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Reaction kinetics and mechanisms of organosilicon fungicide flusilazole with sulfate and hydroxyl radicals



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HIGHLIGHTS

- Flusilazole degradation by photo-Fenton process with solar simulator.
- CG-MS by-products detected show cleavage between silicon and triazole ring.
- A charge transfer reaction mechanism from the fungicide to the radicals is proposed.

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1. Introduction

Advanced oxidation processes (AOPs), like $UV/S_2O_8^{2-}$ and Fenton, have attracted interest for the degradation of organic compounds in wastewater, such as pesticides (Carra et al., 2016;

G R A P H I C A L A B S T R A C T



ABSTRACT

Flusilazole is an organosilane fungicide used for treatments in agriculture and horticulture for control of diseases. The reaction kinetics and mechanism of flusilazole with sulfate and hydroxyl radicals were studied. The rate constant of the radicals with the fungicide were determined by laser flash photolysis of peroxodisulfate and hydrogen peroxide. The results were $2.0 \times 10^9 \text{ s}^{-1}\text{M}^{-1}$ for the reaction of the fungicide with HO[•] and $4.6 \times 10^8 \text{ s}^{-1} \text{ M}^{-1}$ for the same reaction with SO₄⁺⁻ radicals. The absorption spectra of organic intermediates detected by laser flash photolysis of $S_2O_8^2^-$ with flusilazole, were identified as α -aminoalkyl and siloxyl radicals and agree very well with those estimated employing the time-dependent density functional theory with explicit account for bulk solvent effects. In the continuous photolysis experiments, performed by photo-Fenton reaction of the fungicide, the main degradation products were: (bis(4-fluorophenyl)-hydroxy-methylsilane) and the non-toxic silicic acid, diethyl bis(trimethylsilyl) ester, in ten and twenty minutes of reaction, respectively.

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Navarro et al., 2011). This is especially true in the case of the use of light-induced reactions in water treatments. The photo-assisted Fenton reaction ("photo-Fenton" reaction) typically gives faster rates and a higher degree of mineralization than the thermal reaction and can take the advantage of light in the solar spectral region (Faust and Hoigné, 1990). In photo-Fenton process Fe^{2+} ions are oxidized by H_2O_2 and one equivalent HO• is produced. The obtained $Fe(OH)^{2+}$, which is predominant in acidic conditions, act as the light absorbing species that produce another radical while



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the initial Fe²⁺ is regained (Kim and Vogelpohl, 1998). Persulfate reaction mechanism is based on the generation of the strong sulfate radical (SO₄*⁻). One mole of this reagent can be activated to generate 2 mol of sulfate radical under UV irradiation. The sulfate radical is known to be a very strong oxidant, i.e. E (SO₄*⁻) SO₄²⁻) = 2.43 V vs. NHE (Criquet and Leitner, 2009) for degrading organic contaminants in wastewater (Wang and Liang, 2014; Luo et al., 2017).

Flusilazole is the approved name for 1-[[bis(4-fluorophenyl) methyl]silyl]methyl]-1*H*-1,2,4-triazole (CAS No. 85509-19-9). It is a triazole family fungicide with a molecular structure shown in Scheme 1. It is a broad-spectrum fungicide, which exhibits curative and preventative activities and is recommended for use in agriculture and horticulture (T. R. Roberts and Hutson, 1990).

Pesticides can easily accumulate in plants, foods, water reservoirs and food chain, which could be triggered by indiscriminate use. Besides, some of them are resistant to the action of sunlight, temperature, water or microorganisms; as a result of this, plant tissue and environmental samples contain high levels of pesticides residues.

It was suggested that flusilazole is capable of inducing genotoxicity in plants in studies with higher plant models (Ozakca and Silah, 2013). Animal models in vitro and in vivo have been used to investigate the teratogenecity of flusilazole. Some studies reported marked maternal toxicity, growth retardation, and skeletal in rodent models, which contribute to alterations in fetal growth retardation and skeletal development (Farag and Ibrahim, 2007). Besides embryotoxic responses are critically dependent on the timing of exposure during rat embryo development (Dimopoulou et al., 2016).

