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JSWC-D-14-00095R4 Maize and cover crop sequence in the Pampas: effect of fertilization and water stress on the fate of nitrogen
Research Paper
cover crop, maize; nitrate losses; nitrogen balance; tagged nitrogen
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Cover crops are well known for their positive effects on erosion processes, soil organic matter, soil physical properties, weed populations and nitrate leaching. In this work, we evaluated the fate of nitrogen (N) from fertilizer in maize (Zea mays) and then in ryegrass (Lolium multiflorum) as cover crop, in the conditions of the Argentine Pampas. To this end, a field experiment was carried out at the School of Agriculture, University of Buenos Aires, Argentina (34°36′ S, 58°29′ W). The design of the experiment was a factorial with three replications. We applied to maize 2 levels of N (0 and 140 kg N ha-1 (125 lb N ac-1) (ammonium nitrate target with 15N) and two levels of water (50 and 100% of crop evapotranspiration). 15N was determined in both the soil and plants. Maize plants and the soil organic fraction were the main sinks of fertilizer N, depending on the water treatment. The N from fertilizer remaining as nitrates in the soil (0 to 1.50 m [0 to 4.92 ft] depth) at maize harvest was 8% in plots subjected to water stress compared to 3% in the non-water stressed. Nitrogen losses due to volatilization were minor. Total N (soil and fertilizer) accumulated in ryegrass tissues plus nitrates remaining in the soil were higher in cover crop plots than in bare soil (130 vs. 51 kg N ha-1 [116 vs. 45.5 lb N ac-1]). The N in the soil organic matter originating from fertilizer significantly decreased between maize harvest and cover crop harvest. This soil organic N that originated from fertilizer mineralized at high rate (around 47% in 6 months), suggesting it was in more labile. This mineralized N can be subjected to potential losses during following months.
Answers to Reviewer comments line 80 should the average be referred as "regional" instead or "areal"? After an analysis about the differences between "regional" and "areal", we accepted to change to " weather characteristics were around regional average." line 96 replace the "2 levels of N and two levels of water" with "two levels of N and two levels of water" Replaced.

ne 262 replace "denitrificación" with "denitrification" Replaced, the word remained in Spanish.
ne 268 replace "data not showed" with "data not shown" Replaced. Thanks for the attention



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Buenos Aires, August 27, 2014

Drs. Jorge Delgado Editor Journal of Soil and Water Conservation

Dear Dr. Delgado

I am submitting you the manuscript entitled **Fate of nitrogen and nitrate leaching in maize and cover crop sequence in the Pampas**, asking you to consider it for publication in the **Journal of Soil and Water Conservation**, The authors are H. Rimski-Korsakov, M.S. Zubillaga, M.R. Landriscini and myself

Thank you very much for your attention

Sincerely

Raul S. Lavado Professor

Fate of nitrogen and nitrate leaching in maize and cover crop sequence in the Pampas

H. Rimski-Korsakov, M.S. Zubillaga, M.R. Landriscini and R.S. Lavado

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Maize and cover crop sequence in the Pampas: effect of fertilization and water stress on the fate of nitrogen

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4 Abstract: Cover crops are well known for their positive effects on erosion processes, soil 5 organic matter, soil physical properties, weed populations and nitrate leaching. In this work, 6 we evaluated the fate of nitrogen (N) from fertilizer in maize (Zea mays) and then in ryegrass 7 (Lolium multiflorum) as cover crop, in the conditions of the Argentine Pampas. To this end, a 8 field experiment was carried out at the School of Agriculture, University of Buenos Aires, 9 Argentina (34°36′ S, 58°29′ W). The design of the experiment was a factorial with three replications. We applied to maize 2 levels of N (0 and 140 kg N ha⁻¹ (125 lb N ac⁻¹) 10 11 (ammonium nitrate target with ¹⁵N) and two levels of water (50 and 100% of crop evapotranspiration). ¹⁵N was determined in both the soil and plants. Maize plants and the soil 12 13 organic fraction were the main sinks of fertilizer N, depending on the water treatment. The N 14 from fertilizer remaining as nitrates in the soil (0 to 1.50 m [0 to 4.92 ft] depth) at maize 15 harvest was 8% in plots subjected to water stress compared to 3% in the non-water stressed. 16 Nitrogen losses due to volatilization were minor. Total N (soil and fertilizer) accumulated in 17 ryegrass tissues plus nitrates remaining in the soil were higher in cover crop plots than in bare soil (130 vs. 51 kg N ha⁻¹ [116 vs. 45.5 lb N ac⁻¹]). The N in the soil organic matter 18 19 originating from fertilizer significantly decreased between maize harvest and cover crop 20 harvest. This soil organic N that originated from fertilizer mineralized at high rate (around 21 47% in 6 months), suggesting it was in more labile. This mineralized N can be subjected to 22 potential losses during following months.

