

1 *Evidence of ocean warming in Uruguay's fisheries landings: The mean*  
2 *temperature of the catch approach*

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12 **Running page head:** Uruguayan fisheries in a changing climate

13 **Abstract**

14 Distribution, abundance and life history traits of marine fish and invertebrates are increasingly  
15 affected by ocean warming. Consequently, landings of traditional fisheries and their relative species  
16 composition could potentially be modified. The mean temperature of the catch (MTC) concept,  
17 which refers to the average inferred temperature preference of exploited species weighted by their  
18 annual catch for a given area, was applied to Uruguay's industrial fisheries. This approach allowed  
19 assessing evidence of ocean warming in long-term Uruguayan landings (1973-2017), which were  
20 mostly obtained from a major marine warming hotspot. Results showed a marked shift in MTC  
21 through time, with the first 10-15 years characterized by a decreasing trend, but subsequently  
22 increasing steadily over time. Long-term effects of ocean warming have led to a shift from cool-  
23 water to warm-water species in the relative representation of local landings. A significant and  
24 consistent association between sea surface temperature and MTC increase was observed, even  
25 when accounting for other drivers. This study provides the first quantitative evidence supporting  
26 that ocean warming has been increasingly affecting Uruguayan industrial fisheries during the past  
27 decades and call for an urgent need to consider environmental changes to appropriately manage  
28 stocks, particularly those shared with neighboring countries.

29 **Keywords:** climate change; ocean warming; fisheries; mean temperature of the catch; Uruguay

30

## 1 *Introduction*

2 Ocean warming has profound and well documented impacts on fisheries (Perry et al. 2005, Hoegh-  
3 Guldberg & Bruno 2010, Free et al. 2019). The economy and the livelihood of communities who  
4 depend on marine resources for food security are being increasingly threatened by rising ocean  
5 temperature (Allison et al. 2009, Hobday et al. 2016). Marine species respond to ocean warming by  
6 altering their depth (Dulvy et al. 2008) and/or latitudinal range (Cheung et al. 2009, Pinsky et al.  
7 2013), mostly shifting their distributional range poleward (Madin et al. 2012, Pecl et al. 2017). On a  
8 global scale, ocean warming has been inhomogeneous. Hobday & Pecl (2014) identified several  
9 marine hotspots, which are regions where the sea surface temperature (SST) has increased most  
10 rapidly over the period 1950-2000 and are projected to continue changing at a faster rate than the  
11 global average. As marine species respond to ocean warming (Pinsky et al. 2013), these hotspots  
12 could affect the distribution, abundance and life history traits of fish (Pauly & Cheung 2018) and  
13 invertebrates (McLachlan & Defeo 2018), thus affecting catch levels and fisheries targeting  
14 traditional resources. This situation is aggravated by the fact that marine fisheries constitute  
15 complex systems that are also influenced by human-induced changes such as exploitation history  
16 and management tools (Österblom et al. 2017). Thus, disentangling exploitation patterns from  
17 climate change effects represents a major challenge due to the interconnected nature of both  
18 drivers (Alheit et al. 2004).

19 Several approaches have been applied to understand and project impacts of climate change in  
20 fisheries (Barrange et al. 2010), although most of them fell short on critical aspects such as the role  
21 of management measures and resource use in aggravating or mitigating impacts (but see Costello  
22 2018, Gaines et al. 2018). Cheung et al. (2013) developed an innovative and simple approach for  
23 assessing the signature of ocean warming in fisheries catches, based on the “mean temperature of  
24 the catch” (MTC), which refers to the average inferred temperature preference of exploited species  
25 weighted by their annual catch for a given area (e.g. Large Marine Ecosystems (LMEs) or  
26 Economic Exclusive Zones). The underlying ecological theory is that thermal preferences of marine  
27 species can be described by a Gaussian function with an optimal inferred temperature. Thus, an  
28 increase in MTC would indicate more warm-water species in the catch and/or a decrease in the  
29 amount of cool-water species. Recent analyses in the Mediterranean Sea (Tsikliras & Stergiou 2014)  
30 and China’s seas (Liang et al. 2018) supported global observations about the increase in the MTC at  
31 a regional scale. Other studies used fishery-independent data to estimate and compare MTC results  
32 from those obtained using landings data (Keskin & Pauly 2014, Tsikliras et al. 2015), or took into  
33 account habitat characteristics and indirect measures of fishing effects (Maharaj et al. 2018, Tsikliras  
34 et al. 2019).

1 However, if not properly addressed, the MTC (based only on catch data) could lead to misleading  
2 conclusions. MTC analyses should also consider the effects of fishing effort, markets and industry  
3 portfolio diversifications, which can influence the relative composition of the catch.

