AlkB* sequences from this study were sequences from uncultured microorganisms, in accordance with reports from other underexplored coastal and deep sea environments [24, 25]. The matches most often found were *alkB* genes detected in diesel-contaminated Antarctic marine sediments (GenBank acc. no. ACT91224 and related sequences),

Arctic soils contaminated with high-alkane content fuel [75] and natural seepages in deep-sea sediments [24]. Various components including type of pollution and environmental factors such as temperature seem to be shaping the bacterial guilds associated with aerobic alkane biodegradation. Studies are needed in this direction in order to quantitatively assess the relative contribution of the different environmental factors to the structuring of the hydrocarbon-degrading bacterial guilds in the marine environment.

Ushuaia Bay sediments receive small and frequent inputs of refined petroleum products [46, 49]. The addition of a different type of oil (crude oil) resulted in a rapid disappearance of *n*-alkanes and a concomitant shift in bacterial community structure that was different in each nutritional

condition. Other studies have linked preexposure to hydrocarbons to a fast structural and functional response of the microbial community in soils [76, 77] and temperate sediments [14]. As regards to the suitability of the *alkB* gene to reflect the response of the bacterial community to oil exposure, results are still controversial. Hamamura and collaborators were able to detect key alkane-degrading populations in soil with various group-specific *alkB* primers and proposed their use as functional biomarkers [44]. In a study involving quantitative PCR, the proportion of *alkB*-carrying microorganisms was positively correlated to the concentration of *n*-alkanes in soil [42]. On the other hand, in experiments involving temperate coastal sediments, the expression of this gene could not be linked to the community response to oil pollution probably due to the transient nature of the transcripts, although oiled microcosms showed higher diversity of *alkB* genes than controls [78]. We assessed the selection of bacterial populations carrying *alkB* genes after oil exposure, as measured by DNA-based analysis and a time scale of 20 days. The presence of oil resulted in strong amplification and selection of specific *alkB* gene sequences in both conditions (oil and oil-plus-nutrients). In contrast, poor amplification and cloning of unspecific products were observed in the control, suggesting lack of selection of *alkB*-carrying bacterial populations after the dilution of the alkanes present in OR08 sample due to the addition of non-polluted seawater in this slurry. The identification of *alkB* gene sequences from key populations achieved in this study is the first step towards their targeted use as functional biomarkers using molecular techniques with quantitative power, in both basic and applied contexts.

In accordance with the detection of *alkB* gene sequences affiliated with the *Proteobacteria* in the oil-exposed slurry, members of the *Gammaproteobacteria* and *Alphaproteobacteria* were identified as highly abundant by 16 S rRNA gene amplicon pyrosequencing. Of these populations, the ones classified as belonging to the *Sphingibacter* genus (*Gammaproteobacteria*, *Alteromonadales*, formerly *Melitea* [79, 80]) were the most prominent. Bacteria belonging to this genus have been isolated from seawater and described as truly marine as they require salt for growth [81]. Although to the best of our knowledge members of this genus have not been previously linked to hydrocarbon biodegradation, strains from the closely related genus *Marinobacter* [82, 83] are well known to participate in hydrocarbon biodegradation in sediments [19, 20]. The *alkB* genes related to the ones described in *Proteobacteria* that were identified in this slurry could be carried by *Sphingibacter* populations, although this possibility has to be confirmed. Another *alkB* gene strongly selected in this experimental condition, as well as abundant in the sediments obtained at the OR site in two consecutive years, was closely related to an Arctic soil clone. Its deduced protein sequence clustered, among

others, with AlkB from *Nocardioides*. Coincidentally, *Nocardioides* 16 S rRNA sequences were relatively abundant in the oil-exposed slurries and also in the sediment sample, suggesting that this *alkB* gene (OTU #24) could be carried by members of this genus.

In the oil-plus-nutrient amended slurry, the high abundance of *alkB* sequences related to genes from *Proteobacteria* was evident. In accordance, *Alcanivorax* (*Gammaproteobacteria*) and *Thalassospira* (*Alphaproteobacteria*) were detected as very abundant by our phylogenetic-based analysis. The capability of *Alcanivorax* for scavenging nutrients is well known [35, 84]. On the other hand, *Thalassospira*, originally associated with polycyclic aromatic hydrocarbon biodegradation [85, 86], has been recently linked to alkane biodegradation [87]. Therefore, we cannot exclude the possibility that *alkB* gene OTU #6 belongs to this genus. However, the analysis based on the *alkB* gene precludes us from drawing more definite conclusions, due to the possibility of horizontal gene transfer [88], the presence of multiple and divergent *alkB* genes in members of the same genus, characteristic of *Alcanivorax* [89] and lack of catabolic gene information from *Thalassospira*. The other *alkB* gene found in a very high abundance in this slurry was identical at the amino acid level to AlkB from *Rhodococcus* (OTU #23). This result was not in accordance with the 16 S rRNA gene pyrosequencing data, where only 0.25 % of sequences were classified as *Rhodococcus*. This lack of agreement may be due to a bias introduced by the *alkB* primer set, which was designed using a majority of *Rhodococcus* sequences [54]. Alternatively, it is possible that bacteria other than *Rhodococcus* could be carrying this gene, acquired by horizontal gene transfer [90]. Interestingly, this same discrepancy has been recently observed for *Rhodococcus* genes in a metagenomic study of diesel-contaminated Arctic soils [91]. The complementation of functional gene-based information with a phylogenetic approach allowed us to build a more complete picture of the bacterial communities inhabiting Subantarctic sediments and their response to an oil input. For example, sequences classified as belonging to *Oleispira* [34, 92] were detected in Ushuaia Bay sediments in relatively high abundance by 16 S rRNA gene amplicon pyrosequencing. These microorganisms were not selected in any of the slurries, which were performed at 15°C resembling Ushuaia summer temperatures [48]. Both the latitude of Ushuaia and our experimental results suggest that *Oleispira* populations present in these sediments could be more closely related to the psychrophilic bacterium *Oleispira antarctica* [34] than to the recently described *Oleispira lenta*, whose optimum growth temperature is 28°C [92]. Although its role in hydrocarbon biodegradation has been recognized [35, 93, 94], the pathway genes of *Oleispira* have not yet been described. The high-throughput information provided by pyrosequencing was fundamental for the detection of members of this genus in the complex microbial community of

Ushuaia Bay sediments. Other microorganisms for which no *alkB* information is known, or even no biodegradation capabilities have been reported, could also be identified as relevant in our study when using the phylogenetic approach. In the slurries, genera such as *Spongiibacter*, *Maribius*, *Robiginitomaculum*, *Thalassospira*, *Flavobacterium*, *Bizionia* and *Gillisia* were detected in relatively high abundances with respect to the control slurry. These genera constitute candidates for future work involving identification of hydrocarbon biodegradation pathway genes.

Insufficient information is still hampering our understanding of hydrocarbon biodegradation processes in cold marine environments. The results from this study contribute to the knowledge on the identity and behaviour of key alkane-degrading populations from Subantarctic regions.

Acknowledgments LMG, CLL and MSM are recipients of graduate student fellowships from the National Research Council of Argentina (CONICET). ML, HMD and MGC are staff members from CONICET. Grants from CONICET, National Agency for the Promotion of Science and Technology (ANPCyT, Argentina) and Secretary of Science, Technology and Innovation of the Chubut Province (Argentina) supported this research.

