1 Bird community changes associated with cattle raising management in the delta forests of

2 the Paraná River

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9 Abstract

10 Riparian forests and environments close to watercourses support high biodiversity, which may be modified by human activities. In the Paraná River Delta region, cattle raising is 11 12 one of the activities with the greatest impact, altering vegetation structure. These changes are reflected in bird communities that inhabit these environments. We hypothesize that the 13 absence of large herbivores (whether cows or native deer) will produce an increase in the 14 15 vegetation cover of the lower strata of forests, due to the greater height of the herbaceous stratum and greater coverage of the shrub stratum, and that these changes mainly affect the 16 species of birds that use resources associated with these strata. Our objective was to 17 18 understand the changes in richness and abundance of the bird community among different types of cattle management using a functional aggregation approach of bird species. In the 19 areas with cattle, we found less coverage of the shrub layer, lower height of the herbaceous 20 21 layer and fewer climbing vines. We found that changes in richness and abundance of bird 22 community were strongly related to species associated with the lower vegetation strata (ground-feeding guild, shrub guild and low canopy guild), and that the responses of the 23 different guilds were not homogeneous. Understanding the direction of changes in bird 24

communities occurring in response to modifications of the environment, allows us to optimize
conservation efforts. If these efforts are based on conservation of the environment in its
natural state, we should adjust the management of the herbivory to the pristine conditions of
the environment. On the other hand, if conservation efforts are based on particular species or
groups of species associated with certain resources, the correct management of herbivory by
cattle can be essential to obtain successful results.

31 **Keywords**: Anthropic impact, functional guild, riparian forests, vegetation structure.

32 Introduction

Riparian forests and environments close to watercourses support high biodiversity but 33 34 most of these environments are subject to anthropogenic processes, such as fishing, hunting, house construction, and navigation, extraction of firewood and trees, and cattle raising 35 (Croonquist & Brooks 1991, 1993, Berduc et al. 2015, Kandus & Quintana 2016). The latter 36 activity is considered one of the anthropogenic processes that changes the environment the 37 most, because presence or absence of large herbivores like cows modifies vegetation structure 38 39 (Jansen & Robertson 2001, McIntyre et al. 2003, Quintana et al. 2014). Spatial homogenization of the habitat is expected at very high cattle raising intensities, while low or 40 moderate cattle raising levels may increase spatial heterogeneity of the habitat (Fuhlendorf & 41 42 Engle 2001). This increase in spatial heterogeneity of the habitat is positively related to the 43 overall diversity of bird species, both locally and at landscapes scales (Verdú et al. 2000, Frutos et al. 2016, Penteado et al. 2016). 44

Bird communities from around the planet have been reported to display differing
responses to the presence or absence of cattle, and associated grazing pressures. For example,
an Australian study has found with the majority of bird species declining with increasing
grazing pressure (Martin & Possingham 2005). In the United States, bird response (negative or

49 positive) may depended on the intensity of cattle raising (Fuhlendorf & Engle 2001). Similar 50 results have been found in European regions (Laiolo et al. 2004). On the other hand, other studies have found negative responses of bird communities in terms of density and richness 51 after cattle had been excluded (García et al. 2008). Apparently, the effect of cattle density on 52 53 bird communities in the environment depends not only on the loads used, but also on the specific herbivory context of the region, possibly with a marked influence of the recent 54 55 evolutionary processes of the environments in the presence or absence of large herbivores, and 56 of the current populations of native herbivores (García et al. 2008, Okes et al. 2008, Kay et al. 57 2017).

Different species of birds may respond differently to changes or disturbances in the 58 59 environment, so some species may be favored or harmed, and others not altered (Rotenberry & 60 Wiens 1980, Weller 2003, Okes et al. 2008). The guild approach proposing supra-specific groups has been profusely applied in ornithological studies, as an alternative to the more 61 62 complicated species-specific approaches (Croonquist & Brooks 1991, Farías et al. 2007, 63 Bejarano et al. 2011). Species classification into guilds according to resource use, facilitates 64 bird community studies, allowing comparisons regarding functional organization of the guilds (Terborgh & Robinson 1986, Gitay & Noble 1997, Wilson 1999, Blondel 2003). This is a 65 66 functional approach because for species with similar habitat uses and food requirements 67 (similar ecological function), a similar response to spatial-temporal variation in the habitats is expected, allowing to examine differences that may not be observed at the species level or in 68 69 the structure of bird communities.