4 In the Southwestern Atlantic Ocean (SAO), the continental shelf of southern Brazil, Uruguay and  
5 northern Argentina embraces one of the largest marine hotspots (Hobday & Pecl 2014). This  
6 oceanographic complex region is characterized by high seasonal variations modulated by the  
7 confluence of two water masses with different thermohaline characteristics (Matano et al. 2010,  
8 Piola et al. 2018). Subtropical warm waters are carried southward by the Brazil Current along the  
9 shelf break, while subantarctic cold waters are advected northwards by the Malvinas Current (Fig.  
10 1). Their confluence is marked by an array of strongly contrasting water types (Gordon 1984) that  
11 define a very energetic and complex region (Fig. 1). Additionally, the Río de la Plata estuary, the  
12 major freshwater discharge in the area (average discharge of 22,000 m<sup>3</sup> s<sup>-1</sup>; Guerrero et al. 1997),  
13 injects buoyancy and nutrients contributing to system complexity. Furthermore, El Niño Southern  
14 Oscillation (ENSO) has a pronounced effect in the area by altering precipitation regimes, which in  
15 turn modulate the Río de la Plata discharge magnitude (Piola et al. 2005, 2018). The effect of  
16 ENSO on SST anomalies (SSTA) in the SAO is less evident, though lagged northward extensions  
17 of cold water were related with ENSO cold phase and the opposite for ENSO warm phase  
18 (Colberg et al. 2004 and references therein; Ortega & Martínez 2007). Critically, but less  
19 understood, is the atmosphere-ocean coupling during ENSO events (Ortega & Martínez 2007,  
20 Barreiro 2009). This dynamic oceanographic setting, with a high degree of seasonal and interannual  
21 variations, determines a complex horizontal and vertical structure that provides favorable  
22 environmental conditions for high primary productivity and also for spawning and breeding of fish  
23 and invertebrate populations (Ciotti et al. 1995, Acha et al. 2004, Jaureguizar et al. 2016).

24 Several studies at a global scale have projected significant mid and long-term catch losses in the  
25 SAO (Cheung et al. 2009, Blanchard et al. 2017, Cheung et al. 2018), seriously threatening an  
26 already weakened Uruguayan and Argentinean industrial fishery sector (Villasante et al. 2016,  
27 Gianelli & Defeo 2017). The uncertainty of these global model projections at the regional scale is  
28 aggravated by the rapid changes in oceanographic conditions occurring in this hotspot. However,  
29 despite the importance and scope of this phenomenon for fisheries and dependent communities, to  
30 date there are few studies of the impact of climate change on fisheries in this region (Bertrand et al.  
31 2018). Even more uncertain is the understanding of how environmental change and management  
32 policies could interact in aggravating or mitigating climate change impacts (Barange et al. 2010,  
33 Gaines et al. 2018). Additionally, the scarcity of fishery-independent data renders it difficult to  
34 disentangle multiple, potentially confounding factors, such as changes in fishing pressure, market  
35 effects and consumer preferences, from potential climate drivers (Madin et al. 2012, Ortega et al.  
36 2012). Thus, there is an increasing need to provide meaningful evidences of climate change impacts

1 on major fishery resources, with the explicit inclusion of the human dimension. Ideally, this will  
2 assist fishery industry and managers to develop adaptation and mitigation plans for regional  
3 fisheries in a changing climate.

4 In this paper we applied the MTC approach to long-term Uruguayan industrial fishery landings  
5 (1973-2017), which are mostly obtained from one of the most energetic and largest marine  
6 hotspots. An analysis of a variant of the MTC index was also carried out to account for the effect  
7 of the dominant species on the composition of the catch. Results were complemented with detailed  
8 Uruguayan and Argentinean fishery landings for the period 1996-2017, provided by the binational  
9 commission responsible for managing straddling fish stocks in the study area. Finally, the MTC  
10 index was modeled to assess the relative contribution of environmental variables and historical  
11 exploitation patterns of the Uruguayan fishing fleet.

12

13

## 1 **Methods**

### 2 *The Uruguayan industrial fleet*

3 The Uruguayan industrial fleet operates on the Río de la Plata estuary (the widest estuary in the  
4 world) and in the Argentinean-Uruguayan Common Fishing Zone (AUCFZ). This zone is  
5 determined by two arcs of circumference of 200 nm radius, centered at Punta del Este (Uruguay)  
6 and Punta Rasa (Argentina) (Fig. 1). The area of the AUCFZ is 218,718 km<sup>2</sup> and includes almost  
7 entirely the Uruguayan Exclusive Economic Zone (EEZ). The Joint Technical Commission of the  
8 Uruguayan-Argentinean Maritime Front (CTMFM for its acronym in Spanish) is the regional  
9 management and governance body responsible for compiling fishery statistics from both countries,  
10 coordinating research plans and performing joint stock assessments for most commercially  
11 important resources within the AUCFZ (see <http://ctmfm.org/>).

12 Industrial vessels operate, by law, beyond 7 nm from shore. Most Uruguayan vessels are bottom  
13 trawlers. Longlines and pots are used in a lesser extent to target pelagic and benthic resources,  
14 respectively. The industrial fishing sector has traditionally represented >95% of total Uruguayan  
15 landings and currently consists of 50–55 vessels (15–60 m length) with a mean of 360 Gross  
16 Tonnage. More than 50 species are caught by the Uruguayan industrial fishing fleet. However,  
17 landings of Argentine hake (*Merluccius hubbsi*), whitemouth croaker (*Micropogonias furnieri*) and  
18 striped weakfish (*Cynoscion guatucupa*) historically represented more than 85% of total landings  
19 (Gianelli & Defeo 2017). Landings have drastically fluctuated in the long-term following the phases  
20 described by Gianelli & Defeo (2017): development, expansion, stabilization-diversification and a  
21 last phase of declining yields and market contraction.