We would like to thank M. Gil, J. L. Esteves, H. Ocariz, A. Torres and R. Vera for their help during sample collection.

References

- Marshall AG, Rodgers RP (2008) Petroleomics: chemistry of the underworld. *Proc Natl Acad Sci* 105:18090–18095
- NRC National Research Council (2003) Oil in the sea III: inputs, fates and effects. National Academies Press, Washington, D.C
- Peterson CH, Rice SD, Short JW, Esler D, Bodkin JL, Ballachey BE, Irons DB (2003) Long-term ecosystem response to the Exxon Valdez oil spill. *Science* 302:2082–2086
- Short JW, Maselko JM, Lindeberg MR, Harris PM, Rice SD (2006) Vertical distribution and probability of encountering intertidal Exxon Valdez oil on shorelines of three embayments within Prince William Sound, Alaska. *Environ Sci Technol* 40:3723–3729
- Short JW, Irvine GV, Mann DH, Maselko JM, Pella JJ, Lindeberg MR, Payne JR, Driskell WB, Rice SD (2007) Slightly weathered Exxon Valdez oil persists in Gulf of Alaska beach sediments after 16 years. *Environ Sci Technol* 41:1245–1250
- Irvine GV, Mann DH, Short JW (2006) Persistence of 10-year old Exxon Valdez oil on Gulf of Alaska beaches: the importance of boulder-armoring. *Mar Pollut Bull* 52:1011–1022
- Atlas RM (2010) Microbial bioremediation in polar environments: current status and future directions. In: Bej AK, Aislabie J, Atlas RM (eds) Polar microbiology: the ecology, biodiversity and bioremediation potential of microorganisms in extremely cold environments. CRC Press, Boca Raton, pp 373–391
- Coulon F, Chronopoulou P-M, Fahy A, Paissé S, Goñi-Urriza M, Peperzak L, Acuña Alvarez L, McKew BA, Brussaard CPD, Underwood GJC, Timmis KN, Duran R, McGenity TJ (2012) Hydrocarbon biodegradation in coastal mudflats: the central role of dynamic tidal biofilms dominated by aerobic hydrocarbonoclastic bacteria and diatoms. *Appl Environ Microbiol* in press. doi:10.1128/aem.00072-12
- Prince RC, Gramain A, McGenity TJ (2010) Prokaryotic hydrocarbon degraders. In: Timmis KN (ed) Handbook of hydrocarbon and lipid microbiology, vol 3. Springer-Verlag, Berlin Heidelberg, pp 1671–1692
- Röling WF, Milner MG, Jones DM, Lee K, Daniel F, Swannell RJ, Head IM (2002) Robust hydrocarbon degradation and dynamics of bacterial communities during nutrient-enhanced oil spill bioremediation. *Appl Environ Microbiol* 68:5537–5548
- Yakimov MM, Denaro R, Genovese M, Cappello S, D'Auria G, Chernikova TN, Timmis KN, Golyshin PN, Giluliano L (2005) Natural microbial diversity in superficial sediments of Milazzo Harbor (Sicily) and community successions during microcosm enrichment with various hydrocarbons. *Environ Microbiol* 7:1426–1441
- Paisse S, Coulon F, Goñi-Urriza M, Peperzak L, McGenity TJ, Duran R (2008) Structure of bacterial communities along a hydrocarbon contamination gradient in a coastal sediment. *FEMS Microbiol Ecol* 66:295–305
- Alonso-Gutiérrez J, Figueras A, Albaigés J, Jiménez N, Viñas M, Solanas AM, Novoa B (2009) Bacterial communities from shoreline environments (Costa da Morte, Northwestern Spain) affected by the Prestige oil spill. *Appl Environ Microbiol* 75:3407–3418
- Paisse S, Goñi-Urriza M, Coulon F, Duran R (2010) How a bacterial community originating from a contaminated coastal sediment responds to an oil input. *Microb Ecol* 60:394–405
- Zhou HW, Guo CL, Wong YS, Tam NF (2006) Genetic diversity of dioxygenase genes in polycyclic aromatic hydrocarbon-degrading bacteria isolated from mangrove sediments. *FEMS Microbiol Lett* 262:148–157
- Marcial Gomes NC, Borges LR, Paranhos R, Pinto FN, Mendonça-Hagler LC, Smalla K (2008) Exploring the diversity of bacterial communities in sediments of urban mangrove forests. *FEMS Microbiol Ecol* 66:96–109
- Zhou HW, Wong AH, Yu RM, Park YD, Wong YS, Tam NF (2009) Polycyclic aromatic hydrocarbon-induced structural shift of bacterial communities in mangrove sediment. *Microb Ecol* 58:153–160
- Gomes NC, Flocco CG, Costa R, Junca H, Vilchez R, Pieper DH, Krögercklenfort E, Paranhos R, Mendonça-Hagler LC, Smalla K (2010) Mangrove microniches determine the structural and functional diversity of enriched petroleum hydrocarbon-degrading consortia. *FEMS Microbiol Ecol* 74:276–290
- dos Santos HF, Cury JC, do Carmo FL, dos Santos AL, Tiedje J, van Elsas JD, Rosado AS, Peixoto RS (2011) Mangrove bacterial diversity and the impact of oil contamination revealed by pyrosequencing: bacterial proxies for oil pollution. *PLoS One* 6:e16943
- Kostka JE, Prakash O, Overholt WA, Green SJ, Freyer G, Canion A, Delgadino J, Norton N, Hazen TC, Huettel M (2011) Hydrocarbon-degrading bacteria and the bacterial community response in Gulf of Mexico beach sands impacted by the Deepwater Horizon oil spill. *Appl Environ Microbiol* 77:7962–7974
- Peixoto R, Chaer GM, Carmo FL, Araújo FV, Paes JE, Volpon A, Santiago G, Rosado A (2011) Bacterial communities reflect the spatial variation in pollutant levels in Brazilian mangrove sediment. *Antonie Van Leeuwenhoek* 99:341–354
- Rosano-Hernández M, Ramírez-Saad H, Fernández-Linares L (2011) Petroleum-influenced beach sediments of the Campeche Bank, Mexico: diversity and bacterial community structure assessment. *J Environ Manage* in press
- Cui Z, Lai Q, Dong C, Shao Z (2008) Biodiversity of polycyclic aromatic hydrocarbon-degrading bacteria from deep sea sediments of the Middle Atlantic Ridge. *Environ Microbiol* 10:2138–2149
- Wasmund K, Burns KA, Kurtböke DI, Bourne DG (2009) Novel alkane hydroxylase gene (*alkB*) diversity in sediments associated with hydrocarbon seeps in the Timor Sea, Australia. *Appl Environ Microbiol* 75:7391–7398