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24 Key words: cover crop-maize-nitrate losses-nitrogen balance-tagged nitrogen

26 The recovery of fertilizer nitrogen (N) in crops total biomass does not generally 27 exceed 50%, but is even lower after the occurrence of different stresses during the crop 28 cycle (Ma et al. 1995; Gardner and Drinkwater 2009). Water stress, for instance, decreases 29 N uptake by a crop, irrespective of the source of nitrates (fertilizers, soil organic matter, 30 manure), and some nitrates remain in the soil after crop harvest. These nitrates are usually 31 called "residual nitrates". This not recovered N can go into the soil organic fraction, remain as 32 residual nitrates, or be lost by processes such as leaching, volatilization and denitrification. 33 Nitrogen losses have economic consequences and environmental risks. In some agricultural 34 systems, nitrate leaching cause contamination of aquifers (Spalding and Exner 1993). This 35 process is favored by prolonged periods with the soil free of live vegetation, during which 36 nitrates are not taken up by crops (Di and Cameron 2002). Thus, in most agricultural systems, 37 bare fallow is a period susceptible to suffer this kind of losses (Drury et al. 1996; Cambardella 38 et al. 1999). This is typical of production systems dominated by maize, which leaves long 39 fallow periods (Dinnes et al. 2002). If during fallow the soil is covered by growing vegetation, 40 i.e. cover crops (CC), or even weeds, the risk of nitrate leaching is usually reduced (Di and 41 Cameron 2002; Tonitto et al. 2006).

42 Cover crops are defined as crops seeded between cash crops which are not harvested, 43 not incorporated into the soil as green manures, and are not intended to be grazed, such as 44 annual forages (Glossary of Soil Science Terms). They are well known for their ability to 45 control erosion processes (Langdale et al. 1991; Kaspar et al. 2001), increase the soil organic 46 matter content and thus improve the soil physical properties (Reicosky and Forcella 1998; 47 Ding et al. 2006), and reduce weed populations (Williams et al. 2000). Cover crops are also 48 used to decrease the risk of contamination of aquifers by nitrates coming from fertilizers and 49 organic matter mineralization (Meisinger et al. 1991; Di and Cameron 2002; Dinnes et al., 50 2002; Meisinger and Delgado 2002; Tonitto et al. 2006). The effect of CC seems to occur as a two-stage process: i) the absorption of the nitrates of any origin present in the soil and ii) the
release of such N during the next crop cycle (Collins et al. 2007). Cover crops can also reduce
nitrate leaching because their high transpiration rate reduces water percolation (ThorupKristensen at al. 2003).

55 All these facts contribute to the general favorable consensus about the ability of CC to 56 reduce nitrate losses. However, it has also been reported that CC are not efficient or can even 57 increase nitrate leaching in the long term (Berntsen at al. 2006). Their incorporation within a 58 crop sequence can also show negative consequences, reducing the yield of the following crop 59 (Williams et al. 2000; Salmerón et al. 2011). This usually happens due to the competition 60 between the commercial crop and the CC for water and nutrients (Dabney et al. 2001). The 61 most common species used as CC belong to the family Poaceae and, to a lesser extent, to the 62 family Fabaceae. Among Poaceae, annual ryegrass (Lolium multiflorum) is one of the species 63 used as CC in winter fallows because it is characterized by its fast growth and good 64 adaptation to clayey, wet soils (Clark 2007). Our objective was to evaluate the fate of N from 65 fertilizer in aboveground biomass, soil organic matter, soil nitrates and its distribution in the soil 66 profile, and ammonia volatilization, in a maize production system subjected to water stress with 67 ryegrass as a CC, in the conditions of the Argentine Pampas.

68

69 Materials and Methods

A field experiment was carried out in the campus of the School of Agriculture, University of Buenos Aires, Argentina, located at 34°36′ S, 58°29′ W. The soil was a fine, illitic, thermic Vertic Argiudoll (USDA 2006), whose main characteristics in the plow layer, determined using standard techniques (Sparks et al. 1996), were: electrical conductivity 0.08 dS m⁻¹; pH in water 7.3; organic carbon (Walkley and Black) 1.75%; total Kjeldhal N 0.16%; and available phosphorus (Bray #1) 17.1 mg kg⁻¹ (ppm). The experiment was carried out in maize 76 (Zea mays cv FAUBA 209) followed by annual ryegrass (local commercial population of Lolium multiflorum) as CC. This was a short term experiment, like most studies using ¹⁵N 77 78 (Gardner and Drinkwater 2009). The main reason in present experiment was that water was 79 provided by irrigation and the drought was not related to climate variability. Besides, 80 temperature and other weather characteristics were around regional average. Rainfall and 81 other meteorological data during the length of the experiment were taken from a 82 Meteorological Station from the National Weather Service, located less than 500 m (1640 ft) 83 from the experiment location. The annual mean temperature registered in the Station was 16.6 84 °C. January is the hottest month (mean 24.9°C) and July the coldest (mean 11.0 °C). The 85 annual rainfall amounts 1146 mm, with no define seasonality but with great variations among 86 years.

87

88 Maize Cropping. Maize was seeded on November 15, 2005, following conventional tillage, 89 and harvested on May 16, 2006. The plot had maize as a previous crop. Maize was seeded 90 manually in 50 cm (19.7 in) rows with 20 cm (7.9 in) between plants for a total population of 100,000 plants ha⁻¹ (40,469 plants ac⁻¹). Each plot was 2.5 x 7.5 m (8.2 x 24.6 ft) and a 91 92 "microplot" (1.5 x 1.2 m [4.9 x 3.9 ft]) was installed in the center of each fertilized plot 93 (figure 1). To cover maize P requirement, two days before seeding, triple superphosphate (30 94 kg P ha⁻¹ [26.8 lb P ac⁻¹]) was broadcasted in all plots. Weeds, insects and diseases during the 95 crop cycle were controlled when necessary.