### 22 *Fisheries statistics and climatic data*

23 Uruguayan landings within the AUCFZ<sup>1</sup> and the Río de la Plata estuary for the period 1973-2017  
24 were extracted from the National Direction of Aquatic Resources (DINARA for its acronym in  
25 Spanish) reports (compiled in Gianelli & Defeo 2017: Fig. 2a). The analysis was conducted at the  
26 species level or at a higher taxonomic level in which we were confident of the predominance of a  
27 single species (e.g. *Mugil* spp. and *Squatina* spp.). This resulted in 21 species or genera representing  
28 95% of total historical landings. The remaining landings reported in official records were aggregated  
29 into higher taxonomic resolution levels and were not considered in this paper. An additional MTC  
30 variant (Table 1) and analysis were performed by excluding species that comprised >20% of  
31 landings, following Maharaj et al. (2018). This approach allows MTC to be more representative of  
32 species assemblages, avoiding trends that are disproportionately affected by few dominant species.

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<sup>1</sup> Species targeted outside the AUCFZ were excluded from the analysis (i.e. high migratory tunas and billfishes).

1 The temperature preference for each species was obtained from two different sources: (1) Cheung  
2 et al. (2013), which inferred temperature preferences from the overlap of species distribution range  
3 maps with SST data (averaged of 1970 – 2000; see Jones et al. 2012 and Supplementary Online  
4 Material of Chueng et al. 2013 for methods); and (2) FishBase and Sea Life temperature preference  
5 data (Froese & Pauly 2011), based on the Aquamaps (www.aquamaps.org) approach, which uses  
6 descriptors of species relationships with environmental variables and expert judgment to predict  
7 spatial distributions from publicly available global occurrence databases. In order to assess if the  
8 data provided by Cheung et al. (2013) and FishBase were within the physiological temperature  
9 preference ranges derived from *in situ* samples in the region, we reviewed published papers and  
10 local grey literature. MTC was computed as the average of the temperature preference of exploited  
11 fishes and invertebrates species weighted by their annual catch (Cheung et al. 2013):

$$12 \quad MTC_{yr} = \frac{\sum_i^n T_i C_{i, yr}}{\sum_i^n C_{i, yr}}$$

13 where  $C_{i, yr}$  is the catch of species  $i$  in year  $yr$ ,  $T_i$  is the median temperature preference of species  $i$   
14 (described by a Gaussian function) and  $n$  is the total number of species reported from landings  
15 within the AUCFZ. For visualization purposes, the MTC anomaly (relative to the 1973-2017 mean)  
16 is shown hereafter. The cumulative sum of annual MTC anomaly (MTCcs) was used to detect  
17 sustained shifts in the MTC index, marked by changes in slope of the cumulative sum plot (Fiedler  
18 2002 and references therein).

19  
20 SST and SSTA for the grid delimited by the coordinates 34.5-39° S and 52-58.5° W were obtained  
21 from the IRI/LDEO Climate Data Library (ridl.ldeo.columbia.edu 2018, Kalnay et al. 1996,  
22 Reynolds et al. 2002). SST and SSTA were calculated by averaging these variables over 4° x 3° grid  
23 cells of the SAO shelf and the adjacent oceanic region. The cumulative sum of annual mean SSTA  
24 (SSTAcs) was used to detect sustained shifts in climate, marked by changes in slope of the  
25 cumulative sum plot (Fiedler 2002).

### 26 *Temporal pattern in the MTC and modeling*

27 Segmented regression analyses were used to assess possible shifts in the MTC through time and to  
28 detect the breakpoints (year) when MTC changes occur. This procedure was applied to the MTC  
29 index considering all species and also excluding the dominant ones (Argentine hake and  
30 whitemouth croaker, Fig. 2a). Multiple linear regression models were also used to explore the  
31 relationship between the MTC index, the composition of the Uruguayan fishing fleet for each year  
32 and environmental effects (without and with time-lag up to three years). Environmental variables  
33 included SSTA and the Río de la Plata estuary discharge based on data reported by Borús et al.  
34 (2017). On the other hand, the utilization of different fishing métiers and strategies could stabilize  
35 fluctuations in fishing catches (Cline et al. 2017), while maintaining overall efficiency in utilization  
36 of fishery resources and ecosystem functioning (Zhou et al. 2010). Thus, based on vessels

1 characteristics, targeted species and operation areas, an annual diversity index for the Uruguayan  
2 fleet (Simpson index) was estimated using data provided by Marín (2016) and DINARA reports  
3 (Fig. 2b). Variations in the Simpson index could account for major shifts in the diversification  
4 portfolio, switch in target species or specific fishery closures from one year to another.

5  
6 Grounded on the provided background we established our main assumptions regarding the  
7 interaction between the MTC index and environmental and fishing effects, as follows: (1) due to  
8 the rapid changes in SST observed in this region, a clear sign of ocean warming in Uruguayan  
9 landings should be expected; (2) freshwater discharge anomalies could induce variations in the  
10 relative proportion of freshwater, brackish and marine species, leading to MTC changes; (3) fleet  
11 composition could be driven by market effects and portfolio diversification strategies, thus  
12 affecting catch composition and masking the effects of environmental change.