Photolysis (artificial and natural sunlight) on soil surface and aqueous photodegradation is not an important mode of degradation for flusilazole (T. R. Roberts and Hutson, 1990). Despite flusilazole is a widely used fungicide, documented data on their degradation mechanism and intermediate metabolites is lacking. Differences in chemical reactivity between silicon and carbon compounds influence the degradation and oxidative metabolism of otherwise similar compounds. For instance, organic silicon compounds usually lead to naturally occurring silicates as final product, in sharp contrast with the carbon containing analogous compounds (Griessbach and Lehmann, 1999).

In this context, the aim of the present work is to investigate the kinetics and mechanism of the degradation of flusilazole initialed by HO• and SO₄•⁻. These radicals have been generated by laser flash photolysis ($\lambda_{exc.} = 266 \text{ nm}$) of H₂O₂ and Na₂S₂O₈, respectively. Also, continuous-irradiation experiments have been performed by photo Fenton reaction of the fungicide and the nature of the primary degradation products formed have been investigated by GC-MS analysis.

2. Experimental

Reagents. Flusilazole (1-((bis(4-fluorophenyl)methylsilyl)



Scheme 1. Flusilazole.

methyl)-1H-1,2,4-triazole), was purchased from Sigma-Aldrich (Fluka). Sodium peroxodisulfate was obtained from Merck. Ferrous chloride tetrahydrate (FeCl₂·4H₂O) 98% was obtained from Fisher Chemical. H₂O₂ 30% (w/w) in H₂O was supplied by Sigma-Aldrich. Potassium thiocyanate was from Mallinckrodt.

Distilled water was passed through a Millipore system (>18 MX cm, <20 ppb of organic carbon).

Sample Preparation. Flusilazole (F) saturated aqueous solution was prepared to set up suspensions of lower concentrations by dilution. Oxygen-free and oxygen saturated solutions were obtained by bubbling the samples with argon or oxygen, respectively, for 20–30 min before use.

In continuous photolysis experiments the pH of the solution was adjusted to 3.0 using HCl and the sample was mixed using a magnetic stirrer during the process.

Laser Flash Photolysis (LFP) Experiments. LFP experiments were performed by excitation with the fourth harmonic of a Nd:YAG Litron laser (2 ns fwhm, 2 mJ per pulse at 266 nm). The analyzing light from a 350 W Xe arc lamp was passed through a monochromator (PTI 1695) and detected by a 1P28 PTM photomultiplier. Due to the characteristics LFP detection system used, traces with $\lambda > 650$ nm and $\lambda < 290$ nm could not be detected within the experimental error. A cuvette of 10 mm path length was used for the sample container. Decays typically represented the average of 32 laser shots and were stored in a 500 MHz Siglent digital oscilloscope. Solution absorbance was 0.5 at 266 nm. Reactant concentrations were: [Na₂S₂O₈] = 1.0–2.5 × 10⁻² M and flusilazole was in the range 8.49 × 10⁶ M – 8.50 × 10⁻⁵ M.

Freshly prepared solutions were used in order to avoid thermal reactions of peroxodisulfate and hydrogen peroxide with the substrate.

Bilinear regression analysis. For each experimental condition, several absorbance decay profiles at different detection wavelengths were taken. Absorbance is thus a function of wavelength and time. Taking advantage of the linearity of the absorbance with both, concentrations and absorption coefficients, a bilinear regression analysis was applied to the experimental absorption matrix in order to retrieve information on the minimum number of species and on their relative concentration profiles and absorption spectra (San Román and Gonzalez, 1998).

Time-dependent density functional theory calculations. Time dependent density functional calculations, TD-DFT, has been employed in numerous studies as a useful tool to characterize electronically excited states of a large number of molecules (Dell 'Arciprete et al., 2011; Arce et al., 2012). In this work, the hybrid B98 (Schmider and Becke, 1998) density functional combined with the 6-311++G(d,p) basis set, was used to interpret the spectra of the observed transient species. To account for solvent effects, the conductor-like polarizable continuum model, CPCM, was employed using a relative permittivity for H₂O of 78.3553 (Barone and Cossi, 2001). In this way, the interaction between solvent and solute molecules was considered. All calculations have been carried out with the Gaussian 09 package (Frisch et al., 2009). To obtain the theoretical absorption spectrum it is firstly necessary a full optimization of the geometric parameters of the considered groundstate molecule. In all the cases, real frequencies were obtained indicating that molecular structures correspond to stable species. Next, the vertical electronic energies, the associated wavelengths of the band maxima, λ_{max} , and oscillator strengths f were computed at the abovementioned level of theory.