96 The experimental was a factorial (2x2) in randomized complete block design with three 97 replications. We applied to maize two levels of N and two levels of water: *N0:* not N 98 fertilizer was applied; *N140:* N fertilization doses of 140 kg N ha⁻¹ (125 lb N ac⁻¹) and -W: 99 subjected to water stress; +*W:* not subjected to water stress.

100

Nitrogen fertilization was carried out in the phenological stage V6 (Ritchie and

Hanway 1982), applying ammonium nitrate (140 kg N ha⁻¹ [125 lb N ac⁻¹]) manually in bands at 3 to 4 cm (1.2 to 1.6 in) depth. Within the microplots, the same N doses were applied using ¹⁵N tagged fertilizer (1.5% abundance; ¹⁵NH₄¹⁵NO₃) (figure 1). To regulate rain water input in the experiment, a plastic structure was installed on all plots. When rainfall was forecasted, this structure was unfolded and all spaces between the seeding lines were covered by the plastic strips. The structure had a slope toward a ditch, to remove the rain water from the plots. This simple mechanism limited most rain water from reaching the soil.

108 The potential evapotranspiration was calculated to estimate crop water requirement 109 (Penman 1948). A drop irrigation system was installed, which delivered 100% and 50% of the 110 calculated maize water requirement (621 mm [24.4 in and 310 mm [12.2 in]), respectively. 111 During the critical water period for maize (15 days before and after flowering), no water was 112 added to the -W plots. Clearly, the plastic cover did not prevent total flow of rainfall into the 113 soil, but its effect on the water treatments was minor, as tested before the experiment. Water 114 is the main variable factor affecting plant behavior between years. In the present experiment 115 water was provided by irrigation, reasonably solving the lack of environmental replications. 116 Besides, temperature, radiation and other weather factors that affect crop growth were around 117 average. Maize was manually harvested at physiological maturity. Plant material 118 (stems+leaves+cobs+husks and grain) were collected and dry weights measured after drying 119 in an oven at 60°C (140°F) to constant weight. Total N and the proportion of ¹⁵N were 120 determined in plant samples by the Kjeldahl method (Bremner and Mulvaney 1982) and optical 121 emission spectrometry (Fiedler and Proksch 1975), respectively. Soil samples at 0.30 m (0.98 122 ft) intervals up to 1.50 m (4.92 ft) depth were taken at maize seeding and harvest. Nitrate 123 concentration was determined in all soil samples by extraction with 2M KCl and distillation 124 with MgO and Devarda alloy (Keeney and Nelson 1982). The proportion of NO₃-¹⁵N was also 125 determined. Total N and the proportion of ¹⁵N were determined at harvest in top soil samples

by the above mentioned techniques. The N derived from fertilizer assimilated into the soil
organic fraction was calculated by subtracting the NO₃-¹⁵N from the total ¹⁵N. Ammonia
volatilization was determined following the chamber method proposed by Nommik (1973).
Ammonia was trapped in sponges with sulfuric acid-glycerol solution, subjected to airstreams
distillation and determined using a colorimeter (Sparks et al. 1996). Volatilization of
ammonia was determined 4, 15, 34 and 150 days after N fertilization.

The N derived from the fertilizer (NdfF) in the soil or plant compartment (%) and the N derived from the fertilizer (kg N ha⁻¹) were calculated using equation 1 and equation 2, respectively. The natural abundance of ¹⁵N in the commercial ammonium nitrate used to dilute the tagged ammonium nitrate was estimated as 0.366% (IAEA 2001). For the plant and soil compartments of the fertilized treatments, we used the natural abundance of ¹⁵N in the equivalent N0 treatment compartment. The N volatilized from the fertilizer was calculated using equation 3.

139

NdfF in plant or soil (%) =
$$\frac{\% \text{ atom }^{15}\text{N in plant or soil } - \% \text{ atom }^{15}\text{N natural in plant or soil}}{\% \text{ atom }^{15}\text{N in fertilizer } - \% \text{ atom }^{15}\text{N natural in fertilizer}}*100$$

140 (Equation 1)

NdfF in plant or soil $(kg N ha^{-1}) = NdfF$ in plant or soil (%) * N in plant or soil $(kg N ha^{-1}) / 100$ 141

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142 (Equation 2)
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NdfF volatilized = NH_4 -N volatilized in treatment N140 - NH_4 -N volatilized in treatment N0 (kg N ha⁻¹) 143

144 (Equation 3)

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146 The recovery of fertilizer in each compartment was estimated following equation 4.

	Fertilizer recovery (%) = $\frac{NdfF}{N_{fertilizer}}$ *100
147	
148	(Equation 4)
149	
150	Where:
151	NdfF = N derived from the fertilizer in the plant, soil organic N, soil NO ₃ -N, NH ₄ -N _{volatilized}
152	(kg N ha^{-1})
153	N _{fertilizer} = Nitrogen applied by fertilization (kg N ha^{-1})
154	
155	The total recovery of the fertilizer applied was calculated using equation 5.
156	
157	
	N fertilizer (100%) = NdfF plant (%) + NdfF volatilized (%) + NdfF nitrate (%) + NdfF soil organic + + NdfF unaccounted (%)
158	
159	(Equation 5)
160	
161	Where:
162	N $_{\text{fertilizer}} = N$ applied by fertilization (100%).
163	NdfF $_{plant}$ = N from the fertilizer taken by the whole plant at physiological maturity (%).
164	NdfF $_{volatilized} = N$ from the fertilizer volatilized (%).
165	NdfF $_{nitrate}$ = N from the fertilizer remaining as nitrates in the soil (%).
166	NdfF $_{\text{organic}}$ = N from the fertilizer remaining in the soil as organic N (%).
167	NdfF $_{unaccounted} = N$ from the fertilizer not detected in any of the compartments studied (%).
168	