13  
14 Prior to model fitting, all variables were standardized (z-scores). In the case of nested models, the  
15 selection was performed based on likelihood ratio tests and AIC for small samples. The final model  
16 was checked for temporal autocorrelation in the residuals (see Fig. S2 in the Supplement). All  
17 statistical analyses were performed using R version 3.5.2 (R Core Team 2018). It should be noted  
18 that other potential effects of ENSO oscillations on MTC variations were analyzed through several  
19 approaches without any concluding result (see Supplement). Therefore, we do not further discuss  
20 ENSO effects hereafter, but focus in its most obvious manifestations: the Río de la Plata discharge  
21 and SSTA (both directly and indirectly affected by ENSO conditions).

22  
23 For the period 1996-2017, the CTMFM published landings data with detailed taxonomic resolution  
24 (28 species for Uruguay and 25 species for Argentina: Table 1). Thus, we re-estimated MTC for the  
25 period 1996-2017 based on CTMFM landings and compared with the previously described MTC  
26 time series (21 species for Uruguay for the period 1973-2017) through correlation analysis and  
27 paired samples t-tests. Because the CTMFM also compiled Argentinean landings, we estimated  
28 MTC for the entire AUCFZ (Uruguayan plus Argentinean landings for the period 1996-2017: Table  
29 1) and compared with MTC estimated only for Uruguay.

## 1 Results

2 Segmented regression models showed a marked shift in MTC through time, being characterized by  
3 a decreasing trend from 1973 to 1985 and steady increase over time until 2017 (Fig. 3a). By  
4 removing the effect of the two dominant species (*Merluccius hubbsi* and *Micropogonias furnieri*), the  
5 previously observed directional change was reinforced, but the breakpoint occurred later in the  
6 record, in 1999 (Fig. 3b).

7 The regional SSTA increased steadily since the mid-1990s with a clear dominance of positive values  
8 after 1999 (relative to 1973-2017 climatology: Fig. 4a), a trend particularly noticeable in the  
9 temporal analysis of SSTAcS (Fig. 4b). MTC anomaly showed a similar pattern to SSTA (Fig. 4a),  
10 particularly after year 2000. The cumulative sum plot of MTC anomaly and SSTA (Fig 4b) showed a  
11 consistent and positive correlation ( $r = 0.88$ ;  $p < 0.01$ ) and a concurrent shift in the cumulative sum  
12 slope of both variables. Related changes in both regimes suggest a close association between SST  
13 and MTC. This pattern also suggests that between the mid-1990s and 2000 an oceanographic  
14 change occurred in the region, which modulated the MTC index including all species (Fig. 4a) and  
15 also excluding dominant ones (Fig. 3b).

16 Multiple linear models showed a close association between MTC, SSTA and fleet diversity, with and  
17 without time-lags (Table 2). The Río de la Plata discharge anomaly was only statistically significant  
18 when considering 1-year time-lag (Table 2). Based on the different coefficients of determination  
19 obtained for models with and without time-lags (Table 2) and for the sake of parsimony, we  
20 retained the model without time-lag. This model showed that SST and the Simpson index were  
21 significant predictors of MTC (Table 3; Fig. 5a and 5b). The Río de la Plata discharge anomaly was  
22 marginally non-significant ( $p = 0.057$ ) and disregarded from the final model. MTC was positively  
23 related with SSTA and negatively related with the Simpson index (regression slopes: 0.43 and -0.44  
24 respectively). The latter implies that, as a particular type of fleet dominated in a given year targeting  
25 one or a small group of species (low portfolio diversification), MTC decreased. On the other hand,  
26 MTC increased when the fleet was more heterogeneous (high portfolio diversification) and  
27 provided a more balanced catch composition. The almost equal normalized coefficients (with  
28 opposite signs) of SSTA and Simpson index denoted that both variables are equally important as  
29 MTC predictors (Table 3). The final model explained a significant amount of the variance ( $R^2 =$   
30 0.63) and model residuals did not show temporal autocorrelation (Fig. S2).

31  
32 MTC estimated using CTMFM landings data for Uruguay (28 species) and the MTC based on  
33 Gianelli & Defeo (2017: 21 species) exhibited a significant and positive linear correlation ( $r = 0.98$ ;  
34  $p < 0.01$ ) for the period 1996-2016. Although the directional pattern remained the same, the more  
35 detailed taxonomic resolution for Uruguayan landings registered by the CTMFM influenced the  
36 MTC index slightly. The t-test for paired samples showed that MTC based on CTMFM was slightly



1 but significantly lower for the analyzed period ( $t = -9.26$ ,  $p < 0.001$ ). Additionally, the MTC based  
2 on Uruguayan and Argentinean landings (total landings for the AUCFZ) showed a significant linear  
3 increase through time ( $R^2 = 0.77$ ,  $p < 0.001$ , Fig. 6a), consistently with the MTC pattern observed  
4 only for Uruguay for the 1973-2017 period (Fig. 6b).

## 5 **Discussion**

6 This study provides the first quantitative evidence supporting that ocean warming has been  
7 increasingly affecting Uruguayan industrial fisheries during the past decades. Long-term effects of  
8 climate change have led to a shift from cool-water to warm-water species in the relative  
9 representation of Uruguayan landings. Furthermore, a significant and consistent association  
10 between SST and MTC increase was observed, even when accounting for other drivers such as the  
11 composition of the fleet and the influence of freshwater input with large interannual variability.  
12 MTC, which symbolizes a sort of thermometer of annual landings, represents a first attempt to  
13 address climate change effects on fisheries in the SAO.