To compare the prediction with experimental data, theoretical bands were approached with individual Gaussian functions characterized with full width values σ (at the 1/e height of the bands) and summed over all relevant electronic transitions (Cobos and Croce, 2010). As Fig. 3 shows, a satisfactory agreement between

the experimental and the modeled spectra for all species are obtained using a standard value of $\sigma=0.21$ eV.

2.1. Chemical analysis

The concentration of flusilazole was determined by HPLC and reaction products of fenton reactions were analyzed by GC-MS.

Samples were irradiated in a cylindrical glass reactor with a solar simulator (Oriel Instruments, Model 81160) equipped with a 300 W xenon lamp and a cut off 300 nm filter. The prepared solution was 5 ppm of Fe²⁺ and 1.23 \times 10⁻³ M of flusilazole. 50 μ L of H₂O₂ was added to a volume of 250 mL.

HPLC analysis. In order to follow the concentration of the flusilazole in the reaction system, a Merck Hitachi XL Autosystem D-7000 chromatograph with a diode array detector was used. The column used to separate the flusilazole was a LiChroCART[®] 125-4 with reverse phase LiChrospher[®] 100 RP-18 (5 μ m) from Merck. Samples of 0.5 mL were directly taken from the reaction system and they were diluted with 0.5 mL of methanol. The injection volume was 80 μ L. As mobile phase a mixture of water:methanol (20:80) (v/v) were used in a flow 0.8 mL min-1. The detection wavelength was 230 nm and the retention time for the flusilazole was about 5.1 min.

GC-MS. After selected periods of irradiation (10 or 20 min), the reactor was emptied and the treated solution was concentrated by means of solid phase extraction: 100 mL of the aqueous solution were flown though a LiChrolut EN 200 mg (Merck) cartridge and the adsorbed organics were recovered with 3 mL of methanol. A GCMS-QP2010S (Shimadzu) gas chromatograph equipped with a quadrupole mass analyzer was employed to identify major intermediates formed along the process. The temperature program for GC analysis involved a heating method with a constant ramp from 60 °C to 250 °C at a rate of 5°C/min rate. A Meta X5 Teknokroma column was used. The retention time for the flusilazole and the observed products are listed in Table 1.

3. Results and discussion

3.1. Experiments with HO• radicals

3.1.1. Continuous photolysis experiments

Irradiation of 1.23×10^{-3} M aqueous solutions of flusilazole with the solar simulator, shows depletion of the fungicide with the photolysis time. According to Fig. 1, 5 min of irradiation were needed to achieve 80% abatement of flusilazole in the presence of 5 ppm Fe²⁺ and 50 µL of H₂O₂ at pH = 3.

3.1.2. Stable products identification

Aqueous solutions containing Fe^{2+} , H_2O_2 , and flusilazole in the concentrations indicated above, were irradiated for 10 and 20 min with the solar simulator to detect the degradation products formed after reaction 2.

Table 1 summarizes the GC-MS mass peaks and fragments detected for the degradation products of the fungicide.

After 20 min of photo-Fenton reaction with flusilazole, the GC-MS peaks of $R_t < 26.5$ min could not be identified due to the complex reaction system, which lead to a high number of reaction by-products reaching low concentrations.

3.1.3. Kinetic measurements

Hydroxyl radicals (HO•) were generated by LFP of 0.1 M H_2O_2 aqueous solution ($\lambda_{exc.}=266$ nm), reaction 1.

$$\begin{array}{ccc} H_2O_2 & \xrightarrow{h\nu} & HO^{\bullet} \\ F & + & HO^{\bullet} & \longrightarrow \end{array} & \begin{array}{c} 1' \\ 2 \end{array}$$

The rate constant for the reaction of flusilazole with HO• radicals (reaction 2) was assessed by the competition method *k* (F + HO•) = $2.0 \times 10^9 \text{ s}^{-1} \text{M}^{-1}$ (Kozicki et al., 2003) (Dell 'Arciprete et al., 2009) (See Supplementary data). The concentration range of flusilazole solutions used was $0-1.12 \times 10^{-4}$ M.

A degradation mechanism is proposed for the reaction of hydroxyl radicals and flusilazole.