169 *Cover Crop.* The plastic cover to control rainfall was removed after maize harvest and each 170 maize plot was divided in two. A new factorial design with three treatments was defined: 171 Nitrogen fertilization applied to maize (N0 and N140), water applied to maize (-W and +W) 172 and cover crop (with or without CC). Annual ryegrass (Lolium multiflorum) was seeded on May 20, 2006, by broadcasting the equivalent to 30 kg seeds ha⁻¹ (26.8 lb seeds ac⁻¹) in one 173 174 half of the plots (figure 1). The other half remained without vegetation, weeds were controlled 175 manually and pulled weeds were left in the plot. The rainfall and evapotranspiration during 176 the CC cycle amounted to 468 and 448 mm (18.4 and 17.6 in), respectively. During the last 30 days of the experiment, the rainfall was very high (277 mm [10.9 in]) and of great 177 178 intensity.

179 On November 5, 2006, the ryegrass was harvested by cutting 0.25 m² (2.7 ft²) in the center of all half plots, both where the maize was fertilized with ¹⁵N and where it was not. At 180 181 the same time, soil samples at 0.30 m (0.98 ft) intervals up to 1.50 m (4.92 ft) depth were 182 taken in the center of all half plots. Total N and the proportion of ¹⁵N were determined on the plant material and top soil samples; nitrates and the proportion of ¹⁵N were determined in the 183 184 soil samples taken to 150 cm depth. The above-described methods were followed in all cases. The apparent mineralization of the soil organic N coming from the fertilizer (15N) was 185 estimated as the difference between the soil organic ¹⁵N content at maize harvest and at the 186 187 end of the experiment (ryegrass harvest). The immobilization of the ¹⁵N released by the 188 decomposition of maize stubble during the study period in the soil organic fraction was not 189 measured. This could have affected the mineralization quantities estimated.

190 Statistics. The effects of the treatments were statistically analyzed using factorial ANOVA, 191 and the interactions between them and their effects were analyzed. When the interaction was 192 significant between treatments (W and N), comparisons were made between all combinations 193 of N and water application, whereas when the interaction was not significant, comparisons were made between treatments. The least significant difference (LSD) was used to
differentiate means. Differences were considered to be significant at the 5% (p=0.05)
probability level.

197

198 Results and Discussion

199 *Maize plants.* Maize vegetative and total aboveground biomass was significantly higher when 200 high N doses were applied and water supply was enough for crop requirements (N140+W) as 201 compared with the remainder treatments (tables 1 and 2). The other treatments showed no 202 significant differences between them. The water-N interaction agrees with that found by 203 Pandey et al. (2000) but not with that found by Gheysari et al. (2009), who found additive 204 effects on total maize biomass when N doses increased with different water conditions. It is 205 possible this behavior was caused by the lower water stress imposed in that experiment (15 to 206 30 % of non-stressed treatment), as compared with our water conditions (50%). On the other 207 hand, the grain yield increased significantly and additively when water supply and N were adequate. 208

209 Nitrogen concentration in the vegetative biomass (0.45%) and in grains (1.21%) was 210 not significantly different between treatments. Nitrogen accumulation in the total 211 aboveground biomass (vegetative biomass+grains) and grains alone was higher with high 212 fertilizer doses and without water stress (additive effects) (tables 1 and 2). Nitrogen coming 213 from the soil and located in total aboveground biomass was not affected by the treatments 214 applied and was on average 74% of the total N accumulated in that biomass (tables 1 and 2). 215 Treatment affected the accumulation of N coming from fertilizer in total aboveground 216 biomass, which was significantly higher when maize was not subjected to water stress (tables 217 1 and 2). The fertilizer recovery in total aerial biomass was then 24 vs. 46%, of the N applied 218 via fertilization, when maize was or not subjected to water stress, respectively.

219 Nitrogen in soil at maize harvest. The total content of soil nitrates (derived from the soil + 220 derived from fertilizer) at 0 to 1.50 m (0 to 4.92 ft) depth at maize harvest was significantly 221 higher in fertilized treatment than in non-fertilized plots (119 vs. 105 kg NO₃-N ha⁻¹ [106 vs. 94 lb NO₃-N ac⁻¹]; figure 2A). Plots subjected to water stress showed a trend (p=0.067) to a 222 223 higher concentration of nitrates than non-subjected to water stress plots (118 vs. 106 kg NO₃-N ha⁻¹ [105 vs. 95 lb NO₃-N ac⁻¹] respectively). Lower N uptake by maize and lack of nitrate 224 225 movement due to dry conditions can explain this behavior (Gheysari et al. 2009). The soil 226 derived nitrates (non-tagged nitrates, NO₃-¹⁴N) from 0 to 1.50 m (0 to 4.92 ft) presented no 227 effects caused by the treatments imposed. The N derived from fertilizer (tagged nitrates, NO₃-¹⁵N) in the water stress treatments was responsible for the trend shown in figure 2B. The 228 229 content of nitrates derived from the soil plus derived from fertilizer and the nitrates from 230 fertilizer was significantly higher only at 0 to 30 cm (0 to 11.8 in) depth in non-subjected to 231 water stress plots, without effects of N fertilization (N0 or N140). No differences in total 232 nitrates were found at the other soil depths studied.