14 Our results suggest that the AUCFZ is a very vulnerable region to a warming scenario and provide  
15 evidence that the SAO system and its biota are responding consistently with expectations under  
16 ocean warming (Bertrand et al. 2018). The shift from a cold to a warm phase during the 1990s  
17 (Ortega et al. 2016), as well as the increase in speed and frequency of onshore (southeast) winds  
18 (Escobar et al. 2004, Bischoff 2005), could enhance the advection of warm waters from the Brazil  
19 Current to the southern Brazil and Uruguayan shelf. Indeed, Ortega et al. (2016) showed a  
20 poleward shift of the warm water front ( $20^{\circ}$  C; a proxy of the front of tropical waters) of the order  
21 of  $9 \text{ km yr}^{-1}$ , which was positively correlated with the increase in SSTA, suggesting that ocean  
22 warming in this region may be related to changes in broader atmosphere and ocean circulation  
23 patterns. A similar southward displacement of the Brazil-Malvinas Confluence has been reported  
24 based on satellite and in-situ observations (Goni et al. 2011, Lumpkin & Garzoli 2011) and realistic  
25 numerical simulations (Combes & Matano 2014). Numerical simulations indicate that both the  
26 Brazil and the Malvinas currents play a significant role on the intensity of the shelf circulation  
27 (Palma et al. 2008, Matano et al. 2010). Thus, shifts in the Confluence may lead to significant  
28 changes in the circulation of the AUCFZ. Although the marine species that inhabit this region must  
29 be adapted and resilient to these oscillations in oceanographic conditions, the sustained increase in  
30 SSTA could lead to unprecedented changes in the composition and structure of ecological  
31 assemblages, including extensions/contractions of their geographical range (Horta e Costa et al.  
32 2014). This long-term process may be preceded by an increase in the frequency of records of  
33 marine species with tropical affinities (Demicheli et al. 2006, Segura et al. 2009, Martínez et al. 2009,  
34 Izzo et al. 2010) and changes in the prevalence of cold water species towards warm/temperate  
35 water species (Schoeman et al. 2014; Lercari et al. 2018). These results taken together support the  
36 hypothesis of an increase in MTC in this subtropical region (Cheung et al. 2013) that aggravated the

1 crisis of the Uruguayan fishery sector, which is mainly based on the Argentine hake fishery, a cool-  
2 water affinity species.

3 The geographical region analyzed here represents an ecotone between the LME of South Brazil and  
4 Patagonian Shelf. Transitional characteristics, both in oceanography and biodiversity, place the  
5 AUCFZ as a relevant system to assess the effects of climate change on the marine realm. Cheung et  
6 al. (2013) showed that the MTC in the South Brazil LME has been fluctuating in the past decades,  
7 whereas MTC in the Patagonian Shelf LME has been decreasing. Two main factors impact the  
8 environmental conditions in the AUCFZ. Firstly, variations in the along-shore wind stress modulate  
9 seasonal oscillations and the interannual distribution of the Río de la Plata plume (Simionato et al.  
10 2001, Piola et al. 2005, Palma et al. 2008, Matano et al. 2014). In addition, the Subtropical Shelf  
11 Front (STSF, Piola et al. 2000) marks an intense subsurface transition between warm-salty  
12 subtropical shelf waters and cold-fresh subantarctic shelf waters. Both features exert a significant  
13 influence on the distribution of nutrients, suspended matter, zooplankton and ichthyoplankton in  
14 the AUCFZ (Ciotti et al. 1995, Muelbert et al. 2008), and seem to delineate the transition between  
15 the Patagonian Shelf and South Brazil LMEs. At interannual and longer time scales, the meridional  
16 fluctuations of the basin-scale wind fields that modulate the location of the Brazil-Malvinas  
17 Confluence may also impact on: (1) variations in the distribution of Río de la Plata waters; (2) the  
18 location of the STSF; and therefore (3) variations in SST that are reported in this study. Such  
19 changes would have a significant impact on the AUCFZ and the neighboring shelf ecosystem,  
20 suggesting that species inhabiting the boundaries of these LMEs should be resilient to variations,  
21 whereas species with trailing range edges of their geographical distribution should be especially  
22 vulnerable to climate change. Thus, the limits between these two LMEs could be much more  
23 dynamic than previously thought, affecting the geographic extension both for subantarctic and  
24 subtropical species.

25 The removal of dominant species from the analysis showed a marked influence of Argentine hake  
26 (cool-water affinity) and whitemouth croaker (warm-water affinity) landings in modulating MTC  
27 trends. Reduced catch per unit of effort (Lorenzo & Defeo 2015) and landings levels (Gianelli &  
28 Defeo 2017) in Argentine hake for the past 25 years have been largely attributed to a failure in the  
29 management system. However, climate change could be partially responsible for exacerbating the  
30 observed trends. Due to the commercial relevance of the region and its sustained increase in SST, a  
31 contraction in the trailing range edge of Argentine hake may be expected, deserving urgent research  
32 efforts. Furthermore, recent reports from a nearby region, conclude that the reproductive success  
33 indexes of Argentine hake are negatively correlated with SST, particularly in autumn (late larval or  
34 early juvenile phase: Marrari et al. 2018). On the other hand, the wide thermal niche of whitemouth  
35 croaker and its affinity for warm waters determine that this species may be less sensitive to an  
36 increase in regional SST. Other factors such as the Río de la Plata discharge and coastal pollution