We hypothesized that absence of large herbivores (cows) would result in an increase in
vegetation cover of the lower strata of the riparian forest, with greater height of the herbaceous

layer and a denser cover of the shrub stratum, and that these changes would mainly affect
guilds that use resources associated with these strata. Our aim was to understand changes in
richness and abundances of the bird communities under different types of cattle raising
management in riparian forests of the Paraná River delta, using a functional guild approach of
bird species.

77 Materials and methods

78 *Study area*

The study was conducted in floodplain forests in the Pre-Delta region of the Paraná
River, from Pre-Delta National Park (32 ° 03 ' S; 60 ° 38 ' W), south to Islas de Santa Fe
National Park (32 ° 25 ' S; 60 ° 49 ' W), Argentina. The climate in the area is temperate/warm
humid (Kottek et al. 2006). Annual average temperature is about 19 °C and annual rainfall is
~900 mm, with warmer and rainier summer periods than in winter.

84 *Sampling design*

Data were collected in riparian forest under three different types of cattle-raising
management ("type of management" hereinafter): riparian forest on islands with cattle (IwC);
riparian forest in the Pre-Delta National Park on islands with > 20 years of cattle exclosure
(NC); and riparian forest in the Islas de Santa Fe National Park on islands with recent cattle
exclosure (RCE), approximately 4 years before this study began.

The sampling design consisted of four point counts location (separated from each other by at least 250 m) along each of three transects (every transect on a different island) in each of the types of management (9 transects, 36 total point counts location). Transects within the same management type were separated by at least 1,500 meters. Counts were conducted two times (with intervals of 45 days between counts within a season) in summer, autumn, winter, and spring, for three years (2014 - 2016). In total, we performed 864 counts (3 managements x

3 transects x 4 point counts location/transect x 8 times / year x 3 years). The forest structure
was sampled in 15-m x 15-m plots of vegetation centered at each counting point during each
visit. We estimated the percentage of arboreal and shrub layer cover (by visual estimation), the
average height of the herbaceous stratum (taking five random samples with metric rods and
averaging them) and presence or absence of the climbing vines (Matteucci & Colma 1982). *Bird samples*

102 We counted all birds seen or heard during 10 min in a 75 m radius at each point count location. All counts were conducted during the first 4 hours after sunrise and the sampling 103 104 order of the points within each type of management was alternated between the two samples in 105 each season to reduce any bias associated with bird activity and time of day (Frutos et al. 106 2019). We followed Pearman and Areta (2020) for systematics, and categorized residency 107 status of species based on Fandiño and Giraudo (2010): 1) residents (R): species that remain in 108 the area throughout the year; 2) northern austral migrants (NAM): species that nest in Argentina in spring and summer and migrate northward outside of Argentina during the 109 110 autumn-winter period; 3) southern austral migrants (SAM): species that nest in the south of Argentina in spring and summer and that disperse to the north and east of Argentina in autumn 111 and winter; and 4) longitudinal migrants of the west (LMW): species that breed in western 112 113 Argentina in spring and summer and that migrate eastward in Argentina during autumnwinter. 114

115 Data analysis

To evaluate the composition of the bird communities, species were grouped into six functional guilds. The guilds were defined on the basis of vegetation strata and their associated food resources, which are exploited by birds, and by the way the birds obtain food, using sections in the Handbook of the Birds of the World (del Hoyo et al. 1992, 1994, 1996, 1997,

120 1999, 2001, 2002, 2003, 2004, 2005, 2006, 2009, 2010, 2011) and personal notes. Defined
121 guilds were:

122 - Ground-feeding guild: species that obtain their food mainly on the ground, resting on123 it.