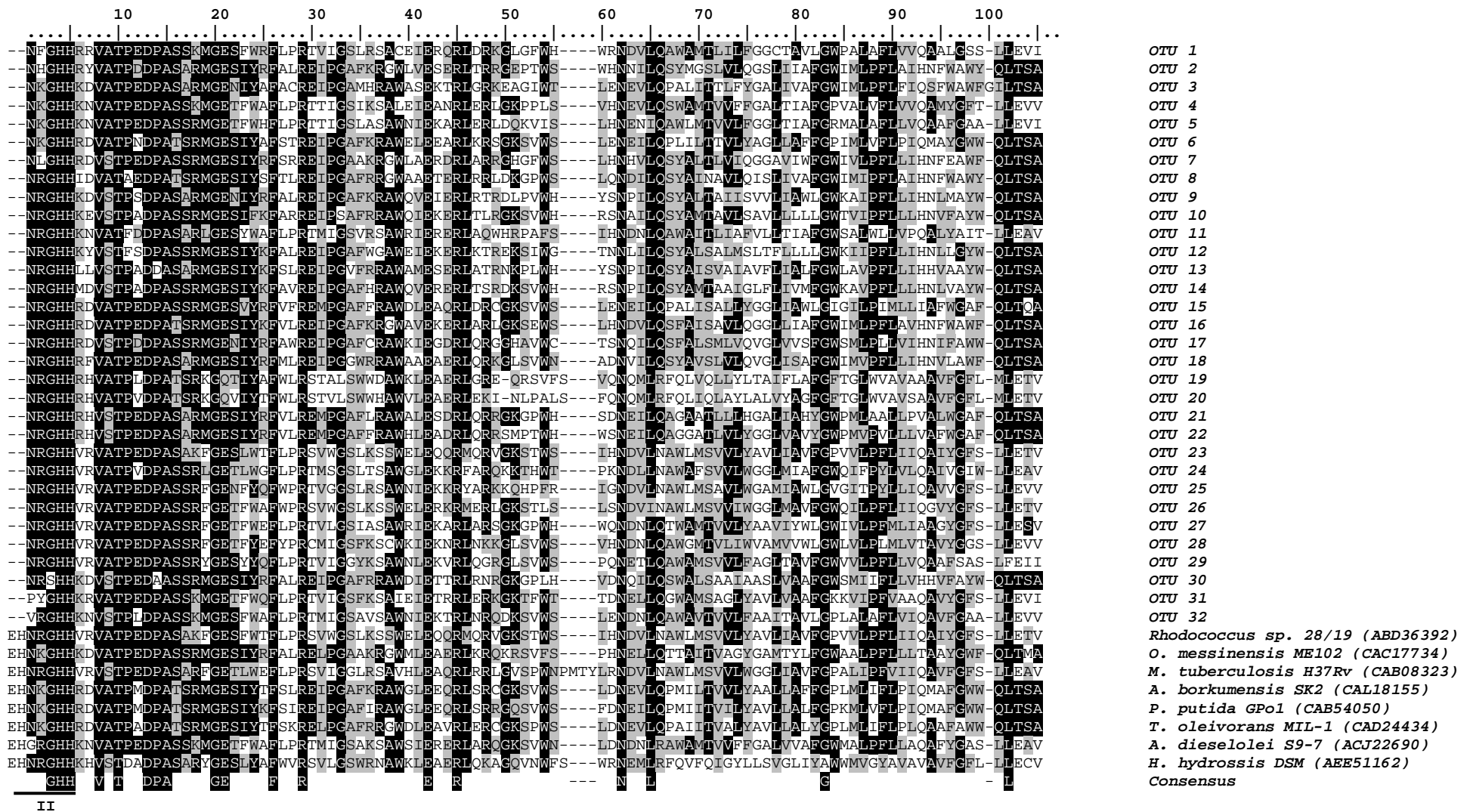
25. Kuhn E, Bellicanta GS, Pellizari VH (2009) New *alk* genes detected in Antarctic marine sediments. *Environ Microbiol* 11:669–673
26. Marcos M, Lozada M, di Marzio WD, Dionisi H (2012) Abundance, dynamics and biogeographic distribution of seven polycyclic aromatic hydrocarbon dioxygenase gene variants in coastal sediments of Patagonia. *Appl Environ Microbiol* 78:1589–1592
27. Head IM, Jones DM, Røling WF (2006) Marine microorganisms make a meal of oil. *Nat Rev Microbiol* 4:173–182
28. Harayama S, Kishira H, Kasai Y, Shutsubo K (1999) Petroleum biodegradation in marine environments. *J Mol Microbiol Biotechnol* 1:63–70
29. Kuhad RC, Gupta R (2009) Biological remediation of petroleum contaminants. In: Singh A, Kuhad RC, Ward OP (eds) *Advances in applied bioremediation*, vol 17. Springer-Verlag, Berlin Heidelberg, pp 173–187
30. Nava V, Morales M, Revah S (2007) Cometabolism of methyl tert-butyl ether (MTBE) with alkanes. *Rev Environ Sci Biotechnol* 6:339–352
31. Yakimov MM, Golyshin PN, Lang S, Moore ER, Abraham WR, Lünsdorf H, Timmis KN (1998) *Alcanivorax borkumensis* gen. nov., sp. nov., a new, hydrocarbon-degrading and surfactant-producing marine bacterium. *Int J Syst Bacteriol* 48:339–348
32. Yakimov MM, Giuliano L, Denaro R, Crisafi E, Chernikova TN, Abraham WR, Luensdorf H, Timmis KN, Golyshin PN (2004) *Thalassolituus oleivorans* gen. nov., sp. nov., a novel marine bacterium that obligately utilizes hydrocarbons. *Int J Syst Evol Microbiol* 54:141–148
33. Golyshin PN, Chernikova TN, Abraham WR, Lünsdorf H, Timmis KN, Yakimov MM (2002) *Oleiphilaceae* fam. nov., to include *Oleiphilus messinensis* gen. nov., sp. nov., a novel marine bacterium that obligately utilizes hydrocarbons. *Int J Syst Evol Microbiol* 52:901–911
34. Yakimov MM, Giuliano L, Gentile G, Crisafi E, Chernikova TN, Abraham W-R, Lünsdorf H, Timmis KN, Golyshin PN (2003) *Oleispira antarctica* gen. nov., sp. nov., a novel hydrocarbonoclastic marine bacterium isolated from Antarctic coastal sea water. *Int J Syst Evol Microbiol* 53:779–785
35. Yakimov MM, Timmis KN, Golyshin PN (2007) Obligate oil-degrading bacteria. *Curr Opin Biotechnol* 18:257–266
36. van Beilen JB, Funhoff EG (2007) Alkane hydroxylases involved in microbial alkane degradation. *Appl Microbiol Biotechnol* 74:13–21
37. Wentzel A, Ellingsen TE, Kotlar HK, Zotchev SB, Throne-Holst M (2007) Bacterial metabolism of long-chain *n*-alkanes. *Appl Microbiol Biotechnol* 76:1209–1221
38. Rojo F (2009) Degradation of alkanes by bacteria. *Environ Microbiol* 11:2477–2490
39. Margesin R, Labbé D, Schinner F, Greer CW, Whyte LG (2003) Characterization of hydrocarbon-degrading microbial populations in contaminated and pristine Alpine soils. *Appl Environ Microbiol* 69:3085–3092
40. Pérez-de-Mora A, Engel M, Schlöter M (2010) Abundance and diversity of *n*-alkane-degrading bacteria in a forest soil co-contaminated with hydrocarbons and metals: a molecular study on *alkB* homologous genes. *Microb Ecol* 62:959–972
41. Powell SM, Bowman JP, Ferguson SH, Snape I (2010) The importance of soil characteristics to the structure of alkane-degrading bacterial communities on sub-Antarctic Macquarie Island. *Soil Biol Biochem* 42:2012–2021
42. Powell SM, Ferguson SH, Bowman JP, Snape I (2006) Using real-time PCR to assess changes in the hydrocarbon-degrading microbial community in Antarctic soil during bioremediation. *Microb Ecol* 52:523–532
43. Salminen JM, Tuomi PM, Jørgensen KS (2008) Functional gene abundances (*nahAc*, *alkB*, *xyIE*) in the assessment of the efficacy of bioremediation. *Appl Biochem Biotechnol* 151:638–652