1 are more relevant for the biology of this species. Acha et al. (2012) have shown that high freshwater  
2 runoff events promote low retention of eggs and larvae, which would affect the reproductive  
3 success of this commercially important species. In addition, interannual variations of Río de la Plata  
4 discharge alter the position and distribution of frontal zones and of those species that use them for  
5 spawning, such as the whitemouth croaker (Jaureguizar et al. 2016). In this sense, it can be  
6 hypothesized that under high freshwater runoff conditions a greater spatial dispersion of  
7 whitemouth croaker will be observed and, therefore resulting in a lower catchability for the fishing  
8 fleet. The fact that the Río de la Plata discharge variations appear to be confined to the region  
9 close to the river mouth (Piola et al. 2008, Matano et al. 2014) could explain the marginal  
10 relationship between discharge and MTC (a community index). However, this predictor could be  
11 especially meaningful for brackish water species, such as the whitemouth croaker (Jaureguizar et al.  
12 2016).

13 Managers and decision makers have left aside the effects of climate change in fisheries to focus on  
14 traditional fisheries management and the pressing threat of overfishing. However, this study  
15 suggests that there is a need to consider the effects of environmental changes to properly manage  
16 stocks located in a dynamic warming hotspot, particularly those shared by neighboring countries.  
17 The consistent results in MTC increase using different landing databases for the entire AUCFZ  
18 during 1996-2016 (considering Uruguayan and Argentinean landings) call for a regional cooperation  
19 strategy. As the responsible governance body, the CTMFM will have to deal with stock  
20 distributions becoming less predictable and traditional fisheries declining or new fisheries emerging.  
21 Existing fisheries management and governance schemes are largely based on the premise that fish  
22 stocks distribution ranges remain relatively static through time (Gutiérrez & Defeo 2013). However,  
23 as already observed in several regions and predicted at a global scale (Pinsky et al. 2018), many fish  
24 stocks will likely shift across national and other political boundaries in coming decades. Thus,  
25 cooperative management will play a critical role for facing conflicts over newly or no longer shared  
26 resources due to the effects of climate change in this region (Villasante & Österblom 2015).

27 Our study leaves open several issues related to: (1) strategies to be explored by the Uruguayan  
28 government and/or its industrial fishery sector to cope with changing oceans; (2) the need to carry  
29 out regional studies by incorporating detailed long-term Argentine landings data and including  
30 fishery independent surveys to provide robust lines of evidence for the AUCFZ; (3) the need to  
31 manage fish stocks based on model projections, not merely on historical and static assumptions  
32 about the distribution and abundance of the species; and finally (4) assess the potential for regional  
33 multinational collaboration on fisheries and ocean sustainability issues. The later issue needs to be  
34 implemented on a scale large enough to capture the impacts of climate change on regional fisheries  
35 of the SAO.

## 36 **Acknowledgements**

1 We are grateful for the support provided by the Inter-American Institute for Global Change  
2 Research (grant CRN 3070), which is supported by the US National Science Foundation (Grant  
3 GEO - 474 1128040), and by Comisión Sectorial de Investigación Científica (CSIC Grupos ID 32).

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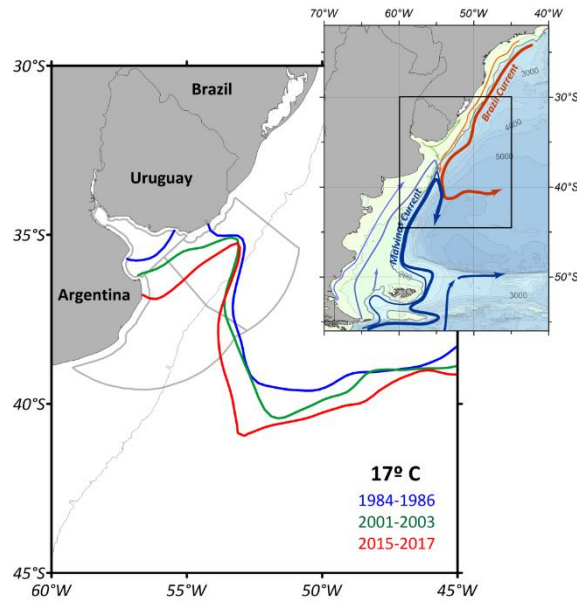
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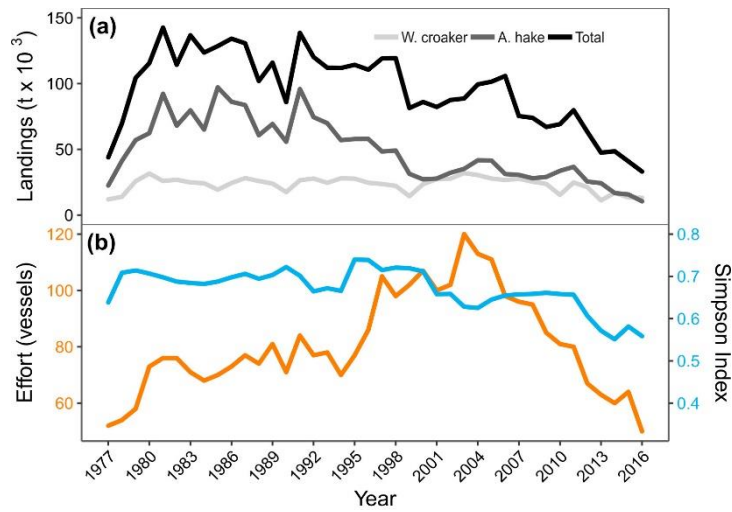
1 **Figures**