- Shrub guild: species that obtain their food mainly in the shrub stratum or at the
height of the herbaceous stratum (usually 0.5 - 3 m).

Low canopy guild: species that obtain their food mainly in the lower part of thecanopy or on upper levels of the shrub layer.

128 - Canopy guild: species that obtain their food mainly in the canopy or at the top of tall129 trees and shrubs.

130 - Trunk-dwelling guild: species that obtain their food mainly by climbing on tree131 trunks.

- Aerial-feeders guild: species that obtain their food mainly by hunting in flight, either
by short maneuvers from a perch or by constant flights.

In order to obtain more representative samples of the bird community that inhabits 134 forests with different management, we combined the observations made at each point count 135 location in a given season. Thus, each sample included the results of six observation events: 136 137 two observation dates per season in three years. We compared richness and abundance of each of the six functional guilds using Mixed Generalized Linear Models (MGLM), considering 138 "Management", "Season" and first-order interactions ("Management x Season") as fixed 139 effects, and "sampling point" (levels = 36) nested in "transect" (9 = levels) as random effect 140 141 variables, recognizing the interdependence of point-count samples (samples repeated at the same point) and their proximity in space (points within the same transect were relatively closer 142 to each other). "Point count" was not evaluated for possible elimination in the model (due to 143

lack of associated variance), but was always part of the models in order to correctly model 144 145 sampling design. To evaluate statistical significance of the fixed effects, we started with a general model with all variables included. We then used a process of stepwise and backwards 146 selection of variables, by comparing nested models and calculating significance through the 147 148 likelihood ratio test (LRT). When first-order interactions were significant, final models were made for each season, totaling four models for each response variable analyzed according to 149 "Management" as a fixed effect variable, and samples nested in the "transect" random variable 150 associated with sampling design. 151

In order to evaluate vegetation structure variables (canopy cover, shrub layer cover,
herbaceous layer height) we considered "Management", "Season", "Year" and all the first and
second-order interactions (Management*Year*Season; Management*Year;

Management*Season; Year*Season) as fixed effects, and the same random effect variables 155 156 mentioned in the previous paragraph. To evaluate statistical significance of the interactions and fixed effects, we started with a general model with all the variables included and the first 157 and second-order interactions. We then used a process of stepwise and backwards selection of 158 variables, by comparing nested models and calculating significance through the likelihood 159 ratio test (LRT). When we found significant second-order interactions for vegetation variables, 160 161 final models were made for each season of each year. Thus, we had twelve models (4 seasons x 3 years) for each dependent variable analyzed, with "management" as a fixed-effect 162 variable, plus the random variables associated with the sampling design. If second order 163 164 interaction were not found, but there was a significant first-order interaction, we analyzed the 165 data using one model for each season of each year, as we mentioned in the previous sentence. In models where the variable "Management" was significant, we used Tukey test (HSD) to 166 make comparisons between pairs of factor levels. These analyses were implemented in R 167

(Team 2017) using the "lme4" package, "glmmADMB" and the "multcomp" package to apply
the Tukey test. In all cases, we report means ± Standard Deviation.

170 **Results**

Canopy cover was similar among the types of management throughout the three years 171 172 of sampling (Fig. 1A; see Appendix A: Table 1). In IwC, we found fewer climbing vines, 173 compared to NC and RCE forests. *Cortaderia selloana*, a tall herb, was included in the shrub 174 layer due to its large size. This was the most abundant species in the shrub stratum in IwC. The shrub stratum only showed greater coverage in RCE with respect to IwC although these 175 differences were not constant during the entire sampling period, with variations in the 176 177 magnitude of the average differences between management, seasons and years and were only important during the autumn season in all years and in winter of 2014 (Fig. 1B, see Appendix 178 179 A: Table 1). The height of the herbaceous stratum also showed variations in the magnitude of 180 the average differences between management types, seasons and years, which was lower in IwC compared to NC and RCE, except during summer in all the years and during winter and 181 spring of 2016, when heights did not differ among management types (Fig. 1C, see Appendix 182 A: Table 1). 183

In the total study area we recorded 20,605 individuals belonging to 130 species, 33 families and 16 bird orders, of which 81 species (20,082 individuals) were grouped into six functional guilds (see Appendix A: Table 3). The remaining 49 species were not analyzed with respect to guild structure, because in many cases they were rare or occasional species (less than five detections during the entire sampling period), or species associated with grassland and/or marsh environments, which use the forests occasionally or exceptionally (for example, during periods of flooding).