44. Hamamura N, Fukui M, Ward DM, Inskeep WP (2008) Assessing soil microbial populations responding to crude-oil amendment at different temperatures using phylogenetic, functional gene (*alkB*) and physiological analyses. *Environ Sci Technol* 42:7580–7586
45. Commendatore MG, Esteves JL (2007) An assessment of oil pollution in the coastal zone of Patagonia, Argentina. *Environ Manage* 40:814–821
46. Esteves JL, Commendatore MG, Nievas ML, Paletto VM, Amín O (2006) Hydrocarbon pollution in coastal sediments of Tierra del Fuego Islands, Patagonia Argentina. *Mar Pollut Bull* 52:582–590
47. Lozada M, Riva Mercadal JP, Guerrero LD, Di Marzio WD, Ferrero MA, Dionisi HM (2008) Novel aromatic ring-hydroxylating dioxygenase genes from coastal marine sediments of Patagonia. *BMC Microbiol* 8:50
48. Gil MN, Torres AI, Amín O, Esteves JL (2011) Assessment of recent sediment influence in an urban polluted Subantarctic coastal ecosystem. Beagle Channel (Southern Argentina). *Mar Pollut Bull* 6:201–207
49. Commendatore MG, Nievas ML, Amín O, Esteves JL (2012) Sources and distribution of aliphatic and polyaromatic hydrocarbons in coastal sediments from the Ushuaia Bay (Tierra del Fuego, Patagonia, Argentina). *Mar Environ Res* 74:20–31
50. Marcos MS, Lozada M, Dionisi HM (2009) Aromatic hydrocarbon degradation genes from chronically polluted Subantarctic marine sediments. *Lett Appl Microbiol* 49:602–608
51. Dionisi HM, Lozada M, Marcos MS, Di Marzio WD (2011) Aromatic hydrocarbon degradation genes from chronically polluted Subantarctic marine sediments. In: Bruijn FJd (ed) *Handbook of molecular microbial ecology II: metagenomics in different habitats*. John Wiley & Sons, Inc., Hoboken, pp 461–473
52. Commendatore MG, Esteves JL, Colombo JC (2000) Hydrocarbons in coastal sediments of Patagonia, Argentina: levels and probable sources. *Mar Pollut Bull* 40:989–998
53. Commendatore MG, Esteves JL (2004) Natural and anthropogenic hydrocarbons in sediments from the Chubut River (Patagonia, Argentina). *Mar Pollut Bull* 48:910–918
54. Olivera NL, Nievas ML, Lozada M, Del Prado G, Dionisi HM, Siñeriz F (2009) Isolation and characterization of biosurfactant-producing *Alcanivorax* strains: hydrocarbon accession strategies and alkane hydroxylase gene analysis. *Res Microbiol* 160:19–26
55. Hall TA (1999) BioEdit: a user-friendly biological sequence alignment editor and analysis program for Windows 95/98/NT. *Nucl Acids Symp Ser* 41:95–98
56. Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ (1990) Basic local alignment search tool. *J Mol Biol* 215:403–410
57. Schloss PD, Westcott SL, Ryabin T, Hall JR, Hartmann M, Hollister EB, Lesniewski RA, Oakley BB, Parks DH, Robinson CJ, Sahl JW, Stres B, Thallinger GG, Van Horn DJ, Weber CF (2009) Introducing mothur: open-source, platform-independent, community-supported software for describing and comparing microbial communities. *Appl Environ Microbiol* 75:7537–7541
58. Chao A (1984) Nonparametric estimation of the number of classes in a population. *Scand J Stat* 11:265–270
59. Chao A, Ma M-C, Yang MCK (1993) Stopping rules and estimation for recapture debugging with unequal failure rates. *Biometrika* 80:193–201
60. Good IJ (1953) The population frequencies of species and the estimation of population parameters. *Biometrika* 40:237–264
61. Thompson JD, Gibson TJ, Plewniak F, Jeanmougin F, Higgins DG (1997) The CLUSTAL_X windows interface: flexible strategies for multiple sequence alignment aided by quality analysis tools. *Nucleic Acids Res* 25:4876–4882
62. Tamura K, Peterson D, Peterson N, Stecher G, Nei M, Kumar S (2011) MEGA5: molecular evolutionary genetics analysis using maximum likelihood, evolutionary distance, and maximum parsimony methods. *Mol Biol Evol* 28:2731–2739