Tertiary amines are reactive sites for HO• radical attack (Das and von Sonntag, 1986) (Dell 'Arciprete et al., 2009). The simplest oxidation process for a tertiary amine is single electron transfer from the non-bonding electron pair of the amine to form a radical cation, followed by a α -CH bond cleavage or hydrogen atom transfer. The process will be denoted as hydrogen-atom transfer if the electron and proton acceptor are the same identity (Sumalekshmy and Gopidas, 2005; Suresh Das and V. Suresh, 2001).

Scheme 2 show a possible reaction pathway, which involves an initial electron transfer by HO• radicals to yield a radical cation in the N tertiary atom of triazole ring followed by elimination of H⁺ ion to yield a more stable α aminoalkylradical.

3.2. Experiments with $SO_4^{\bullet-}$ radicals

It is not possible to detect organic intermediates generated by reaction of hydroxyl radicals with flusilazole using the indirect method. In order to confirm the first steps of the mechanism proposed in Scheme 2, LFP experiments initiating the oxidation of flusilazole with highly oxidative species such as sulfate radicals $(SO_4^{\bullet-})$ were performed.

3.2.1. Kinetic measurements

In LFP experiments of aqueous $1.0-2.5 \times 10^{-2}$ M peroxodisulfate solutions, a transient specie was formed in the wavelength range of 330–500 nm (reaction 3). The obtained decay rate and spectrum are in agreement with those reported in the literature for SO₄-⁻ (Choure et al., 1997).

$$S_{2}O_{8}^{-2} \xrightarrow{hv} 2SO_{4}^{-}$$

$$F + SO_{4}^{-} \longrightarrow 4$$

 $Fe^{2+} + H_2O_2 \longrightarrow Fe(OH)^{2+} + HO^{\bullet}$ $Fe(OH)^{2+} + hv \rightarrow Fe^{2+} + HO^{\bullet}$ $HO^{\bullet} + F \longrightarrow 2$ 2

Table 1

Degradation	products.







Fig. 1. Fungicide depletion (1.23 \times 10 $^{-3}$ M) in photolysis experiments with light of solar simulator at pH = 3.

The SO₄•[–] radical absorbance decay at detection λ (ΔA), fit to Eqn (1) which shows both first and second—order components (See Supplementary data):

$$\Delta A = \frac{a}{b(\lambda)\exp(at) - c(\lambda)} \tag{1}$$

Where: *a* is the apparent first order decay constant (K_{app}), and $b(\lambda) = (a / \Delta A_0) + c(\lambda)$. The second order constant $2k/\epsilon(\lambda)$ is calculated from $c(\lambda) = 2k/\epsilon(\lambda)l$, where $\epsilon(\lambda)$ is the absorption coefficient at a certain wavelength, and *l* is the optical path. ΔA_0 is absorbance at t = 0.

Photolysis experiments of sodium persulfate aqueous solutions in the presence of flusilazole ($4.25 \times 10^{-5} \text{ M} > [\text{F}] > 8.47 \times 10^{-5} \text{ M}$), showed formation of transient species in the wavelength range from 300 to 650 nm. The decay kinetics of the traces observed strongly depend on wavelength. Therefore, a bilinear regression analysis was used to retrieve kinetic and spectroscopic information of the individual species in the reaction mixture.

Sulfate radical anion is the only specie that absorbs at $\lambda = 450$ nm. Sulfate radical traces show a first order decay with an apparent rate constant, k_{app}, increasing with flusilazole concentration, as depicted in the inset of Fig. 2.

The slope of the straight line in Fig. 2, yields the rate constant k for the reaction of sulfate radicals with flusilazole (reaction 4) k $(F + SO_4^{-}) = (4.6 \pm 0.4) \times 10^8 \text{ s}^{-1}\text{M}^{-1}$.



3.2.2. Intermediates formed after the reaction of SO_4 - radicals with flusilazol

At high flusilazole concentrations the decay rate of $SO_4^{\bullet-}$ radicals becomes too fast to be observed in our time resolution range. On the other hand, laser flash photolysis of flusilazole solutions in the absence of persulfate showed no traces, within our experimental error, in the UV and visible range. Therefore, under these conditions, the observed traces may be assigned to organic transients formed after reaction of $SO_4^{\bullet-}$ with flusilazole (reaction 4).

To study the spectra and decay rates of the organic intermediates, experiments were performed with solutions containing the highest concentration of flusilazole used (8.47 \times 10⁻⁵ M) and 2.5 \times 10⁻² M of persulfate.