The content of total soil organic N including the ¹⁵N derived from the fertilizer at 0 to 233 234 30 cm (0 to 11.8 in) depth was not affected by the treatments imposed. Conversely, the N 235 fraction in the soil organic matter derived from fertilizer was higher in the -W treatment than in the +W treatment (78 and 51 kg N ha⁻¹ [70 and 46 lb N ac⁻¹]), which means 56 and 37% of 236 237 the applied fertilizer, respectively (figure 3). Stressed plants, with low N recovery efficiency, 238 left a high concentration of N in the soil, which was immobilized by soil biota and mainly 239 accumulated in the soil organic fraction. Our data were relatively greater that those found by 240 Reddy and Reddy (1993) and Portela et al. (2006), who found 10 to 30% of the applied N in the 241 soil organic matter at harvest of maize not subjected to limiting conditions.

242 Volatilization during maize. Accumulated volatilization during the maize cycle was243 significantly higher when fertilizer was applied, with no differences between water

treatments. Volatilization in non- fertilized treatments was the natural volatilization from the soil (3.2 kg NH₃-N ha⁻¹ [2.9 lb NH₃-N ac⁻¹]) while fertilized treatments lost 7.1 kg NH₃-N ha⁻¹ (6.3 lb NH₃-N ac⁻¹) on average. Based on these data, we estimated that the N volatilized from fertilizer was 3.9 kg NH₃-N ha⁻¹ (3.5 lb NH₃-N ac⁻¹) during the maize cycle. These losses were lower than other data observed in the Pampas region (Palma et al. 1998) and could be attributed to the N source (ammonia nitrate), the depth fertilizer applied and/or the occurrence of adequate humidity in the soil.

251 Fate of N from fertilizer during maize cycle. Maize biomass was the main sink of N from 252 fertilizer in plants not subjected to water stress, being the soil organic fraction the second 253 most important sink: 47% and 37%, respectively (figure 4). When plants were subjected to 254 water stress, we observed the opposite trend: the soil organic fraction was the main sink for 255 the N from fertilizer (56%) followed by maize plants (24%). The N from fertilizer remaining 256 as residual nitrates (0 to 1.50 m [0 to 4.92 ft] depth) was higher in -W plots (8%) than in +W 257 plots (3%). Volatilization was a minor sink of N (3%), with no differences between water 258 treatments. Around 10% of the N applied was not recovered. Roots were also a sink of this 259 unrecovered fraction. Although roots were not quantified in the present study, in an 260 experiment carried out near the experimental site, the amount in roots was around 3% of the 261 applied N (Rimski-Korsakov et al. 2012). The remaining unrecovered N can be attributed to 262 losses by denitrification, leaching below the studied depth, or inaccuracies of the methods 263 applied. Nitrogen losses via denitrification are currently low in the area (<0.6% of applied N) 264 (Palma et al. 1997; Sainz Rozas et al. 2001).

265

Cover crop plants. Nitrogen fertilization and water stress applied to maize showed little effect
on the biomass produced by the CC (table 3). Only a trend (p=0.067) to higher biomass
production of ryegrass when previous maize was fertilized was found (table 3). The N

concentration (1.89% on average) (data not shown) and total N accumulation in the CC crop
aerial biomass was not significantly affected by maize fertilization or water status (table 3).
No significant differences were observed in the N accumulated derived from the soil or from
the fertilizer in the CC (table 3). The N accumulated in CC aerial biomass was 78 kg N ha⁻¹
(70 lb N ac⁻¹) being 4.22 kg N ha⁻¹ (3.77 lb N ac⁻¹) derived from the fertilizer applied to the
maize (table 3).

275 Nitrogen in soil at cover crop harvest. The total nitrates (derived from soil [non-tagged]+ 276 derived from fertilizer [tagged]) from 0 to 1.50 m (0 to 4.92 ft) depth present a trend to show 277 interaction (p=0.068) between the water treatments (+W and -W) on the previous maize and the 278 subsequent CC (+CC and -CC) (tables 4 and 5). The lowest nitrate content in the soil was found 279 when previous maize non-subjected to water stress and the CC was grown; the other treatments 280 showed no differences between them. The nitrates derived from fertilizer in the whole soil profile 281 (0 to 1.50 m [0 to 4.92 ft]) or at any depths showed no significant differences caused by the 282 treatments, being on average 2.6 kg NO₃-¹⁵N ha⁻¹ (2.32 lb NO₃-¹⁵N ac⁻¹) (table 4). About 1.27 kg 283 (2.80 lb) of the NO₃- 15 N derived from the fertilizer was found in the top soil (0 to 0.30 m [0 to 284 0.98 ft] depth) after the CC harvest (figure 5). At the end of the experiment, the nitrate 285 concentration in the soil covered by the CC and in the bare soil was similar. Working in 286 contrasting rainfall years, Willumsen and Thorup-Kristensen (2001) found results similar to 287 those of the present experiment. These authors attributed this result to nitrate leaching in the 288 wet period and the opposite in the dry year, when the uncovered soil showed high nitrate 289 content. A month before the end of the present experiment, high rainfall was recorded (277 290 mm [10.9 in]). Such precipitation could have displaced the residual nitrates deeper in the profile, mainly in uncovered soil. The CC retained 78 kg N ha⁻¹ (70 lb N ac⁻¹) in their biomass 291 292 (table 3), which would be not available to suffer leaching losses. The CC acted as an efficient 293 trap, avoiding losses of residual nitrates.