63. Schloss PD, Westcott SL (2011) Assessing and improving methods used in operational taxonomic unit-based approaches for 16S rRNA gene sequence analysis. *Appl Environ Microbiol* 77:3219–3226
64. Pruesse E, Quast C, Knittel K, Fuchs BM, Ludwig W, Peplies J, Glöckner FO (2007) SILVA: a comprehensive online resource for quality checked and aligned ribosomal RNA sequence data compatible with ARB. *Nucleic Acids Res* 35:7188–7196
65. Schloss PD (2010) The effects of alignment quality, distance calculation method, sequence filtering, and region on the analysis of 16S rRNA gene-based studies. *PLoS Comput Biol* 6:e1000844
66. Huse SM, Welch DM, Morrison HG, Sogin ML (2010) Ironing out the wrinkles in the rare biosphere through improved OTU clustering. *Environ Microbiol* 12:1889–1898
67. Simpson EH (1949) Measurement of diversity. *Nature* 163:688
68. Bray JR, Curtis CT (1957) An ordination of the upland forest communities of Southern Wisconsin. *Ecol Monogr* 27:325–349
69. Wang Q, Garrity GM, Tiedje JM, Cole JR (2007) Naïve Bayesian classifier for rapid assignment of rRNA sequences into the new bacterial taxonomy. *Appl Environ Microbiol* 73:5261–5267
70. United Nations Environment Programme (1992) Determination of petroleum hydrocarbons in sediments. In: UNEP (ed.) Reference methods for marine pollution studies, vol. 20. UNEP/IOC/IAEA, pp. 1–75.
71. Altschul SF (1991) Amino acid substitution matrices from an information theoretic perspective. *J Mol Biol* 219:555–565
72. Smits TH, Rothlisberger M, Witholt B, van Beilen JB (1999) Molecular screening for alkane hydroxylase genes in Gram-negative and Gram-positive strains. *Environ Microbiol* 1:307–317
73. Shanklin J, Whittle E (2003) Evidence linking the *Pseudomonas oleovorans* alkane ω -hydroxylase, an integral membrane diiron enzyme, and the fatty acid desaturase family. *FEBS Lett* 545:188–192
74. Kim SJ, Kwon KK (2010) Marine, hydrocarbon-degrading Alphaproteobacteria. In: Timmis KN (ed) Handbook of hydrocarbon and lipid microbiology, vol 3. Springer-Verlag, Berlin Heidelberg, pp 1707–1714
75. Bell TH, Yergeau E, Martineau C, Juck D, Whyte LG, Greer CW (2011) Identification of nitrogen-incorporating bacteria in petroleum-contaminated arctic soils using [^{15}N]DNA-based stable isotope probing and pyrosequencing. *Appl Environ Microbiol* 12:4163–4171
76. Greenwood PF, Wibrow S, George SJ, Tibbett M (2009) Hydrocarbon biodegradation and soil microbial community response to repeated oil exposure. *Org Geochem* 40:293–300
77. Kauppi S, Romantschuk M, Strömmér R, Sinkkonen A (2010) Natural attenuation is enhanced in previously contaminated and coniferous forest soils. *Environ Sci Pollut Res Int* 19:53–63
78. Paise S, Duran R, Coulon F, Goñi-Urriza M (2011) Are alkane hydroxylase genes (*alkB*) relevant to assess petroleum bioremediation processes in chronically polluted coastal sediments? *Appl Microbiol Biotechnol* 92:835–844
79. Graeber I, Kaesler I, Borchert MS, Dieckmann R, Pape T, Lurz R, Nielsen P, von Döhren H, Michaelis W, Szewzyk U (2008) *Spongiibacter marinus* gen. nov., sp. nov., a halophilic marine bacterium isolated from the boreal sponge *Haliclona* sp. 1. *Int J Syst Evol Microbiol* 58:585–590
80. Jang GI, Hwang CY, Choi H-G, Kang S-H, Cho BC (2011) Description of *Spongiibacter borealis* sp. nov., isolated from Arctic seawater, and reclassification of *Melitea salexigens* Urios *et al.* 2008 as a later heterotypic synonym of *Spongiibacter marinus* Graeber *et al.* 2008 with emended descriptions of the genus *Spongiibacter* and *Spongiibacter marinus*. *Int J Syst Evol Microbiol* 61:2895–2900
81. Urios L, Agogué H, Intertaglia L, Lesongeur F, Lebaron P (2008) *Melitea salexigens* gen. nov., sp. nov., a gammaproteobacterium from the Mediterranean Sea. *Int J Syst Evol Microbiol* 58:2479–2483
82. Gauthier MJ, Lafay B, Christen R, Fernandez L, Acquaviva M, Bonin P, Bertrand JC (1992) *Marinobacter hydrocarbonoclasticus* gen. nov., sp. nov., a new, extremely halotolerant, hydrocarbon-degrading marine bacterium. *Int J Syst Bacteriol* 42:568–576
83. Márquez MC, Ventosa A (2005) *Marinobacter hydrocarbonoclasticus* Gauthier *et al.* 1992 and *Marinobacter aquaeolei* Nguyen *et al.* 1999 are heterotypic synonyms. *Int J Syst Evol Microbiol* 55:1349–1351
84. Schaefer S, Martins dos Santos VA, Bartels D, Bekel T, Brecht M, Buhrmester J, Chernikova TN, Denaro R, Ferrer M, Gertler C, Goesmann A, Golyshina OV, Kaminski F, Khachane AN, Lang S, Linke B, McHardy AC, Meyer F, Nechitaylo T, Puhler A, Regenhart D, Rupp O, Sabirova JS, Selbitschka W, Yakimov MM, Timmis KN, Vorholter FJ, Weidner S, Kaiser O, Golyshin PN (2006) Genome sequence of the ubiquitous hydrocarbon-degrading marine bacterium *Alcanivorax borkumensis*. *Nat Biotechnol* 24:997–1004
85. Zhao B, Wang H, Li R, Mao X (2010) *Thalassospira xianhensis* sp. nov., a polycyclic aromatic hydrocarbon-degrading marine bacterium. *Int J Syst Evol Microbiol* 60:1125–1129
86. Kodama Y, Stiknowati LI, Ueki A, Ueki K, Watanabe K (2008) *Thalassospira tepidiphila* sp. nov., a polycyclic aromatic hydrocarbon-degrading bacterium isolated from seawater. *Int J Syst Evol Microbiol* 58:711–715
87. Jiménez N, Viñas M, Guiu-Aragónés C, Bayona J, Albaigés J, Solanas A (2011) Polyphasic approach for assessing changes in an autochthonous marine bacterial community in the presence of *Prestige* fuel oil and its biodegradation potential. *Appl Microbiol Biotechnol* 91:823–834
88. Wang F, Xu M, Xiao X (2011) Isolation and characterization of alkane hydroxylases from a metagenomic library of Pacific deep-sea sediment In: De Bruijn, F (ed.) Handbook of molecular microbial ecology II: metagenomics in different habitats, vol. 2. Wiley-Blackwell, pp. 475–479
89. Wang L, Wang W, Lai Q, Shao Z (2010) Gene diversity of CYP153A and AlkB alkane hydroxylases in oil degrading bacteria isolated from the Atlantic Ocean. *Environ Microbiol* 12:1230–1242
90. Tourova T, Nazina T, Mikhailova E, Rodionova T, Ekimov A, Mashukova A, Poltarau A (2008) *alkB* homologs in thermophilic bacteria of the genus *Geobacillus*. *Mol Biol* 42:217–226
91. Yergeau E, Sanschagrin S, Beaumier D, Greer CW (2012) Metagenomic analysis of the bioremediation of diesel-contaminated Canadian high arctic soils. *PLoS One* 7:e30058
92. Wang Y, Yu M, Austin B, Zhang X-H (2012) *Oleispira lenta* sp. nov., a novel marine bacterium isolated from Yellow sea coastal seawater in Qingdao, China. Antonie van Leeuwenhoek in press
93. Coulon F, McKew BA, Osborn AM, McGenity TJ, Timmis KN (2007) Effects of temperature and biostimulation on oil-degrading microbial communities in temperate estuarine waters. *Environ Microbiol* 9:177–186
94. Golyshin PN, Ferrer M, Chernikova TN, Golyshina OV, Yakimov MM (2010) *Oleispira*. In: Timmis, KN (ed.) Handbook of hydrocarbon and lipid microbiology, vol. 3. Springer-Verlag Berlin Heidelberg, pp. 1755–1764