A bilinear analysis of the absorbance matrix (see Experimental Section above), shows that the system could be described by two absorbing species in the wavelength range from 290 to 640 nm, named transient 1 and transient 2. The first transient absorbs at wavelengths of 300 nm, and the second one has two maxima at

350-360 nm and 640 nm, respectively.

In order to gain further insight into the transients' nature, the reaction with molecular oxygen was investigated. To that purpose, LFP experiments under argon, air and oxygen saturated atmosphere with 2.5 \times 10^{-2} M $S_2O_8^{2-}$ solutions and 8.47 \times 10^{-5} M of flusilazole were performed. The absorption spectra of the organic transient in the range 300–400 nm (transient 1), were similar in all three cases. within the error of the determination (See Fig. S1 at Supplementary data). The decay profiles of this transient fit to a mixed first and second order reaction rate law in the presence of oxygen (absorbance traces fit to Eqn (1)). Mainly a second order decay was observed in Ar- saturated solutions (see equation at Supplementary data). The first order decay rate constant was $2.2 \times 10^3 \text{ s}^{-1}$ and 3.4×10^3 s⁻¹ in air- and oxygen-saturated samples, respectively. The second order component $2k/\epsilon_{(300nm)} \sim 6 \times 10^6$ cm s⁻¹ was independent on the presence of oxygen. The presence of a second order term in the decay traces might indicate a recombination of transient 1.



Fig. 2. Apparent rate constants, K_{app} vs. [F] obtained in experiments with 2.5×10^{-2} M sodium peroxodisulfate $Na_2S_2O_8$ aqueous solutions. **Inset**: Absorbance profiles at 450 nm obtained in experiments with 2.5×10^{-2} M peroxodisulfate $S_2O_8^{-2}$ solutions in the presence (dark grey curve) and absence (light grey curve) of 4.25×10^{-5} M of F. The full lines stand for the fittings to Eqn (1).

The second transient had two absorption maxima at 350 nm and 640 nm with a decay profile that fits a mixed first and second order equation (Eqn (1)). No dependence of both first and second order constants with the concentration of dissolved oxygen was observed (see Table 2). However, the absorbance intensity of the second transient was dependent on the concentration of dissolved oxygen.

The one-electron charge transfer reaction from amine nitrogen to sulfate radicals to yield a radical cation in the N atom has been reported (Dell 'Arciprete et al., 2011). On the other hand, the rate constants for the SO_4 -⁻ hydrogen-abstraction from C–H bonds were smaller than those measured here for the reactions of sulfate radical with the fungicide (10^5 to 10^7 s⁻¹ M⁻¹) (Neta et al., 1988).

Nitrogen-centered radical cations do not efficiently react with molecular oxygen, as observed in other organic transients (Bosio et al., 2005; Dell'arciprete et al., 2007). It is discussed in literature that α -aminoalkyl radicals are formed after the elimination of H⁺ from nitrogen-centered radical cations; this α -aminoalkyl radicals are highly stabilized by the free electronic pair of the vicinal nitrogen (Padmaja et al., 1993; Luke et al., 2003; Ito et al., 2009). Considering that the amine-type nitrogen atom in the triazole moiety of flusilazole present H-atoms α to nitrogen, identification of the observed transients as α -aminoalkyl radicals is strongly suggested. Furthermore, α -aminoalkyl radicals show significant absorbance up to 500 nm and react with molecular oxygen (Dell 'Arciprete et al., 2009; Hiller and Asmus, 1983; Lalevée et al., 2007), in agreement with the behavior of the first transient herein observed.

The second transient signal grows quickly, reaching its maximum value about 13 μ s after the laser shot, and it decays within 500 μ s (See Fig. S3 at Supplementary data). Besides, its formation depends on the concentration of dissolved oxygen in the sample. This behavior may be explained by a reaction mechanism in which this transient is formed via the reaction of the first



Fig. 3. Absorption spectra in arbitrary units of transient 1 (black circles) and transient 2 (grey circles) obtained at 1×10^{-5} sec from bilinear analysis of LFP experiments with 2.5×10^{-2} M $S_2O_8^{2-}$ and 8.47×10^{-5} M Flusilazole in argon saturated solutions. The full, dotted and dashed lines stand for the calculated TD-DFT spectra.

transient with oxygen. Peroxyl radicals generated by reaction of α aminoalkyl radicals with oxygen could lead to the formation of siloxyl radicals, which are assigned to the second transient, as will be discussed later in the manuscript.