2

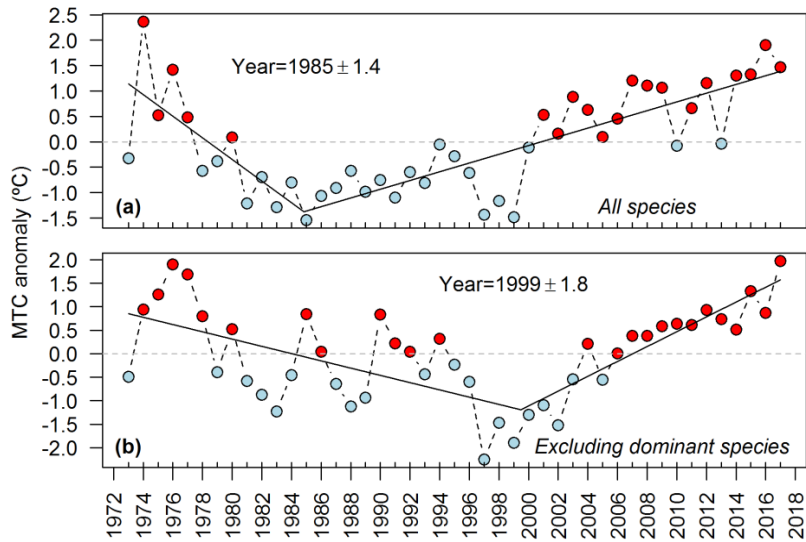
3 **Fig. 1.** Mean location of the 17 °C isotherm (proxy of the latitudinal shift of the warm waters front) derived from sea  
 4 surface temperature (Reynolds et al. 2007) averaged for the periods 1984-1986 (blue), 2001-2003 (green), 2015-2017 (red).  
 5 The Argentinean-Uruguayan Common Fishing Zone is delimited by the grey polygon. The 200 m isobath is highlighted in  
 6 gray. The upper panel shows a schematic representation of major ocean currents in the Southwest Atlantic Ocean, with  
 7 depth (m) indicated by background shading (adapted from Piola et al. 2018).

8



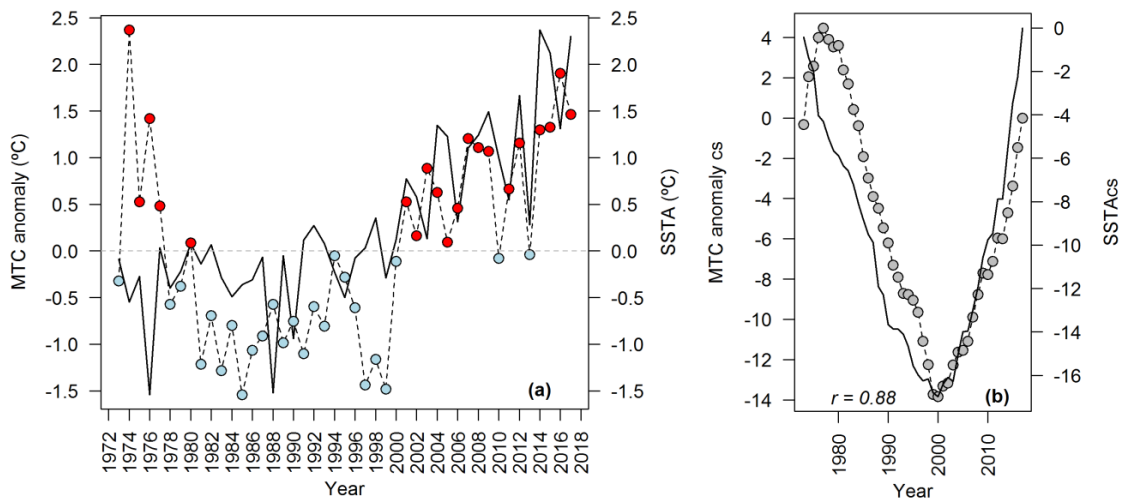
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10 **Fig. 2.** Long-term (1977-2016) variations in: (a) Uruguayan industrial fleet landings (total and, specific for Argentine hake  
 11 (A. hake) and whitemouth croaker (W. croaker)); (b) fishing effort, expressed as the number of vessels operating each  
 12 year (orange line) and Simpson's Index based on the composition of the fishing fleet (light blue line). Landings and  
 13 fishing effort time series were not linearly correlated ( $r = 0.22$ ,  $p = 0.17$ ).