Online Resource 1 Hydrocarbon content and *alkB* PCR clone library information of coastal sediment samples from Ushuaia Bay

Sample		Hydrocarbon content ($\mu\text{g/g dws}$) ^b			Aliphatic diagnostic indices				<i>alkB</i> PCR clone libraries		
type	name ^a	TRA ^c	UCM ^d	TAH ^e	<i>n</i> C17/Pr ^f	<i>n</i> C18/Ph ^g	$\Sigma\text{LMW}/\Sigma\text{HMW}$ ^h	TRA/UCM	<i>N</i> ⁱ	<i>S</i> _{obs} ^j	Coverage
Ushuaia Bay sediments	EM06	38.78	308.4	347.2	1.32	1.13	na	0.13	50	14	0.90
	OR06	5.09	28.03	33.12	1.27	1.48	na	0.18	30	1	1
	OR07	8.74	68.26	77.00	1.27	1.34	na	0.13	46	12	0.87
	OR08	4.06	19.44	23.50	1.19	1.43	na	0.21	41	5	0.95
	OR08s	2.06	17.64	19.70	1.35	1.34	na	0.12	35	8	0.94
Crude oil	O	na	na	na	2.17	3.18	4.09	0.76	-	-	-
Experimental systems	expOR08-c	na	na	na	1.00	1.20	na	0.16	-	-	-
	expOR08-O	na	na	na	1.99	2.82	2.34	0.41	26	5	0.96
	expOR08-ON	na	na	na	0.02	0.20	0.78	0.12	25	3	0.96

na: not applicable; ^a sediment samples are named as site followed by the last two digits of the sampling year. Experimental systems are named as expOR08-c: control slurry (nor oil neither nutrients added), expOR08-O: oil-exposed slurry, expOR08-ON: oil-plus-nutrient amended slurry; ^b hydrocarbon content values are expressed as $\mu\text{g/g}$ dry weight sediment; ^c total resolved aliphatics; ^d aliphatic unresolved complex mixture; ^e total aliphatic hydrocarbons (TRA + UCM); ^f ratio between *n*-C17 alkane and pristane; ^g ratio between *n*-C18 alkane and phytane; ^h ratio between low and high molecular weight hydrocarbons; ⁱ number of sequences; ^j number of OTUs defined at 0.20 cutoff distance at the amino acid level.



Online Resource 2 Alignment of deduced AlkB sequences from clones obtained in this study and related sequences from public databases. Only one representative sequence per OTU is shown. The conserved histidine box II [1, 2] is indicated. The beginning of histidine box II in the sequences from this study is not included, as it was included in the forward primer [3]. Residues $\geq 50\%$ identical and similar are shaded in black and gray, respectively. CLUSTAL, MEGA5 and BioEdit softwares were used to build and shade the alignment.

1. Shanklin J, Whittle E, Fox BG (1994) Eight histidine residues are catalytically essential in a membrane-associated iron enzyme, stearyl-CoA desaturase, and are conserved in alkane hydroxylase and xylene monooxygenase. *Biochemistry* 33: 12787-12794
2. Smits TH, Rothlisberger M, Witholt B, van Beilen JB (1999) Molecular screening for alkane hydroxylase genes in Gram-negative and Gram-positive strains. *Environ Microbiol* 1: 307-317
3. Olivera NL, Nievas ML, Lozada M, del Prado G, Dionisi HM, Siñeriz F (2009) Isolation and characterization of biosurfactant-producing *Alcanivorax* strains: hydrocarbon uptake strategies and alkane hydroxylase gene analysis. *Res Microbiol* 160: 19-26

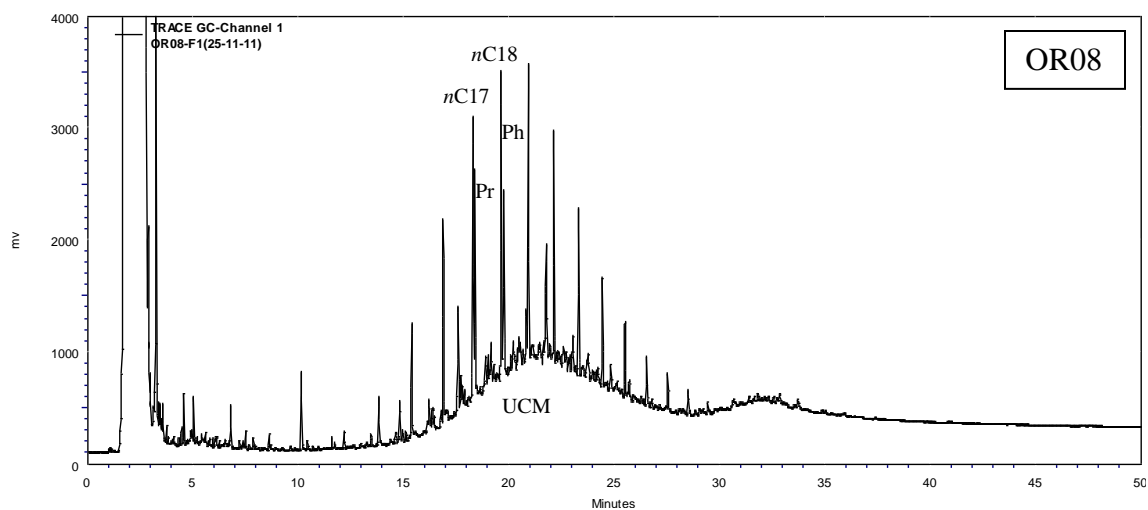
Online Resource 3 BLAST results of the *alkB* gene OTUs identified in this study. OTUs are ordered by relative abundances. For each OTU, the first BLAST match and the sequence from the closest isolated microorganism is indicated, with GenBank accession numbers in parentheses. Percent identity at the amino acid level is shown. *N*: number of sequences in the OTU; *n*: number of samples where the OTU was detected; samples correspond to: S (sediment), O (oil-exposed slurry), ON (oil-plus-nutrient amended slurry).

<i>OTU</i>	<i>N</i>	<i>n</i>	<i>sample</i>	<i>BLAST match</i>	<i>% Identity</i>	<i>e-value</i>
29	44	3	S	<i>alkB</i> Forest soil clone alkB22CB (HM441106)	67	3.00E-41
				<i>alkB</i> <i>Alcanivorax dieselolei</i> NO1A (AAT91719)	64	4.00E-41
8	38	2	S	<i>alkB</i> Antarctic sediment clone SRE-19 (ACT91224)	73	8.00E-45
				<i>alkB</i> <i>Ralstonia</i> sp. PT11 isolate 4 (ABA55801)	57	6.00E-35
24	26	2	S	<i>alkB</i> Arctic soil clone GO0VNXF07IJMIO (JF407716)	80	8.00E-55
				<i>alkB</i> <i>Gordonia</i> DSM 44995 (ACZ17528)	66	7.00E-41
	11	1	O	<i>alkB</i> Arctic soil clone GO0VNXF07IJMIO (JF407716)	89	4.00E-61
				<i>alkB</i> <i>Nocardia</i> sp. 3.2-VPr (AEH76914)	70	9.00E-43
6	7	1	S	<i>alkB</i> Soil clone OTU28 (ACB11552)	93	2.00E-62
				<i>alkB</i> <i>Marinobacter</i> sp. S17-4 (ACT31523)	85	3.00E-57
	2	1	O	<i>alkB</i> Soil clone OTU28 (ACB11552)	93	9.00E-63
				<i>alkB</i> <i>Marinobacter</i> sp. S17-4 (ACT31523)	85	2.00E-57
	12	1	ON	<i>alkB</i> France sediment clone cl3 (CCA29153)	96	6.00E-66
				<i>alkB</i> <i>Kordiimonas gwangyangensis</i> S20-13 clone 1 (ACJ22725)	91	6.00E-63
23	5	3	S; O	<i>alkB</i> <i>Rhodococcus</i> sp. 28/19 (ABD36392)	100	3.00E-67
	12	1	ON	<i>alkB</i> <i>Rhodococcus</i> sp. 28/19 (ABD36392)	99	2.00E-66
7	16	3	S;O	<i>alkB</i> Antarctic sediment clone SRE-19 (ACT91224)	67	1.00E-41
				<i>alkB</i> <i>Ralstonia</i> sp. PT11 isolate 4 (ABA55801)	62	2.00E-35
17	16	2	S	<i>alkB</i> Timor deep-sea sediment clone G17 (ACU43522)	70	1.00E-48
				<i>alkB</i> <i>Kordiimonas gwangyangensis</i> S20-13 clone 1 (ACJ22725)	58	1.00E-35