3.2.3. Reaction pathways

As discussed before, a charge transfer pathway leads to the formation of sulfate anions and the radical cation of flusilazole, which upon elimination of H^+ leads to α -aminoalkyl radical.

To help identify the nature of the observed transients, the TD-DFT spectra of α -aminoalkyl radicals in carbon (structures *a* and *b*) and the radical cation were calculated.

The coincidence obtained between the experimental first transient spectrum and those calculated for the *a* structure of α -aminoalkyl radicals, indicates that the observed transient may be assigned to the specie where carbon radical lies in the carbon atom vicinal to silicon in flusilazole molecule (Fig. 3). Radical cation in the N triazole atom is proposed in the first reaction step of Scheme 2. A reactive intermediate with a TD-DFT spectra absorption bellow 240 nm with $\lambda_{max.} = 210$ nm, was not detectable in the time window of our experimental LFP setup.

Calculations for the *b* structure of α -aminoalkyl radical in carbon, shows a spectrum with a low absorption coefficient at 300 nm (Fig. 3).

The mechanism discussed above strongly confirms that the more reactive group of flusilazole is the triazole ring, as was proposed according to the by-products detected by GC-Mass in photo-Fenton reactions.

The α -aminoalkyl radicals (-HC•-N<) (Hiller and Asmus, 1983; Baciocchi et al., 2004) may either further react with molecular oxygen to yield peroxyl radicals (*c structure*), or recombine leading to the eventual production of flusilasole-amine dimer (Hasegawa et al., 1988) (See Scheme 3).

Peroxyl radicals (*c structure*) may produce siloxyl radicals through a mechanism that involves silicon pentacoordinated

Table 2

First order decay component (a) and second order one (b) obtained from the fit of decay traces of transient 1 and 2 by Eqn (1).

	а	$b=2\mathrm{k}/arepsilon(\lambda)$
Transient 1 (λ_{max} ~300 nm)	$2.2 \times 10^3 \text{ s}^{-1}$ (Air)	$6 \times 10^6 \text{cm} \text{s}^{-1}$ (bubbling Ar, air and oxygen)
Transient 2 (λ_{max} ~350 and 640 nm)	$1 \times 10^4 \text{ s}^{-1}$ (bubbling Ar, air and oxygen)	$1\times10^{6}~\text{cm}~\text{s}^{-1}$ (bubbling Ar, air and oxygen)

intermediates as shown in Scheme 3. A similar mechanism was proposed by Corriu and co-workers for the hydrolysis of tetracoordinate silicates initiated by a nucleofilic attack leading to pentacoordinate silicon intermediate (Corriu et al., 1991).

Formation of transient d (d structure) is strongly suggested, due to the high affinity of oxygen to silicon. Si–O• radical could be depleted by reaction with hydrogen peroxide or other hydrogen donor in the media, leading to the formation of silanol.

The TD-DFT calculated spectra of peroxyl radical (structure *c*), has a wavelength maximum around 270 nm, not observed in the experimental conditions of this setup. However, the second transient with λ_{max} at 350 and 640 nm, is very similar to the one calculated called *d* radical which it is also shown in Fig. 3 and may be assigned to that specie.

3.3. Reaction of HO• radicals with flusilazol in continuous photolysis experiments. Reaction pathways

Based on the detailed reaction mechanism for the first steps of reactions of $SO_4^{\bullet-}$ with flusilazole with identification of organic transients formed, a complete pathway for the initial photochemical products and primary steps of HO• radical oxidation of flusilazole is proposed in Scheme 4 y 5 yielding compound 1, compound 2 and 2' (GC-MS products).

The proposed mechanism is further supported by the fact that the rate constants reported for these reactions are on the order of those found here for the reaction of HO• with the fungicide (Hiller and Asmus, 1983; Ross et al., 1998).