The ground-feeding guild presented the greatest richness and relative abundance of birds in the total study area. The trunk-dwelling and aerial-feeders guilds presented the least richness and relative abundance of birds (Fig. 2A). In IwC, the ground-feeding, canopy and aerial-feeders guilds presented the highest relative abundance respect to others managements types (Fig. 2). In RCE, the shrub and low canopy guilds presented the highest relative abundance (Fig. 2).

197 <u>Ground-feeding guild</u>. The species richness by point count location was higher in IwC 198 than in RCE and NC in all seasons. Total abundance of individuals by point count location 199 was greater in IwC than RCE and NC in all seasons, except in the autumn where IwC was 200 similar to NC, and RCE had lower values (Table 1; see Appendix A: Table 2).

201 Shrub guild. The species richness by point count location did not present differences 202 between the different types of management in any of the four seasons. The abundance of 203 individuals by point count location was lower in IwC than in RCE and NC in autumn and 204 winter seasons, while in summer there were no differences and in spring it was higher in RCE 205 than in IwC and NC (Table 1; see Appendix A: Table 2).

Low canopy guild. The species richness by point count location showed no differences between management types in summer, but was lower in IwC than RCE and NC during autumn and winter, while in spring it was higher in RCE than IwC and NC). The abundance of individuals by point count location did not differ between management types during the summer, but was lower in IwC than RCE and NC during autumn, winter, and spring (Table 1; see Appendix A: Table 2).

<u>Canopy guild</u>. The species richness by point count location did not differ between
 management types in any season. The abundance of individuals by point count location was
 lower in NC during summer and winter with respect to IwC and RCE, in autumn it was higher

in IwC with respect to the other managements, and in spring it did not present differencesbetween types of managements (Table 1; see Appendix A: Table 2).

Aerial-feeders guild. The species richness by point count location was lower in IwC than in RCE and NC only during the winter season, while there were no differences between management types in other seasons. The abundance of individuals by point count location did not differ between management types (Table 1; see Appendix A: Table 2).

221 <u>Trunk-dwelling guild.</u> The species richness and the abundance of individuals by point 222 count location showed no differences between management types in any season (Table 1; see 223 Appendix A: Table 2).

224 Discussion

The ground-feeding guild presented the greatest differences with respect to type of 225 226 cattle raising management, and in general, species that make up this guild were positively 227 associated with IwC. The shrub guild presented the higher abundance values detected during cold seasons (autumn and winter) in NC and mainly in RCE may respond to a greater supply 228 of food resources associated with the lower strata, with higher values of shrub cover in RCE, 229 230 and little developed in IwC. The low canopy species showed a strong association with RCE 231 and NC. The main differences in richness and abundance values occurred in cold seasons, 232 possibly due to the greater supply of food resources associated with climbing vine that colonized most of tree trunks in RCE and NC, which were not present in IwC. Many species 233 of the canopy guild are migratory species that reside in the area during warm seasons. This 234 235 explains the decrease in abundance of this guild during the cold seasons in RCE, as they migratory species were very abundant in that management during the warm seasons. The 236 species of the aerial-feeders guild generated great changes in community at a temporal scale 237 (among seasons), since most of them are migrant species that reside in the area during the 238

warm seasons. The group of species of the <u>trunk-dwelling guild</u> was very stable with respect
to their abundances in different types of management and different seasons. Our bird species
classification into functional guilds reflected structural changes in vegetation strata, so we
recommend their implementation in future studies that focus in structural changes of the
environment.