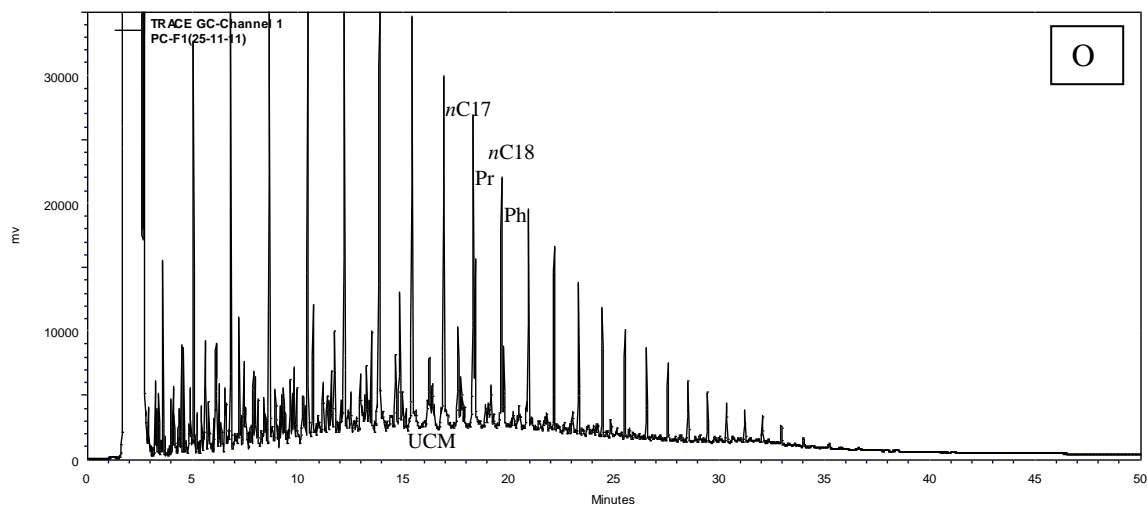
<i>OTU</i>	<i>N</i>	<i>n</i>	<i>sample</i>	<i>BLAST match</i>	<i>% Identity</i>	<i>e-value</i>
5	9	2	S	<i>alkB</i> <i>Alcanivorax hongdengensis</i> A-11-3 (ACD47155)	74	3.00E-47
1	6	1	O	<i>alkB</i> Salinisphaeraceae bacterium PC39 (ADE08461)	73	1.00E-47
2	5	1	S	<i>alkB</i> Antarctic sediment clone SRE 19 (ACT91224)	84	3.00E-55
				<i>alkB</i> <i>Alcanivorax borkumensis</i> s17-15 (ACJ22718)	62	8.00E-35
11	4	1	S	<i>alkB</i> <i>Alcanivorax dieselolei</i> S9-7 (ACJ22690)	62	5.00E-37
14	4	1	S	<i>alkB</i> Antarctic sediment clone SRE_F11 (ACT91241)	68	7.00E-44
				<i>alkB</i> <i>Ralstonia</i> sp. PT11 isolate 1 (ABA55798)	55	1.00E-32
18	4	1	S	<i>alkB</i> Timor deep-sea sediment clone G17 (ACU43522)	72	4.00E-41
				<i>alkB2</i> Salinisphaeraceae bacterium PC39 (ADE05603)	62	6.00E-30
21	4	2	S	<i>alkB</i> Forest soil clone SA08 (ABA55775)	67	6.00E-35
				<i>alkB</i> <i>Alcanivorax dieselolei</i> 6-D-6 (ACJ22764)	59	3.00E-29
13	3	1	S	<i>alkB</i> Timor deep-sea sediment clone F33 (ACU43488)	73	7.00E-47
				<i>alkB</i> <i>Ralstonia</i> sp. PT11 isolate 5 (ABA55802)	58	3.00E-33
20	3	1	S	<i>alkB</i> <i>Haliscomenobacter hydrossis</i> DSM 1100 (AEE51162)	50	2.00E-22
28	3	2	S	<i>alkB2</i> <i>Alcanivorax dieselolei</i> Qtet3 (ADN21388)	66	1.00E-45
32	3	1	S	<i>alkB</i> Soil clone alkB36mpn (ACZ64725)	73	6.00E-46
				<i>alkB</i> <i>Alcanivorax dieselolei</i> S9-7 (ACJ22690)	71	8.00E-45
12	2	1	S	<i>alkB</i> Pacific sediment metagenomic clone 9E7 (CAJ01370)	64	2.00E-37
				<i>alkB</i> <i>Ralstonia</i> sp. PT11 isolate 1 (ABA55798)	51	2.00E-26
26	2	1	S	<i>alkB</i> <i>Rhodococcus</i> sp. H1 (ACR56750)	82	2.00E-53
3	1	1	S	<i>alkB</i> Barley field soil clone alkG4-12 (ABB90688)	83	8.00E-56
				<i>alkB</i> <i>Ralstonia</i> sp. PT11 isolate 2 (ABA55799)	65	3.00E-36
4	1	1	S	<i>alkB</i> <i>Alcanivorax hongdengensis</i> A-11-3 (ACD47155)	71	6.00E-44
9	1	1	S	<i>alkB</i> Timor deep-sea sediment clone F33 (ACU43488)	70	3.00E-46
				<i>alkB</i> <i>Alcanivorax</i> sp. PA3 (ACD63080)	55	2.00E-31

<i>OTU</i>	<i>N</i>	<i>n</i>	<i>sample</i>	<i>BLAST match</i>	<i>% Identity</i>	<i>e-value</i>
10	1	1	S	<i>alkB</i> Antarctic sediment clone SRE-19 (ACT91224)	65	8.00E-40
				<i>Pseudomonas aeruginosa</i> SL1 (ACJ65013)	57	5.00E-28
15	1	1	ON	<i>alkB</i> Arctic soil clone GO0VNXF07IG6EW (JF408383)	75	2.00E-39
				<i>alkB</i> <i>Alcanivorax dieselolei</i> II-D-3 (ACJ22770)	69	1.00E-34
16	1	1	S	<i>alkB</i> Antarctic sediment clone SRE-19 (ACT91224)	73	2.00E-49
				<i>alkB</i> <i>Marinobacter</i> sp. S17-4 (ACT31523)	66	6.00E-39
19	1	1	S	<i>alkB</i> <i>Haliscomenobacter hydrossis</i> DSM 1100 (AEE51162)	54	1.00E-25
22	1	1	S	<i>alkB</i> Arctic soil clone GO0VNXF07IG6EW (JF408383)	69	2.00E-44
				<i>alkB</i> <i>Alcanivorax borkumensis</i> s17-15 (ACJ22718)	59	2.00E-35
25	1	1	S	<i>alkB</i> <i>Arthrobacter</i> sp. 7.19-Vpa (AEH76916)	96	4.00E-65
27	1	1	S	<i>alkB</i> <i>Alcanivorax venustensis</i> ISO4 (AAT91720)	71	9.00E-48
30	1	1	S	<i>alkB</i> Timor deep-sea sediment clone H6 (ACU43509)	72	1.00E-34
				<i>alkB1</i> <i>Salinisphaeraceae</i> bacterium PC39 (ADE05602)	59	3.00E-28
31	1	1	S	<i>alkB</i> Forest soil clone alkB15XB2 (ADO21745)	87	2.00E-59
				<i>alkB</i> <i>Acinetobacter venetianus</i> 1-D-2 (ACJ22758)	73	1.00E-49