294 The content of N derived from fertilizer in the soil organic fraction at CC harvest 295 showed no significant differences between treatments, although it was significantly lower than 296 N coming from fertilizer at maize harvest (p=0.001) (figure 3). On average, 35 kg (77 lb) of 297 the N derived from fertilizer was found at the end of the experiment retained in the soil 298 organic fraction. This immobilized N could be taken by the next crop after soil organic matter 299 mineralization. The apparent mineralization of the fertilizer retained in soil organic N during 300 the CC cycle, was on average 47%. Even taking the inaccuracies into account, this amount was 301 order of magnitude greater than the mineralization of soil organic matter in the Pampas's soil 302 (4-5 % yearly) (Alvarez and Alvarez 2000). The N from fertilizer applied to maize remains in 303 the labile fractions of the soil organic matter, thus being more easily mineralized (Alvarez and 304 Alvarez 2000). Then, the N from fertilizer retained in the soil organic fraction could attenuate 305 nitrate leaching but only in the short term. This high mineralization rates show that this N 306 would be released swiftly and that when no other crop retains it, it can be leached.

307 The N retained in CC plants (table 3) plus nitrates remaining in the soil (table 4) showed 308 that the CC treatment accumulated significantly more N (p<0.001) than the nitrates remaining in the uncovered soil during the fallow (130 vs. 51 kg N ha⁻¹ [116 vs. 45 lb N ac⁻¹] respectively). 309 310 The N derived specifically from the fertilizer remaining in plants plus nitrates remaining in the 311 soil was significantly higher in the CC treatment (p=0.03). When N in the soil organic fraction 312 was included, no significant differences in the soil-plant system were recorded among the studied 313 factors, probably due to the organic component magnitude which dilutes the magnitude of other 314 sinks.

315

316 Summary and Conclusions

317 In the conditions of present research, results showed that the main sinks of fertilizer N were

318 maize plants in non-water stressed treatments and the soil organic fraction in the water

319	stressed. Residual nitrates were low (8% vs. 3%, according to water stress). Volatilization was
320	always a minor sink. The CC took up nitrates from the soil leaving fewer nitrates subjected to
321	losses (i.e. leaching) and the N accumulated in CC will be available for the subsequent crop.
322	The N from fertilizer retained in the organic matter is quickly mineralized (around 47% in 6
323	months), releasing N subjected to possible losses in the months ahead.
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428 Figure 1

429 Diagram showing a plot boundaries and the microplot location, and within them the fertilized

430 band, the place of the cylinders for volatilization determination, the sampling plants, the

431 sampling soil site and the rows of maize

432

433 Figure 2

434 Nitrate content from 0 to 150 cm (0 to 59 in) depth at maize harvest. The experiment was

435 carried out in the Argentinean Pampas. Maize was seeded and 2 nitrogen fertilizer doses 0 and

436 140 kg N ha⁻¹ were applied. ¹⁵N tagged fertilizer was used. The crop was subjected to water

437 stress or not subjected to water stress by drop irrigation. Bars: standard error

438

439 Figure 3

Nitrogen located in the soil organic matter at 0 to 30 cm (0 to 11.8 in) depth, derived from the fertilizer at maize and cover crop harvest. Maize was seeded and 2 nitrogen fertilizer doses 0 and 140 kg N ha⁻¹ were applied. ¹⁵N tagged fertilizer was used. The crop was subjected to water stress or not subjected to water stress by drop irrigation. After the maize a cover crop was seeded. Bars: standard error.

445

446 Figure 4

Sinks of the N from the fertilizer in each treatment relative to the N applied via fertilization.
Plant: N accumulated in the total aboveground biomass of maize at harvest; Volatilized:
ammonia-N volatilized from fertilization to crop harvest; Nitrates: nitrates-N from 0 to 150 cm
(0 to 59 in) depth, at maize harvest; Organic N: ¹⁵N measured in the soil organic pool, 0 to 30
cm (0 to 11.8 in) depth; Unaccounted: fraction of the applied N, non-recovered.

452

453 Figure 5

454 Nitrates derived from fertilizer from 0 to 150 cm (0 to 59 in) depth at cover crop harvest. The

- 455 experiment was carried out in the Argentinean Pampas. After maize harvest, annual ryegrass
- 456 was seeded. It was not fertilized and received only rainfall water.
- 457 Bars: Standard error.

- 1 -

Maize biomass production at harvest, total N absorbed (production x N concentration), N
derived from the soil (NdfS) and N derived from fertilizer (NdfF), in aerial vegetative biomass
(stems+leaves+cobs+husks) and grains. N0: not N fertilizer was applied; N140: N fertilization
doses of 140 kg N ha-1 and -W: subjected to water stress; +W: not subjected to water stress.
The standard error is given between parentheses.