As shown in Scheme 4, the proposed reaction mechanism might involve the formation of silicon pentacoordinated intermediates after reaction of the α -aminoalkyl radical with oxygen. This

intermediate may explain the formation of identified products 1 and 2'. Besides, compound 1 could be produced by hydrolysis of compound 2' (Tuazon et al., 2000) or by depletion of Si–O• radical with a hydrogen donor, such as hydrogen peroxide. Siloxyl radical intermediate could be proposed considering that the reaction mechanism with hydroxyl radical is analog to that described of flusilazole with sulfate radical (see Schemes 2 and 3).

Disproportionation reaction of peroxyl radicals (Scheme 5), may lead to the formation of hydroxyl and keto derivatives of flusilazole, neither of which was detected by GC-MS spectroscopy. Compound 2 detected, is justified by hydroxyl radical attack on triazole ring of keto derivatives of flusilazole, followed by hydrolysis.

The observation that no nitrogen derivates were detected by GC-MS could be understood by a fast rupture of silicon pentacoordinated intermediates followed by oxidation of the triazole ring (Lhomme et al., 2007; Guillard et al., 2002; Le Campion et al., 1999).

Twenty minutes photo-fenton reaction with flusilazole leads to the formation of products with higher molecular weight than flusilazole. GC-MS spectra for retention times > 28 min showed formation of compounds containing Si-O-C and Si-O-Si groups, such as silicic acid, diethyl bis(trimethylsilyl) ester. The main route to obtain siloxane functional group is by condensation of two silanols like those formed in the first minute's fenton reaction, as showed in Scheme 3. Besides, molecules containing Si-O-C groups are proposed to be generated as by-product in the same scheme (compound 2'). Silicic acid, diethyl bis(trimethylsilyl) ester has no fluorine content in its structure. It indicates that between ten and twenty minutes of photo-Fenton process, loss of fluoride starts taking place.

GC-MS peaks of R_t < 26.5 min are not easy identified in 20 min reaction. This might be because the reaction system is too complex





to recognize low molecular weight by-products. Amat et al. (2009) described a similar complex system were a great number of low molecular weight CG-MS products generated by fast cleavage of molecules, which did not allow the elucidation of a detailed mechanism for the degradation of the chemicals.

In flusilazole metabolism in rats, goat, hens, apples, wheat, grapes and soil (Mastovska, 2007; T. R. Roberts and Hutson, 1990) and in environmental fate studies of the insecticide, the main degradation product detected was [Bis(4-fluorophenyl)methyl] silanol ($C_{13}H_{12}F_2OSi$) of molecular weight 250 g/mol. Other polar metabolites containing silanol groups were detected in the tissue residues and excretion samples analyzed in rats, goats, and hens.

Silicic acid, diethyl bis(trimethylsilyl) ester is a non toxic compound that is found in root and inner bark extract of some medicinally important plants against human pathogens with antioxidant activity (Anjaneyulu et al., 2013; Hosseinihashemi et al., 2015; Kathirvel et al., 2014).

This compound is detected in GC-MS spectrum after the photo Fenton degradation of flusilazole, what suggests that degradation of that organosilicon fungicide by this process is an important route of decomposition yielding non toxic by-products. Cleavage between the silicon and the triazole ring to form the silanol is the major metabolic route in plants, which may be further metabolized to the silane diols or to disiloxane (Mastovska, 2007). The results in twenty minutes photo fenton experiments showed that a siloxane (a silane ether) is produced, in line with the reported metabolic route.

4. Conclusions

The fungicide flusilazole chemically react with hydroxyl radical and sulfate radical anion with rate constants of 2.0 \times 10⁹ s⁻¹ M⁻¹ and 4.6 \times 10⁸ s⁻¹ M⁻¹, respectively.

Both reactions involve a charge transfer from the fungicide to the radicals, characteristic for tertiary amines. Proton elimination from the ethyl carbon vecinal to the triazole N atom yields α -aminoalkyl radicals detected as the main transients formed. The absorption spectra of the transients observed in the range 300–700 nm could be assigned to carbon centered radicals and siloxyl radicals in accordance to the DFT calculated spectra and with data reported by Dell 'Arciprete et al. (2011). The amidine nitrogen of the molecule is the preferred site of attack by HO• and SO₄•-



Scheme 5.

radicals.

In the degradation process, the presence of the element silicon in the fungicide compound is essential to generate the intermediates proposed in the reaction path. The by-products identified, could further yield silicates as the final product of degradation.

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Appendix A. Supplementary data

Supplementary data related to this article can be found at https://doi.org/10.1016/j.chemosphere.2017.09.134.

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