Reactions of bird communities to different disturbances that modify access to soil 244 245 resources and low vegetation strata, such as fire and cattle raising, have been studied in different regions of the planet (Laiolo et al. 2004, Powell 2008), where species that exploit 246 247 ground resources generally benefit from herbivory, while species associated with shrubs and 248 high grass are negatively affected. More complex studies have evaluated the response of communities to heterogeneity of patches with different fire and herbivory management, 249 250 finding species that are favored by such heterogeneity, while others are preferentially 251 associated with certain patches (Fuhlendorf et al. 2006, Isacch & Cardoni 2011). A recent research concluded that, on an international scale, grazing causes a decrease 252 253 in diversity at all trophic levels, except detritivores (Filazzola et al. 2020). However, 254 researchers from different regions of the planet suggest that grazing exclusion may have negative consequences for biodiversity, mainly for organisms that have evolved in 255 256 environments with high herbivory pressure and where populations of native herbivores have decreased or have become extinct (Fuhlendorf & Engle 2001; Laiolo et al. 2004; Fuhlendorf et 257 al. 2006; Cingolani et al. 2008; Coppedge et al. 2008; Ahlering & Merkord 2016; Ferreira et 258 259 al. 2020).

In areas with different types of cattle raising within the floodplain of the Paraná River, we found that changes in bird community composition were strongly related to species associated with low vegetation strata (ground-feeding guild, shrub guild, and canopy guild),

263 and that responses of different guilds were dissimilar. Ground-feeding species showed a 264 positive association with cattle raising, while shrub stratum and low canopy species showed a positive association with absence of cattle raising, mainly with recently excluded areas. On the 265 other hand, guild species composed of species less strongly associated with vegetation strata 266 267 modified by type of management showed little or no differences between sites. Understanding 268 the direction of changes in biodiversity associated with environmental modifications allows us to optimize conservation efforts. If such efforts are based on conservation of environment in 269 270 its natural state, we should adjust herbivory management to pristine conditions of the environment (for our study area it was discussed in Frutos et al. 2020). On the other hand, if 271 272 conservation efforts are based on particular species or groups of species associated with particular resources, correct management of the herbivory may be fundamental to obtaining 273 successful results. 274

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Author contributions: A.E.F. and C.I.P conceived the project. A.E.F performed the data
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287 Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, atXXXXX.

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Fig. 1. Mean values ± SE of canopy cover (A), shrub layer cover (B) and herbaceous layer height
(C) per point count location in each of four seasons in 2014, 2015 and 2016 in albardón forests
under three different types of cattle-raising management. In cases where there were differences
between the different types of management, the letters indicate similarities (equal letters) and
differences (different letters) between management types (Tukey test).

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Fig. 2. Relative abundance [%] and total richness (in italics) for the different guilds in the total
study area (A); Relative abundance [%] and total richness for the different types of cattle-raising
management in albardón forests for Ground-feeding guild (B); Shrub guild (C); Low canopy
guild (D); Canopy guild (E); Trunk-dwelling guild (F); Aerial-feeders guild (G).

453

454 **Table 1.** Mean values \pm SE of total number of individuals and total richness of birds species for 455 each guild per point count location in each of four seasons in 2014, 2015 and 2016 in albardón forests under three different types of cattle-raising management. Shadings represent differences 456 between types of management (result of a posteriori comparisons by Tukey test); the gray 457 458 triangle indicates that the value of this treatment did not show differences with the others. IwC: Island with Cattle; RCE: Recent Cattle Exclosure; NC: Non Cattle (> 20 years of cattle 459 460 exclosure). Each count point was sampled 6 times per season (2 samples per season during the 461 years 2014, 2015 and 2016).