Treatments	Biomass	N total	NdfS	NdfF
	(kg DM ha ⁻¹)	(kg N ha ⁻¹)	(kg N ha ⁻¹)	(kg N ha ⁻¹)
Aerial vegetati				
N0-W	7,490 (419) b	31.2 (1.8)	31.2 (1.8)	
N0+W	6,669 (255) b	27.6 (1.0)	27.6 (1.0)	
N140-W	7,767 (933) b	38.7 (7.9)	31.1 (5.5)	7.6 (2.3)
N140+W	10,421 (100) a	51.5 (4.2)	34.9 (5.0)	16.6 (0.9)
Grains				
N0-W	6,799 (543)	77.9 (10.8)	77.9 (10.8)	
N0+W	8,058 (1,011)	94.5 (12.6)	94.5 (12.6)	
N140-W	9,153 (838)	114.9 (12.5)	88.8 (8.0)	26.0 (5.9)
N140+W	13,301 (577)	171.1 (9.4)	122.4 (11.5)	48.6 (6.4)
Total abovegro	ound biomass			
N0-W	14,289 (838) b	109.1 (12.5)	109.1 (12.5)	
N0+W	14,727 (1,267) b	122.2 (11.6)	122.2 (11.6)	
N140-W	16,921 (1,762) b	153.6 (20.3)	119.9 (13.2)	33.7 (8.1)
N140+W	23,723 (663) a	222.6 (5.8)	157.3 (6.5)	65.2 (5.8)

484

485 Letters indicate significant differences between treatments (p<0.05) and the occurrence of a486 positive interaction.

488 Analysis of variance probability values for the effects of water stress (W) and nitrogen (N)

489	availability. Differences were o	considered to be significant at the 5%	(p=0.05) probability level.
	······································		

	Treatments		Interaction
	W	Ν	WxN
	p value	p value	<i>p</i> value
Aerial vegetative biomass			
Biomass production	0.142	0.010	0.018
N concentration	0.953	0.103	0.934
Total N absorbed	0.342	0.013	0.112
N absorbed from soil	0.970	0.384	0.367
N absorbed from fertilizer	0.107		
Grains			
Grains production	0.009	0.002	0.090
N concentration	0.575	0.127	0.949
Total N absorbed	0.032	0.005	0.180
N absorbed from soil	0.073	0.145	0.490
N absorbed from fertilizer	0.032		
Total aboveground biomass			
Biomass production	0.023	0.003	0.037
Total N absorbed	0.038	0.003	0.122
N absorbed from soil	0.091	0.118	0.369
N absorbed from fertilizer	0.038		

Ryegrass aerial biomass production and total N absorbed (production x N concentration), N derived from the soil (NdfS) and N derived from fertilizer (NdfF), in aerial biomass at ryegrass harvest and analysis of variance probability values for the effects of water (W) and nitrogen (N) availability. The water and N treatments are those applied to the preceding maize. NO: not N fertilizer was applied; N140: N fertilization doses of 140 kg N ha-1 and -W: subjected to water stress; +W: not subjected to water stress. The ryegrass was neither fertilized nor irrigated. The standard error is given between parentheses. Differences were considered to be significant at the 5% (p=0.05) probability level.

Treatments	Aerial Biomass	Total N	NdfS	NdfF
	(kg DM ha ⁻¹)	(kg N ha ⁻¹)	(kg N ha ⁻¹)	(kg N ha ⁻¹)
N0-W	2,560 (370)	48.53 (2.68)	48.53 (2.68)	
N0+W	4,053 (352)	83.06 (9.06)	83.06 (9.06)	
N140-W	4,677 (806)	80.82 (17.39)	77.54 (16.65)	3.27 (0.77)
N140+W	5,429 (1,021)	100.09 (19.64)	94.93 (18.29)	5.16 (1.52)
ANOVA p val	lues			
W	0.201	0.144	0.134	0.475
Ν	0.067	0.174	0.228	
WxN	0.652	0.651	0.594	

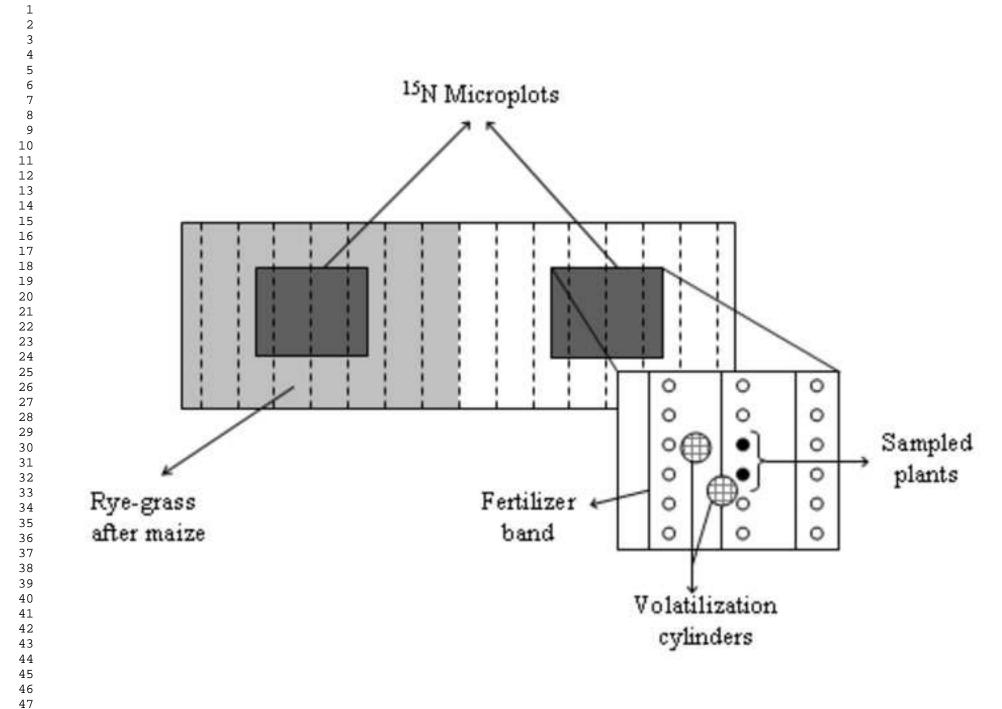
508 Content of total nitrates, nitrates derived from the soil (Nitrates dfS) and nitrates derived from 509 the fertilizer (Nitrates dfF) from 0 to 150 cm (0 to 59 in) to depth (kg NO₃-N ha⁻¹) at ryegrass 510 harvest. N0: not N fertilizer was applied; N140: N fertilization doses of 140 kg N ha-1; -W: 511 subjected to water stress; +W: not subjected to water stress; +CC: with cover crop; -CC: 512 without cover crop. The standard error is given between parentheses.