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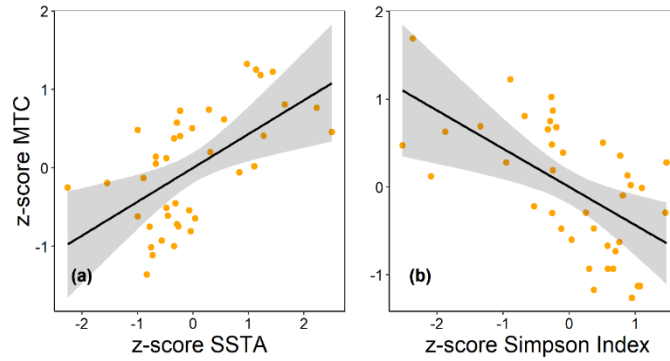
2 **Fig. 3.** Mean temperature of the catch (MTC; °C) anomaly for Uruguayan landings (1973-2017) in the AUFZ for (a) all  
 3 species landed ( $n = 21$ ; 95% of total landings); and (b) excluding *Merluccius hubbsi* and *Micropogonias furnieri*, which represent  
 4 nearly 70% of historical landings. The linear fits of segmented regressions are shown to highlight the increase in MTC  
 5 since (a) 1985 ( $R^2 = 0.62$ ); and (b) 1999 ( $R^2 = 0.52$ ). Red and light-blue dots represent positive and negatives anomalies in  
 6 MTC, respectively.



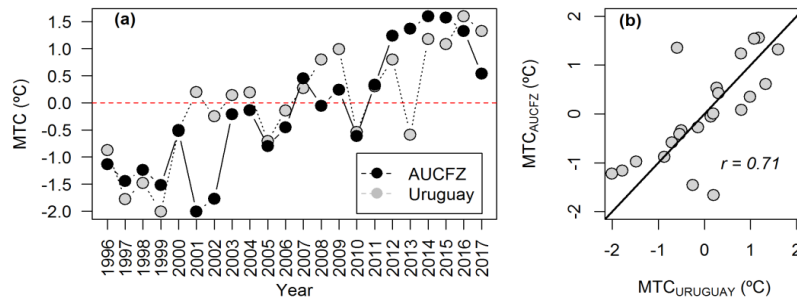
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8 **Fig. 4.** Long-term variations (1973-2017) in: (a) annual mean temperature of the catch (MTC; °C, dashed line) anomaly  
 9 and sea surface temperature anomaly (SSTA in °C; solid line). Red and light-blue dots represent positive and negatives  
 10 anomalies in MTC, respectively; (b) cumulative sum of mean temperature of the catch anomaly (MTC anomaly cs; grey  
 11 dots and dashed line) and sea surface temperature anomaly (SSTAcs; solid line).

12



**Fig. 5.** Mean partial effects of: (a) sea surface temperature anomaly (z-score SSTA); and (b) Simpson index applied to the fishing fleet each year (z-score Simpson Index), in the mean temperature of the catch (z-score MTC). Gray shadows indicate the 95% confidence interval. All variables were standardized to allow comparisons of slopes between predictors.



**Fig. 6.** (a) Mean temperature of the catch (MTC) anomaly for the Argentinean-Uruguayan Common Fishing Zone (°C; black dots) and Uruguay (°C; grey dots) based on CTMFM landings data. (b) Correlation between both time series shown in (a).

**Table 1.** MTC variants used in the study. Differences rely on different landing databases used for MTC calculation, the timeframe, the taxonomic resolution and the countries involved. CTMFM: Joint Technical Commission of the Uruguayan-Argentinean Maritime Front (CTMFM for its acronym in Spanish), AUCFZ: Argentinean-Uruguayan Common Fishing Zone.

MTC variants	Landings database used	Timeframe	Species considered
MTC	Gianelli & Defeo (2017)	1973-2017	21
MTC excluding main species	Gianelli & Defeo (2017)	1973-2017	19
MTC <sub>AUCFZ</sub> (Uruguay and Argentina)	CTMFM	1996-2017	28

**Table 2.** Results of linear models between MTC and environmental predictors, without and with time-lag (up to three years). For all models, the Simpson Index (z-score) for fishing fleet diversity was considered without time lag. S.E.: standard error. RdIP: Río de la Plata.

Variable	No time-lag		1-year time-lag		2-year time-lag		3-year time-lag	
	Estimate	S.E.	Estimate	S.E.	Estimate	S.E.	Estimate	S.E.
z-score SSTA	0.40**	0.14	0.25	0.15	0.43**	0.14	0.47**	0.15
z-score Simpson Index	-0.46**	0.11	-0.55***	0.14	-0.46**	0.13	-0.52***	0.13
z-score RdIP discharge	-0.19	0.09	-0.23*	0.109	-0.16	0.11	-0.08	0.11
Intercept	0.00	0.09	-0.01	0.101	0.02	0.10	0.060	0.11
	R <sup>2</sup> : 0.66		R <sup>2</sup> : 0.61		R <sup>2</sup> : 0.65		R <sup>2</sup> : 0.64	

**Table 3.** Linear model relating the Mean Temperature of the Catch (MTC) with the sea surface temperature anomaly (SSTA) and the fishing fleet diversity (Simpson Index). S.E.: standard error. Significance levels: \*\*\*<0.001; \*\*<0.01; \*<0.05.

Variable	Estimate	S.E.	IC95%	p-value
z-score SSTA	0.43	0.14	[0.14-0.72]	0.0043**
z-score Simpson Index	-0.44	0.14	[-0.72-0.15]	0.0040**
Intercept	0	0.10	[-0.19-0.19]	