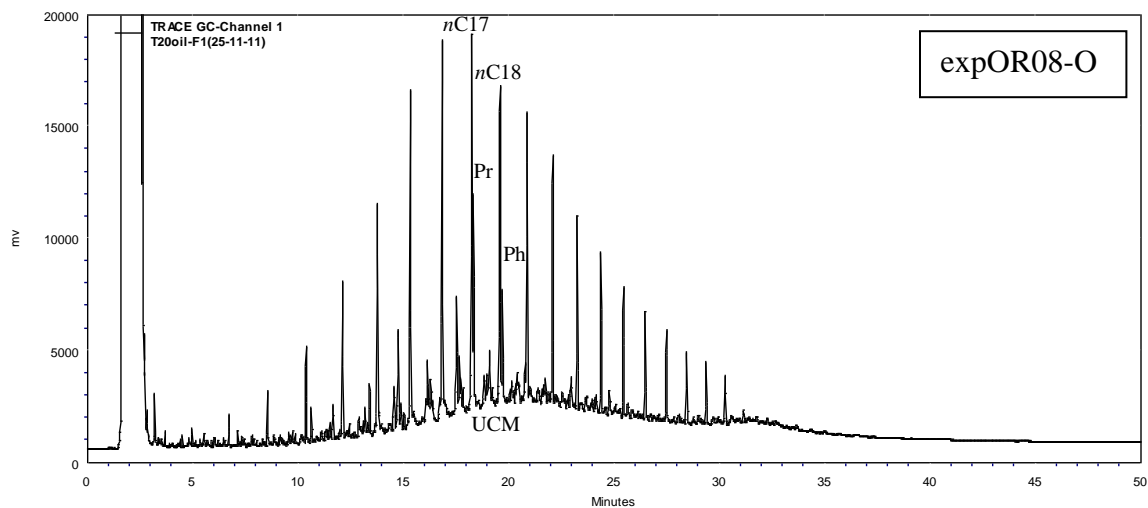
a



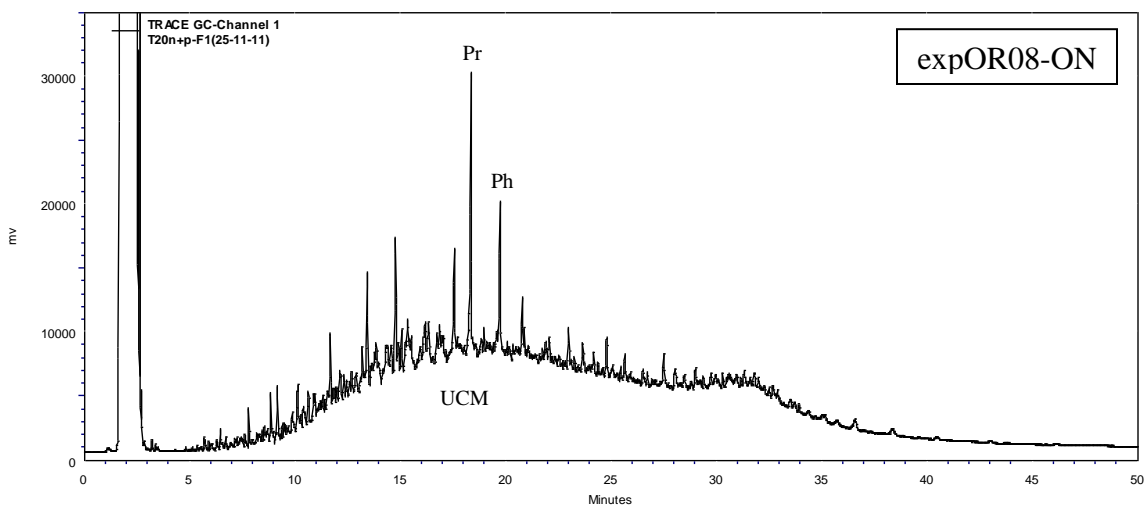
b



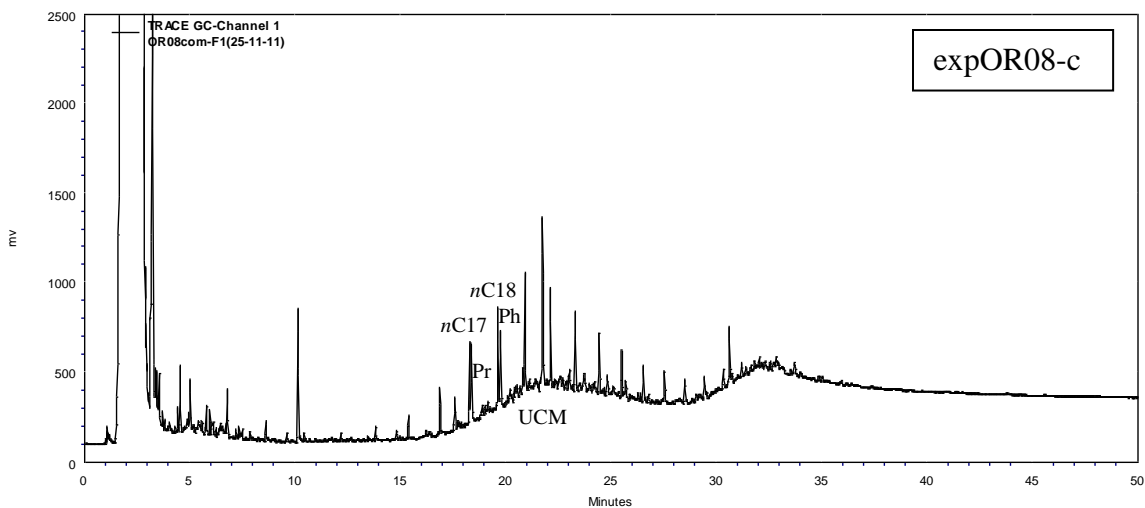
c



d



e



Online Resource 4 Profiles obtained by high resolution gas chromatography of the aliphatic hydrocarbon fraction of sediments and experimental systems. **a.** OR08 sediment sample from Ushuaia Bay. **b.** crude oil added to the experimental systems. **c.** oil-exposed slurry (expOR08-O) after twenty days of exposure. **d.** oil-plus-nutrient amended slurry (expOR08-ON) after twenty days of exposure. **e.** control slurry (expOR08-c, neither oil nor nutrients added) after twenty days. Representative compound abbreviations are indicated above the corresponding peaks. UCM: unresolved complex mixture