Treatments	Total nitrates	Nitrates dfS	Nitrates dfF
	(kg NO ₃ -N ha ⁻¹)	(kg NO ₃ -N ha ⁻¹)	(kg NO ₃ -N ha ⁻¹)
N0-W+CC	47.4 (8.8)	47.4 (8.8)	
N0-W-CC	44.0 (3.6)	44.0 (3.6)	
N0+W+CC	40.4 (2.0)	40.4 (2.0)	
N0+W-CC	39.0 (9.7)	49.6 (9.7)	
N140-W+CC	76.7 (16.8)	73.6 (16.1)	3.1 (0.8)
N140-W-CC	66.9 (8.8)	63.8 (8.6)	3.1 (0.2)
N140+W+CC	45.8 (14.3)	43.9 (13.9)	1.9 (0.4)
N140+W-CC	61.6 (23.6)	59.2 (22.4)	2.4 (1.3)

- - •

523 Analysis of variance probability values for the effects of water stress (W) and nitrogen (N) 524 availability and cover crop (CC) presence for content of total nitrates, nitrates derived from the 525 soil (Nitrates dfS) and nitrates derived from the fertilizer (Nitrates dfF) from 0 to 150 cm (0 to 526 59 in) depth at ryegrass harvest. Differences were considered to be significant at the 5% 527 (p=0.05) probability level.

Treatments	Total NO ₃ -N	NO ₃ -N dfS	NO ₃ -N dfF
	<i>p</i> value	<i>p</i> value	<i>p</i> value
N	0.177	0.227	
W	0.440	0.447	0.212
NxW	0.473	0.481	
CC	0.535	0.533	0.735
NxCC	0.993	0.989	
WxCC	0.068	0.063	0.658
NxWxCC	0.498	0.499	



- 48 49

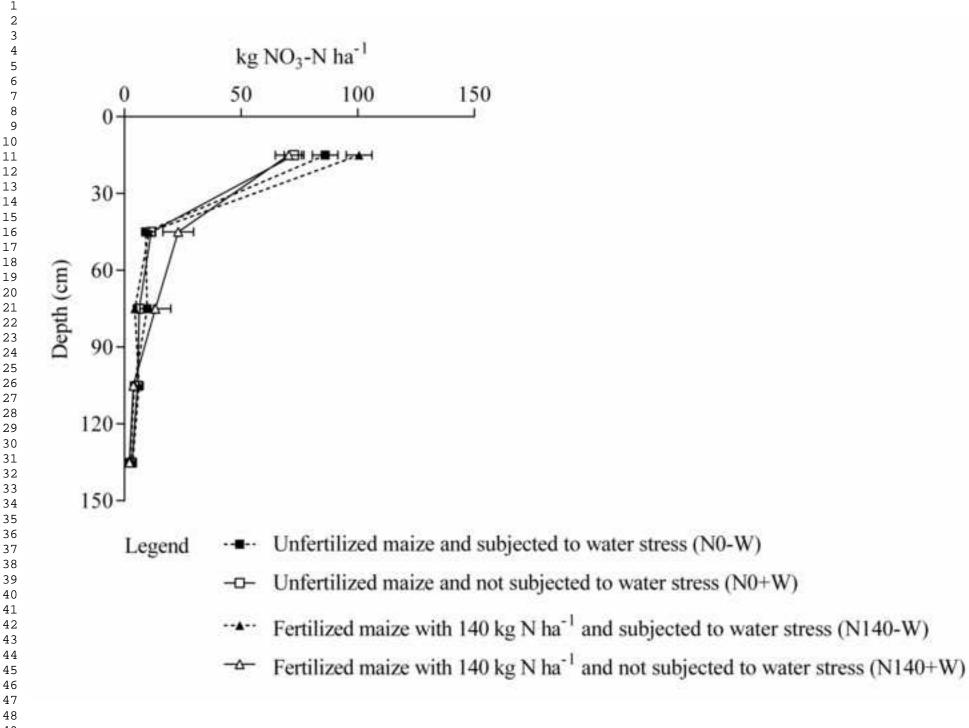


Figure 2B Click here to download Figure: Figure 2B.tif