		Summer			Autumn			Winter			Spring		
		IwC	RCE	NC									
Average abundance per sample	Ground-feeding guild	68.8 ± 3.0	42.7 ± 2.3	39.8 ± 1.3	57.3 ± 5.0	39.3 ± 3.2	47.8 ± 1.7	62.7 ± 3.9	45.6 ± 3.3	41.4 ± 2.1	82.9 ± 3.9	52.8 ± 2.3	52.7 ± 1.2
	Shrub guild	19.5 ± 1.1	21.8 ± 1.4	19.4 ± 1.1	17.0 ± 0.8	22.8 ± 1.5	22.0 ± 0.9	14.4 ± 1.0	25.7 ± 1.9	22.8 ± 1.4	19.1 ± 1.2	27.5 ± 1.8	20.8 ± 1.4
	Low canopy guild	18.5 ± 1.6	25.4 ± 1.7	23.8 ± 1.1	9.4 ± 1.0	23.9 ± 1.9	20.5 ± 1.2	11.1 ± 1.0	23.6 ± 2.4	22.3 ± 1.0	20.3 ± 1.4	33.9 ± 1.5	29.1 ± 1.6
	Canopy guild	26.5 ± 1.1	26.5 ± 1.1	17.9 ± 1.6	26.8 ± 2.1	19.1 ± 0.6	16.6 ± 2.0	26.3 ± 2.0	20.8 ± 1.3	14.7 ± 1.3	26.8 ± 2.4	28.8 ± 1.3	22.3 ± 1.2
	Aerial-feeders guild	9.0 ± 1.2	8.3 ± 1.3	9.3 ± 1.9	4.4 ± 0.7	4.3 ± 1.0	1.8 ± 0.5	10.6 ± 1.4	8.7 ± 1.3	5.9 ± 1.2	19.7 ± 2.0	13.6 ± 1.6	13.7 ± 1.2
	Trunk-dwelling guild	10.8 ± 0.9	10.7 ± 1.5	10.7 ± 1.2	11.9 ± 1.2	12.0 ± 1.7	13.2 ± 1.2	12.1 ± 1.1	12.5 ± 1.2	12.4 ± 0.9	10.7 ± 1.1	14.2 ± 1.4	12.2 ± 1.0
Average richness per sample	Ground-feeding guild	16.2 ± 0.4	11.1 ± 0.6	10.2 ± 0.5	12.4 ± 0.4	10.1 ± 0.4	10.4 ± 0.5	14.2 ± 0.4	11.4 ± 0.4	10.8 ± 0.6	17.1 ± 0.4	13.4 ± 0.4	13.0 ± 0.7
	Shrub guild	5.1 ± 0.4	4.9 ± 0.3	4.6 ± 0.4	3.7 ± 0.3	4.6 ± 0.3	3.4 ± 0.3	3.8 ± 0.3	5.1 ± 0.4	4.0 ± 0.2	5.6 ± 0.6	6.2 ± 0.3	4.6 ± 0.4
	Low canopy guild	5.0 ± 0.3	5.8 ± 0.4	5.5 ± 0.4	3.3 ± 0.3	5.4 ± 0.4	5.3 ± 0.4	3.8 ± 0.3	5.5 ± 0.2	5.1 ± 0.3	5.8 ± 0.3	8.1 ± 0.4	6.2 ± 0.4
	Canopy guild	6.2 ± 0.3	5.7 ± 0.3	6.2 ± 0.4	5.3 ± 0.4	5.3 ± 0.3	5.3 ± 0.3	5.6 ± 0.3	4.6 ± 0.3	5.0 ± 0.3	5.8 ± 0.3	6.3 ± 0.4	6.4 ± 0.2
	Aerial-feeders guild	3.8 ± 0.3	3.1 ± 0.4	3.4 ± 0.5	1.4 ± 0.2	1.8 ± 0.3	1.2 ± 0.2	1.3 ± 0.1	2.4 ± 0.1	2.1 ± 0.2	5.3 ± 0.5	4.7 ± 0.3	4.6 ± 0.3
	Trunk-dwelling guild	3.6 ± 0.2	3.8 ± 0.4	3.6 ± 0.2	3.9 ± 0.4	4.0 ± 0.3	4.7 ± 0.3	3.8 ± 0.2	4.1 ± 0.3	4.3 ± 0.3	4.0 ± 0.2	4.4 ± 0.3	4.3 ± 0.3